Proposed Special Review Decision

Santé

PSRD2018-02

Special Review of Thiamethoxam Risk to **Aquatic Invertebrates: Proposed Decision for** Consultation

Consultation Document

(publié aussi en français)

15 August 2018

This document is published by the Health Canada Pest Management Regulatory Agency. For further information, please contact:

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ISSN: 2561-6366 (online)

Catalogue number: H113-30/2018-2E (print)

H113-30/2018-2E-PDF (PDF version)

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1.0 Introduction

The Pest Management Regulatory Agency (PMRA) initiated a special review of thiamethoxam under subsection 17(1) of the *Pest Control Products Act* based a preliminary analysis of available information on the concentrations and frequency of detection of thiamethoxam in aquatic environments.

As required by subsection 18(4) of the *Pest Control Products Act*, the PMRA has evaluated the aspects of concern that prompted the special review of pest control products containing thiamethoxam. The aspect of concern for this review is to assess potential risk to aquatic invertebrates exposed to thiamethoxam applied as a seed, foliar or soil treatment.

2.0 Uses of Thiamethoxam in Canada

Appendix I lists all thiamethoxam products with agricultural uses that are registered under the authority of the *Pest Control Products Act* as of May 2018 that were subject to this special review. Thiamethoxam is currently found in 18 agricultural end-use products to which aquatic invertebrates may be exposed. These products may be used in greenhouses (peppers and ornamentals), as a seed dressing (various cereal, pulse and vegetable crops, sunflower, and potato as a seed piece treatment), foliar spray application (ornamentals, potato, pome fruit, stone fruit, bush berries, caneberries, and various vegetable crops), or in-furrow drench (potato, various vegetable crops). Foliar spray applications can be made by ground boom, airblast or aerial sprayers, depending on crop. Appendix II lists all registered uses of Commercial Class end-use products containing thiamethoxam that were subject to this special review.

3.0 Aspects of Concern that Prompted the Special Review

This special review was initiated on 23 November 2016, at the same time the PMRA's proposed cyclical re-evaluation decision was published for imidacloprid (PRVD2016-20). The aquatic risk assessment for imidacloprid identified risks of concern to aquatic invertebrates. Thiamethoxam shares the same mode of action with a similar toxicity profile. Available monitoring data indicated that thiamethoxam was being detected at concentrations and frequencies in aquatic environments that may pose a risk to aquatic invertebrates. A preliminary assessment was conducted to determine if a special review was required. Based on the available fate, toxicity and water monitoring information for thiamethoxam, there were reasonable grounds to believe that the potential risk to aquatic invertebrates from the use of thiamethoxam may exceed the PMRA's level of concern under the current conditions of use.

The initiation of the special review was announced in REV2016-17, *Initiation of Special Reviews: Potential Environmental Risk to Aquatic Invertebrates Related to the Use of Clothianidin and Thiamethoxam*. The aspect of concern for this special review is to assess potential risk to aquatic invertebrates exposed to thiamethoxam applied as a seed, foliar or soil treatment.

4.0 PMRA Evaluation of the Aspects of Concern

The PMRA required the registrant to submit all available data that are relevant to the environmental fate of thiamethoxam, including Canadian surface water monitoring data, and to its toxicity to aquatic invertebrates. In addition, the PMRA requested the same information from provinces and other relevant federal departments and agencies, in accordance with subsection 18(2) of the *Pest Control Products Act*. In response to PMRA's requests, information was received related to the aspect of concern.

Additional data supplied by the registrant included information on the environmental fate of thiamethoxam in soil and water as well as the ecotoxicity of thiamethoxam and its major transformation products to aquatic invertebrates. Data on thiamethoxam toxicity to aquatic invertebrates generated by Environment and Climate Change Canada (ECCC) and by academic researchers was included for this special review. A comprehensive literature review of current data relevant to the special review provided additional ecotoxicity data for thiamethoxam. In total, the PMRA considered acute ecotoxicity data for 44 species of aquatic invertebrates and chronic data for 8 species, as well as higher-tier community-based endpoints from two studies. Environmental incidents for aquatic invertebrates were not identified in North America.

Published and unpublished Canadian freshwater monitoring data were received from federal and provincial governments and acamedic researchers, registrant companies, and members of Agriculture and Agri-Food Canada's Multi-stakeholder Environmental Monitoring Working Group. Freshwater monitoring data consisted of several robust datasets often with large numbers of samples taken at high frequencies from agricultural areas from 2010 to 2017.

Key Findings

The environmental assessment showed that, in aquatic environments in Canada, thiamethoxam is being measured at concentrations that are harmful to aquatic insects. These insects are an important part of the ecosystem, including as a food source for fish, birds and other animals. Based on currently available information, the continued use of thiamethoxam outdoors in agricultural areas in Canada is not sustainable. For more information on Health Canada's proposed decision for this special review of thiamethoxam, refer to Section 5.0.

Risk Assessment Conclusions

In conducting environmental risk assessments, it is the PMRA's policy to always consider both monitoring data (when available) and estimated environmental concentrations (EECs) generated using water models as part of its overall risk assessment. Although valid monitoring data are considered preferable to modelled EECs, the weight given to these data varies depending on the circumstances.

When determining the most appropriate toxicity endpoints for consideration in the risk assessment, the PMRA considers both registrant submitted studies and publically available studies. The ecotoxicity data is considered in a tiered approach, which consists of the following:

- the endpoint of the most sensitive species,
- a species sensitivity distribution when enough data points are available, and
- mesocosm studies which considers effects at the community level.

For thiamethoxam, Species Sensitivity Distributions (SSDs) for both acute and chronic exposure in freshwater environments were determined. In addition, two acceptable mesocosm studies were available to assess the concentrations at which community level effects would be observed. For the chronic assessment, the endpoints from the most sensitive mesocosm study, the SSD and the most sensitive single species were considered in a weight-of-evidence approach in the risk assessment.

A major transformation product of thiamethoxam in soil is clothianidin, another registered insecticide, which is also toxic to aquatic invertebrates. The risk assessment conclusion for thiamethoxam considers thiamethoxam alone and not the combination of thiamethoxam and clothianidin. Given that both pesticides are registered for use on many of the same crops, it is often not possible to determine whether concentrations of clothianidin measured in water are a result of the transformation of thiamethoxam, a result of the use of clothianidin as an insecticide, or a combination of the two. Concurrent with this special review of the risk of thiamethoxam to aquatic invertebrates, the PMRA has conducted a separate special review of the risk of clothianidin to aquatic invertebrates. The clothianidin special review is published in PSRD2018-01.

The risk assessment based on the modelling results indicates that exposure to thiamethoxam poses a minimal acute risk to freshwater invertebrates; however, on a chronic basis, exposure to thiamethoxam poses a risk to freshwater invertebrates. Typically, modelling inputs and assumptions are conservative and the EECs generated are likely to be higher than actual concentrations present in waterbodies. For thiamethoxam, however, the range of surface water EECs predicted from modelling overlaps with the range of concentrations measured in surface freshwater bodies.

Thiamethoxam concentrations measured in Canadian waterbodies did not exceed the acute level of concern. Chronic risks to freshwater invertebrates were identified based on robust Canadian monitoring data sets. Monitoring data likely provide an underestimate of acute exposure, as sampling typically does not capture peak concentrations.

Thiamethoxam concentrations detected in the following areas frequently exceeded the chronic SSD endpoint for freshwater invertebrates (the registered methods of application of clothianidin are listed in parentheses):

- Corn and soybean growing regions (seed treatment),
- Potatoes (seed treatment, soil application or foliar spray), and
- Vegetables (seed treatment or foliar spray, depending on the type).

Concentrations of thiamethoxam occasionally exceeded the chronic SSD endpoint for freshwater invertebrates in a few waterbodies located in areas where orchards occupy large portions of the cultivated area of the watershed.

The chronic SSD endpoint was exceeded in wetlands primarily associated with seed treatment uses in the Prairies; however, there was uncertainty surrounding the duration of exposure. Therefore, in the Prairie region, there is insufficient evidence to conclude that thiamethoxam uses are acceptable.

Concentrations detected in some waterbodies located in regions growing potatoes, mixed vegetables and corn/soybean exceeded the PMRA level of concern based on the mesocosm endpoint for periods of weeks to months. This chronic exposure may result in effects at the community level, including changes in insect species abundance and emergence. Concentrations of thiamethoxam exceeding the community-level endpoint were also detected in other cropgrowing regions, however, they were sporadic and of short duration. The occurrence of thiamethoxam concentrations at or above the community-level endpoint may have significant impacts on community invertebrate structure which is a primary protection goal of the PMRA.

No Canadian monitoring data for thiamethoxam in marine or estuarine water were available to exclude risks to marine/estuarine invertebrates.

4.1 Fate and Behaviour in the Environment

A summary of all available information pertaining to the fate and behaviour of thiamethoxam in the environment is provided in Appendix III. The environmental fate and behaviour of thiamethoxam are summarized as follows:

- Thiamethoxam will come in contact with soil when it is applied directly on the ground, sprayed on foliage, or when thiamethoxam contained in the seed coating moves away from the seed into the surrounding soil. The length of time that thiamethoxam will persist in soil depends on various factors including soil type. In certain fields, thiamethoxam may persist long enough to carryover from one growing season to the next.
- Major products formed from the microbial degradation of thiamethoxam in soil are 1-(2-chloro-thiazol-5-ylmethyl)-3-methyl-*N*-nitroguanidine [CGA 322704 (clothianidin)] and 3-(2-chloro-thiazol-5-ylmethyl)-5-methyl[1,3,5]oxadiazinan-4-one (CGA 355190), both of which may also persist in soil. CGA 322704 has been found in rotational crops.
- Thiamethoxam can leach through the soil profile and has been detected in groundwater. The transformation product CGA 322704 (clothianidin) has been found in both soil pore water and in groundwater. Another transformation product, CGA 355190, has been found sporadically in soil pore water but was not detected in groundwater.
- Thiamethoxam may enter the aquatic environment through spray drift or runoff. Thiamethoxam readily dissolves in water and is not expected to enter the air or break down by chemical reactions with water molecules in waters of environmentally relevant pH.

- In water, thiamethoxam is expected to dissipate relatively quickly if exposed to sunlight. In the absence of sunlight, thiamethoxam will be broken down more slowly by microbes. In the laboratory, thiamethoxam is non-persistent to moderately persistent in water systems containing sediment. Under more realistic conditions in an outdoor study, thiamethoxam was non-persistent.
- Thiamethoxam and its transformation product CGA 322704 (clothianidin) are frequently found in surface waters located in Canadian agricultural areas.
- Major products formed from the break down of thiamethoxam in water of high pH (alkaline conditions) include CGA 355190 and 1-(2-chloro-thiazol-5-ylmethyl)-3-nitrourea (NOA 404617), which further breaks down to 2-chlorothiazoly-5-lmethyl-amine (CGA 309335). The major products 3-methyl-1,3,5]oxadiazinan-4-ylideneamine (CGA 353042) and carbonyl sulfide are formed in the presence of sunlight. In the presence of microbes, thiamethoxam breaks down to 3-(2-chloro-thiazol-5-ylmethyl)-5-methyl[1,3,5]oxadiazinan-4-ylideneamine (NOA 407475), which is found primarily in sediments.
- Residues relevant in the aquatic environment include thiamethoxam, and the major products CGA 353042 and NOA 407475 (both in water and sediment). CGA 355190, NOA 404617 and CGA 309335 may be relevant in alkaline systems; however such conditions are not common in the natural environment. High amounts of carbonyl sulfide are not expected in aquatic systems. CGA 322704 (clothianidin) formed from the breakdown of thiamethoxam in soils can leach to groundwater and it can be transported to waterbodies through runoff.

4.2 Mode of Action

Thiamethoxam is a second-generation neonicotinoid insecticide. Thiamethoxam is classified by the Insecticide Resistance Action Committee (IRAC) as a Group 4A mode of action insecticide. It acts via contact exposure or ingestion by binding to the nicotinic acetylcholine receptor sites in the central nervous system of insect pests. While the enzyme acetylcholinesterase normally breaks down acetylcholine to terminate signals from these receptors, it does not readily break down neonicotinoid insecticides. The prolonged stimulation of the cholinergic nerves leads to paralysis and eventually death. Neonicotinoids are known to have greater affinity for the insect nicotinic acetylcholine receptors than those of birds or mammals. The reason for this is that nicotinic acetylcholine receptors are different in insects and vertebrates thus affecting the ability to bind nicotinoids (described in detail in Tomizawa and Casida, 2003 and 2005).

4.3 Toxicity to Aquatic Invertebrates

A summary of thiamethoxam toxicity data available for aquatic invertebrates is presented in Table A.3-8 for the technical grade active ingredient and end-use products formulated with thiamethoxam alone, and in Table A.3-9 for transformation products of thiamethoxam. Toxicity information was assessed from registrant-generated studies, government and academia-generated studies and published studies in the open literature. Endpoints for acute toxicity studies with aquatic invertebrates were reported as either EC₅₀ or LC₅₀ values. Sub-lethal EC₅₀ endpoints

were generally characterized by immobilization of the animal. As immobilization often occurred, followed by mortality in test subjects, several of the reported EC₅₀ values included both immobilization and mortality effects, which are identified in Table A.3-8 and Table A.3-9. In the cases where the observed effect was due to mortality alone, the LC₅₀ is provided. As immobility can significantly impact the survival of an aquatic invertebrate in the natural environment, EC₅₀ and LC₅₀ values are considered as an equivalent measure of mortality for this group of animals.

4.3.1 Thiamethoxam and Its End-use Products

Freshwater invertebrates

Thiamethoxam toxicity to freshwater invertebrates differs according to taxanomic group. Crustaceans belonging to Cladocera and Copepoda orders are generally less sensitive, with acute thiamethoxam endpoints for a variety of species, such as *Daphnia magna*, generally ranging from $> 25~000~\mu g$ a.i./L to $> 106~000~\mu g$ a.i./L (Table A.3-8). However, exposure to formulated product can be highly toxic to *Daphnia magna* (48-h EC₅₀ = 27.3 μg a.i./L). This suggests that components of the formulations may be contributing to the toxicity.

Thiamethoxam is, however, very highly toxic to moderately toxic to crustaceans belonging to the ostracod, amphipod, isopod and decapod groups. Acute endpoints based on observed immobilization and/or mortality resulted in the most sensitive EC₅₀/LC₅₀ values reported from studies ranging from 84 μg a.i./L (*Asellus aquaticus*, 48-h EC₅₀ immobilization) to 4775μg a.i./L (*Caecidotea* sp., 96-h EC₅₀ immobilization).

Thiamethoxam is generally non-toxic to rotifers, molluscs and annelids. Acute endpoints based on observed immobilization and/or mortality resulted in EC₅₀/LC₅₀ values ranging from > 691 μ g a.i./L (*Lampsilis fasciola*, 48-h LC₅₀) to > 100 000 μ g a.i./L (24 – 48-h EC₅₀s for several species; e.g., *Lymnea stagnalis*).

Thiamethoxam is highly to very highly toxic to freshwater insects including Diptera, Ephemeroptera, Odonata, Hemiptera, Trichoptera and Coleoptera. The most sensitive acute toxicity endpoint for freshwater invertebrates is for the mayfly, *Neocloeon triangulifer* (96-h $EC_{50} = 5.5 \mu g \text{ a.i./L}$). In Raby et al. (2018), which directly compared median sub-lethal effects and lethal effects, EC_{50} values were equal to or lower than LC_{50} values as immobilization generally occurred earlier and at lower concentrations than mortality (Table A.3-8).

Freshwater invertebrates are highly sensitive to chronic (long-term or repeated) exposure of thiamethoxam. Sub-lethal effects, including reductions in reproduction capacity, growth, emergence and sex ratios of insects, were observed at concentrations far below acute median effect concentrations for immobilization and/or lethality. Chronic aquatic exposure data were available for cladocerans, amphipods, molluscs and ephemeropteran and dipteran insects. The most sensitive endpoints seen among these species ranged from 0.43 µg a.i./L (*Cloeon dipterum*, 28-d EC₁₀ immobilization) to 51 000 µg a.i./L (*Daphnia magna*, 21-d NOEC reproduction).

Two studies with sediment exposure to dipteran insects (chironomid sp.) were available for use in the risk assessment. Endpoints from these studies were expressed relative to concentrations in the sediments, overlying water and/or interstitial pore water. Endpoints reported relative to overlying water concentrations were not considered for the risk assessment due to low

recoveries, most likely due to the study design which requires continual renewal of untreated overlying water to maintain adequate water quality for the test organisms. The most sensitive endpoint based on sediment concentrations was 43 µg a.i./kg dry weight (dw) sediment for (*Chironomus riparius*, 30-d NOEL emergence/development). The most sensitive endpoint based on pore water concentrations was 120 µg a.i./L (*C. riparius*, 10-d NOEC dry weight).

Marine invertebrates

Acute toxicity data for marine invertebrates were only available for the Eastern oyster, *Crassostrea virginica*, and the mysid shrimp, *Americamysis bahia*. Thiamethoxam is practically non-toxic to marine molluscs, and is moderately toxic to *A. bahia* (96-h EC₅₀ = 4500 μ g a.i./L). Chronic exposure to thiamethoxam also resulted in significant reduction in survival of *A. bahia* (NOEC survival = 560 μ g a.i./L).

Mesocosm studies

Registrant-submitted studies

Two higher tier studies were conducted with zooplankton, macroinvertebrate and periphyton communities. Studies of effects of thiamethoxam investigated under microcosm/mesocosm conditions simulating aquatic environments show that certain components of the invertebrate community are particularly at risk.

In a microcosm study with a single application of thiamethoxam at test concentrations of 0 (control), 1, 3, 10, 30 and 100 µg a.i./L, no significant adverse effects on phytoplankton community structures or individual populations were observed after 93 days, at time-weighted average (TWA) concentrations up to 34 µg a.i./L (100 µg a.i./L nominal) (PMRA# 2712709). Time-weighted average concentrations were determined by the PMRA for this study due to the disappearance of thiamethoxam from the microcosms by Day 14. Transient effects were noted for some phytoplankton species but no dose-response relationship was observed. Thiamethoxam reduced species abundance of some zooplankton taxa at the highest test concentration, but effects were not considered to be significant. A NOEC_{community} of 9.4 μg a.i./L (TWA-concentration) was determined based on significant reductions in chironomid emergence at the 34 µg a.i./L treatment on Day 15. However, numbers were comparable with those seen in the controls on all other sampling occasions. The lack of difference from controls beyond Day 15 may be due to additional recruitment as the microcosms were not closed to the environment. However, conclusions regarding recovery cannot be made as these results are representative of a single application scenario and thiamethoxam can be applied up to three times per season by foliar application. In addition, due to very low abundance in control and treatment ponds, conclusions could not be made on ephemeropterans, which are known to be sensitive to neonicotinoids.

An additional outdoor mesocosm study with a naturally occurring diverse species assemblage (invertebrates, plants, algae) examined the impact of thiamethoxam exposure on mayfly (ephemeropteran) larval abundance and emergence over 35 days (PMRA# 2681280). Thiamethoxam was applied nine times over the course of the study to maintain nominal treatment rates at concentrations of 0.0 (control), 0.1, 0.3, 1.0, 3.0 and 10 µg a.i./L. Overall timeweighted average concentrations were 101% of nominal (range: 93 – 108%) among all test enclosures; therefore, results were based on nominal exposure levels. Significant reductions in larval abundance and emergence of the mayfly species *Cloeon dipterum* occurred at

concentrations of 1.0 µg a.i./L and above. A NOEC of 0.3 µg a.i./L was determined for effects on larval abundance and emergence of *C. dipterum*. It was also noted in the study that recovery was not clearly demonstrated and was unlikely if exposure would have continued beyond the 35-day study period. This NOEC is considered a valid community-level endpoint to cover the most sensitive invertebrate population observed in thiamethoxam aquatic field studies.

Published literature micro- or mesocosm studies

In a recently published study, Basley and Goulson 2018 (PMRA# 2861918) examined the ability of aquatic invertebrates to colonize aquatic habitats at environmentally relevant concentrations of either thiamethoxam or clothianidin in small-scale outdoor microcosm treatments. Microcosm containers (14 L) were filled with loamy soil with no history of neonicotinoid use and 10 L of fresh tap water and exposed to nominal concentrations of 0 (control), 0.1, 1, 3, 7, 10 and 15 μ g a.i./L of analytical grade pesticide. Microcosms were housed outdoors with no cover to allow for colonization of flying insects and left in-situ for 33 – 38 days, beginning in late August. Invertebrate populations quantified included Ostracoda (likely to have come from the soils) and Chironomidae and Culicidae dipterans.

There was a significant relationship in invertebrate abundance across thiamethoxam exposure concentrations, with a general pattern of reduced numbers at higher concentrations for Chironomidae larvae, Culex larvae and pupae and Ostracoda. The strongest trend in decreasing abundance with concentration was with chironomid larvae; however, variability in abundance was very high, with peak numbers occurring at the lowest treatment level, making it difficult to establish a true NOEC. Ostracoda were the only taxa to show significant reductions relative to controls, occurring at the highest thiamethoxam concentration. The NOEC determined for this species was 10 µg a.i./L. However, variability in abundance in controls was high, and as treatment concentrations were not verified analytically and test conditions were not monitored throughout the study, the PMRA will consider these results in a qualitative manner only.

4.3.2 Thiamethoxam Transformation Products

For a complete listing of thiamethoxam transformation products, including common identifier codes and chemical names, along with a summary of where they are formed, see Table A.3-7. Acute toxicity data were available for the major thiamethoxam transformation products CGA 322704 (clothianidin), CGA 355190, NOA 407475, NOA 459602, CGA 282149, NOA 404617 and NOA 421275. Chronic toxicity data were available for CGA 282149, CGA 353042, NOA 459602, CGA 322704 (clothianidin), SYN 501406 (or, NOA 501406) and NOA 407475 (Table A.3-9).

CGA 322704 (clothianidin) is the most toxic of all thiamethoxam transformation products; it is highly to very highly toxic to freshwater insects and certain crustaceans, though it is practically non-toxic to *Daphnia*. Among toxicity studies submitted with CGA 322704 (clothianidin) for the registration of thiamethoxam, the most sensitive acute toxicity endpoint was a 48-h EC₅₀ = 7 μ g/L for Coleoptera (*Dytiscidae* sp.) and the most sensitive chronic toxicity endpoint is a 28 d-NOEC = 0.55 μ g/L for the emergence of *Chironomus riparius* (a freshwater sediment-dwelling invertebrate) exposed to clothianidin through water application. A full assessment of all available toxicity studies of clothianidin on aquatic invertebrates is available under PSRD2018-01.

Among the data available for the remaining thiamethoxam transformation products, CGA 355190 is moderately toxic to *C. riparius* (48-h EC₅₀ = 4100 μ g/L), while NOA 407475 is slightly toxic to *Daphnia magna* 48-h (EC₅₀ = 82 900 μ g/L). All other transformation products are practically non-toxic to freshwater invertebrates on an acute basis. On a chronic basis, SYN 501406 (NOA 501406), was the most toxic to freshwater invertebrates (28-d NOEC emergence = 1100 μ g/L for *C. riparius* in treated water).

4.4 Risks to Aquatic Invertebrates

The environmental risk assessment integrates the environmental exposure and ecotoxicology information to estimate the potential for adverse effects on non-target species. This integration is achieved by comparing exposure concentrations with concentrations at which adverse effects occur. Estimated environmental concentrations (EECs) are concentrations of pesticide in various environmental media, such as food, water, soil and air. The EECs are estimated using standard models which take into consideration the application rate(s), chemical properties and environmental fate properties, including the dissipation of the pesticide between applications. For this special review, ecotoxicology information includes acute and chronic toxicity data for various aquatic invertebrates. Toxicity endpoints used in risk assessments may be adjusted to account for potential differences in species sensitivity as well as varying protection goals (i.e., protection at the community, population, or individual level).

Initially, a screening level risk assessment is performed to identify pesticides and/or specific uses that do not pose a risk to non-target organisms, and to identify those groups of organisms for which there may be a potential risk. The screening level risk assessment uses simple methods, conservative exposure scenarios (for example, direct application at a maximum cumulative application rate) and sensitive toxicity endpoints. A risk quotient (RQ) is calculated by dividing the exposure estimate by an appropriate toxicity value (RQ = exposure/toxicity), and the risk quotient is then compared to the level of concern (LOC). For aquatic invertebrates, the PMRA's LOC is equal to a RQ = 1. If the screening level risk quotient is below the level of concern, the risk is considered negligible and no further risk characterization is necessary. If the screening level risk quotient is equal to or greater than the level of concern, then a refined risk assessment is performed to further characterize the risk. A refined assessment takes into consideration more realistic exposure scenarios (such as drift to non-target habitats) and might consider different toxicity endpoints. Refinements may include further characterization of risk based on exposure modelling, monitoring data, results from field or mesocosm studies, and probabilistic risk assessment methods. Refinements to the risk assessment may continue until the risk is adequately characterized or no further refinements are possible.

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4.4.1 Thiamethoxam Endpoints

For the assessment of risk, toxicity endpoints for the available aquatic invertebrate species tested were used as surrogates for the wider range of species that can be exposed following treatment with clothianidin. The PMRA takes a tiered approach in determining risk based on the availability of data. When limited data are available and a Species Sensitivity Distribution (SSD) cannot be derived, the most sensitive endpoint identified for a single species is used. When sufficient laboratory data are available to determine an SSD, the HC₅ value (the 5th percentile of the SSD) is used to identify the concentration which is expected to be protective of 95% of the species in the community. When outdoor semi-field or field studies conducted under relevant exposure and environmental conditions are available, the endpoints from these studies may be used preferentially, as they can more closely approximate community-level effects in the natural environment. Table 1 outlines the different thiamethoxam endpoints considered in the current risk assessment.

For freshwater invertebrates, the most sensitive acute endpoint was a 96-h EC₅₀ value for the mayfly larvae *Neocloeon triangulifer* (5.5 μ g a.i./L). For assessing risk, acute single-species endpoints are divided by a factor of two (2) to account for potential differences in species sensitivity as well as protection at the community or population level. The most sensitive chronic endpoint was a 28-day EC₁₀ based on immobilization for the mayfly larvae *Cloeon dipterum* (0.43 μ g a.i./L).

Sufficient laboratory toxicity data were available for freshwater invertebrates to determine acute and chronic HC₅ values for either the acute EC₅₀/LC₅₀ endpoints or the chronic NOEC or EC₁₀/EC₂₀ endpoints. For acute studies reporting both EC₅₀ and LC₅₀ values, large differences were observed between the EC₅₀ (immobility) and LC₅₀ (mortality) values (i.e., EC₅₀s < LC₅₀s) for several species (Table A.3-8), a result that is likely characteristic of the time dependent nature of thiamethoxam toxicity. For neurotoxic substances, such as neonicotinoids, paralysis may result in altered behaviour and increased susceptibility to drift in flowing waters, which may ultimately affect survival in the environment (Raby et al. 2018). In cases where both an EC₅₀ and LC₅₀ were reported, the more sensitive endpoint was chosen for the SSD. Acute and chronic toxicity endpoints were available for 37 and 7 freshwater invertebrate species, respectively. Corresponding acute and chronic HC₅ values (with 90% CI) were 9.0 (3.4 – 19.0) μ g a.i./L and 0.026 (3.5 × 10⁻⁵ – 0.63) μ g a.i./L, respectively. Further details regarding the calculation of HC₅ values are provided in Appendix IV.

The most sensitive community-level endpoint available from a freshwater mesocosm study was a 35-d NOEC of $0.30~\mu g$ a.i./L based on reductions in mayfly abundance and emergence. This study was scientifically sound and was used in the risk assessment.

For marine invertebrates there were an insufficient number of species to determine HC₅ values for acute or chronic endpoints. Risks were assessed for the most sensitive endpoints for individual species as shown in Table 1

Table 1 The different endpoints considered in the thiamethoxam risk assessment for aquatic invertebrates.

Endpoint	Value (µg a.i./L) with confidence interval, where available	Comments	
Freshwater			
Acute most sensitive sp.	2.775	Calculated as 5.55 µg a.i./L divided by 2 ¹ based on 96-h EC ₅₀ <i>N. triangulifer</i> .	
Acute HC ₅	9.0 (3.4 – 19.0)	Calculated by PMRA (n = 37).	
Chronic most sensitive sp.	0.43	28-d EC ₁₀ C. dipterum	
Chronic HC ₅	$0.026 (3.5 \times 10^{-5} - 0.63)$	Calculated by PMRA (n = 7). Uncertainty was identified for this endpoint based on number of available species.	
Mesocosm	0.30	35-d NOEC	
Marine			
Acute most sensitive sp.	2250	Calculated as 4500 μg a.i./L divided by 2 ¹ based on 96-h EC ₅₀ <i>A. bahia.</i>	
Chronic most sensitive sp.	560	28-d NOEC A. bahia	

¹ For assessing risk, acute single-species endpoints are divided by a factor of two (2) to account for potential differences in species sensitivity as well as protection at the community or population level.

Comparison to other reference values

The PMRA's reference values used for assessing risk are compared with reference values available from the public literature in Table 2. In their preliminary aquatic risk assessment, the USEPA (2017) determined risk to aquatic invertebrates based on the most sensitive acceptable endpoints for acute and chronic invertebrate species. These same species were considered in the PMRA risk assessment, but in the case of freshwater invertebrates, the PMRA considered additional acceptable endpoints for derivation of SSDs.

The PMRA-calculated acute HC₅ of 9.0 µg a.i./L for combined immobilization and mortality effects (EC₅₀/LC₅₀ endpoints) for all taxa is similar to acute HC₅ values in two recent reviews by Finnegan et al. (2017) and Raby et al. (2018). Finnegan et al. (2017) report a range of acute HC₅ values for different aquatic invertebrate taxonomic groupings, with comparable HC₅ values of 5.1 and 22.9 μg a.i./L for all invertebrates based on EC₅₀ and LC₅₀ endpoints respectively (Table 2). Acute thiamethoxam HC₅ values for insects alone from this study were 1–2 orders of magnitude lower (more sensitive) than for distributions that either combined all invertebrate taxa or excluded insects (Finnegan et al. 2017). Raby et al. (2018) also report similar acute HC₅ values of 6.09 μg a.i./L for immobilization and 12.29 μg a.i./L for mortality based on EC₅₀ and LC₅₀ endpoints, respectively, using toxicity data that primarily consisted of invertebrates but also included fish, plant/algae and amphibian species. An acute thiamethoxam HC₅ value of 427 µg a.i./L for crustaceans alone from Whiteside et al. (2008) was significantly higher (less sensitive) than for PMRA's acute HC₅ of 9.0 µg a.i./L for all invertebrates. The PMRA recognizes the potential differences in sensitivity between taxa, but did not determine separate SSDs for insect and non-insect taxa in order to identify potential impacts on the complete invertebrate community.

Mineau and Palmer (2013) report an acute HC_5 of 0.74 μg a.i./L that is an order of magnitude lower than other acute endpoints reported above. This HC_5 however was based on a limited number of freshwater and marine invertebrate species and is therefore not representative of the larger thiamethoxam dataset available for the aquatic invertebrate community. Therefore, the HC_5 value determined by the PMRA is considered to be more representative of anticipated field effects following the application of thiamathoxam.

The PMRA's acute HC_5 estimate based on sub-lethal and lethal effects is an order of magnitude higher (less sensitive) than the lower confidence limit of the lethality-based HC_5 for neonicotinoids (0.2 μ g/L) recommended for the protection of aquatic invertebrates by Morrissey et al. (2015). This value was derived using 24–96-h LC_{50} values available for six neonicotinoid active ingredients, which were standardized and weighted by molecular mass to imidacloprid. The HC_5 estimate of Morrissey et al. (2015) is however, largely weighted by the influence of imidacloprid, which makes up 66% of the 178 acute endpoints considered.

The PMRA chronic reference value for thiamethoxam of $0.026~\mu g$ a.i./L based on the HC₅ is an order of magnitude lower than the USEPA reference value of $0.74~\mu g$ a.i./L based on the most sensitive species endpoint. There are no chronic SSD reference values for thiamethoxam alone to compare against, but the PMRA chronic HC₅ is similar to the lower confidence limit of the chronic HC₅ for neonicotinoids ($0.035~\mu g/L$) recommended for the protection of aquatic invertebrates, derived using chronic EC₅₀/LC₅₀ endpoints for clothianidin, imidacloprid and thiacloprid (Morrissey et al. 2015).

Table 2 Comparison of PMRA's thiamethoxam reference values with those from the open literature.

Source (PMRA#)	Reference Value (µg a.i./L)	Comments
Thiamethoxam		
PMRA	Freshwater: 9.0 (acute HC ₅) 0.026 (chronic HC ₅)	Freshwater HC ₅ values based on EC ₅₀ /LC ₅₀ values for 37 species (acute) and on NOEC, EC ₁₀ /EC ₂₀ values for 7 species (chronic).
	5.5 (acute single species) 0.43 (chronic single species)	Freshwater single species values: 96-h EC ₅₀ (<i>N. triangulifer</i>); chronic 28-d EC ₁₀ immobilization (<i>C. dipterum</i>).
	Marine: 4500 (acute) 560 (chronic)	Marine endpoints based on lowest single species values. Acute: 96-h EC ₅₀ (<i>A. bahia</i>); chronic: 28-d NOEC (<i>A. bahia</i>).
USEPA (2017) (PMRA# 2862809)	Freshwater: 35 (acute) 0.74 (chronic) Marine: 6900 (acute) 1100 (chronic)	Reference values for risk assessment are based on the lowest acceptable single-species endpoints for each. Acute: 48 – 96-h EC ₅₀ /LC ₅₀ ; Chronic: NOEC.
Raby et al. (2018) (PMRA# 2842540)	6.09 (acute immobilization) 12.3 (acute mortality)	Combined freshwater and marine HC_5 . Data include acute $48-96$ -h EC_{50} or LC_{50} values for invertebrates from authors' study plus additional taxa from the literature including fish (LC_{50} values) and plants/algae

Source (PMRA#)	Reference Value (µg a.i./L)	Comments
Thiamethoxam		
		(EC/IC ₅₀ values).
Finnegan et al. (2017) (PMRA# 2764640)	5.1 (acute immobilization; all invertebrates) 22.9 (acute mortality; all invertebrates)	Freshwater HC_5 values. Review of registrant-generated invertebrate studies (acute $48-96$ -h EC_{50} or LC_{50} values). Additional EC_{50} and LC_{50} based HC_5 values provided for insects only and all invertebrates, excluding insects.
Mineau and Palmer (2013) (PMRA# 2526820)	0.74 (acute)	Combined freshwater and marine HC ₅ (5 species, 48 – 96-h; crustaceans and insects).
Whiteside et al. (2008) (PMRA# 2862805)	427 (acute)	Freshwater HC ₅ for crustaceans only (24 – 96-h EC/LC ₅₀ values)
Combined neonicotinoids		
Morrissey et al. (2015) (PMRA# 2538669)	0.2 (acute) 0.035 (chronic)	Lower confidence intervals of HC ₅ values from SSDs generated from 42 species (acute 24 – 96-h LC ₅₀ values) and 18 species (chronic 7 – 39-d EC ₅₀ /LC ₅₀ values). SSDs included six neonicotinoid compounds (acute) or clothianidin, imidacloprid and thiacloprid (chronic) standardized and weighted by molecular mass to imidacloprid.

4.4.2 Screening Level Assessment

Estimated environmental concentrations

Screening level EECs for thiamethoxam and its transformation products in water were calculated assuming a reasonable conservative scenario of direct application into waterbodies of 80 cm depth. The pesticide is assumed to be instantaneously and completely mixed within the waterbody. EECs for transformation products assume a 100% transformation from parent. The 80-cm waterbody was chosen to represent a permanent body of water to assess the risk to aquatic invertebrates that depend on a permanent waterbody. The screening level calculation is intended to be a simple, conservative estimate of thiamethoxam and transformation products concentrations in a surface waterbody.

For the initial conservative screening level assessment, EECs were calculated based on the highest maximum annual application rates among all use types and crops. Details on derivation of the cumulative annual application rates for determining EECs can be found in Table A.5-1, Appendix V. The screening level assessment considered the highest foliar cumulative application rate of 178.1 g a.i./ha for apples applied by airblast sprayer, and the highest seed treatment rate of 150 g a.i./ha for a variety of vegetables. In addition, to determine the lower limit of potential risk, the lowest annual rate among all crops of 4.5 g a.i./ha for seed treatment application to sorghum was also considered. Screening level EECs for thiamethoxam transformation products assumed that 100% of the thiamethoxam EEC in 80 cm of water is converted to the transformation product in question, adjusted for the molecular weight ratio of transformation product to thiamethoxam. Screening level EECs for thiamethoxam and its major transformation products in surface waters of 80-cm depth are provided in Table A.5-2.

Assessment of risk

Thiamethoxam

The screening level risk assessment for aquatic invertebrates is presented in Table A.3-10. Acute exposure to thiamethoxam at the highest seed treatment and foliar application rates may present a risk to freshwater invertebrates. Risk quotients exceeded the PMRA's level of concern (LOC) of 1 (RQ \geq 1) based on both the acute HC₅ (RQs up to 2.5) and the acute EC₅₀ for the most sensitive species, *Neocloeon triangulifer* (RQs up to 8.1). Thiamethoxam is not expected to pose an acute risk to freshwater invertebrates at the lowest seed treatment rate (RQs \leq 0.2). Risk to freshwater invertebrates from chronic exposure to thiamethoxam is expected to be greater than for short-term acute exposure. Screening level risk quotients exceeded the LOC for all application rates based on both the chronic HC₅ (RQs 22–854) and the chronic EC₁₀ for the most sensitive species, *Cloeon dipterum* (RQs 1.3–52). For estuarine/marine invertebrates, acute and chronic risk quotients did not exceed the LOC (RQs <1) for all application rates. Foliar application and seed treatment at the proposed rates are therefore not expected to pose a risk to marine invertebrates.

Transformation products

A screening level risk assessment was performed for water exposures of major transformation products identified from laboratory transformation studies with thiamethoxam (Table A.3-11). Based on acute toxicity studies conducted with *Daphnia magna*, *Asellus aquaticus*, *Cloeon dipterum*, Dytiscidae and *Chironomus riparius* and conservative EEC estimates, exposures to CGA 355190, NOA 407475, NOA 459602, CGA 282149, NOA 404617 and NOA 421275 are not expected to pose a risk to aquatic invertebrates (RQs < 1). Laboratory toxicity studies conducted with *Daphnia magna* and *Chironomus riparius* indicate that chronic exposures to CGA 282149, CGA 353042, NOA 407475, NOA 459602 and SYN 501406 (NOA 501406) are also not expected to pose a risk to aquatic invertebrates (RQs < 1).

The transformation product CGA 322704 (clothianidin) may pose acute and chronic risks to aquatic invertebrates. Based on studies conducted with CGA 322704 (clothianidin) submitted for the registration of thiamethoxam and a conservative assumption of 100% transformation of thiamethoxam to clothianidin for calculation of EECs, adjusted for molecular weight, the LOC was exceeded for both acute (RQs up to 5.4) and chronic exposures (RQ = 35). However, this assessment does not consider the full range of available toxicity data for clothianidin, which is addressed under the risk assessment for clothianidin in PSRD2018-01.

4.4.3 Refined Risk Assessment

Aquatic organisms can be exposed to thiamethoxam as a result of spray drift into an aquatic environment during application and through runoff from the application site. To further characterize potential aquatic risk, inputs from both are assessed separately. As no risk to estuarine/marine invertebrates was identified at the screening level, the refined risk assessment is for aquatic invertebrates in freshwater habitats only.

Spray drift risk assessment

The risk to aquatic invertebrates was further characterized by taking into consideration the concentrations of thiamethoxam that could be deposited through spray drift in aquatic habitats

that are 1 m downwind from the treatment area. End-use products containing thiamethoxam are applied by a variety of foliar spray methods that may result in spray drift, including field sprayer, airblast and aerial sprayer applications. The maximum amount of spray that is expected to deposit 1 m downwind from the application site during application by field and aerial sprayers with an ASAE (American Society of Agricultural and Biological Engineers) S572.1 fine spray droplet size is 11% and 26% respectively. For early and late airblast applications, 74% and 59% of spray is expected to deposit 1 m downwind from the application site, respectively. Given the variation in percent drift off site for each of the application methods, the assessment of potential risk from spray drift was assessed for the maximum cumulative application rate for each foliar application method: for field sprayers, a single application of 150 g a.i./ha for outdoor ornamentals; for airblast spray, a cumulative rate of 178.1 g a.i./ha (2 × 96.25 g a.i./ha with a 10-day interval) for apple; for aerial spray, a cumulative rate of 68.3 g a.i./ha (3 × 25.38 g a.i./ha with a 7-day interval) for dry beans. Details on the derivation of maximum cumulative rates are provided in Table A.5-1. Estimated environmental concentrations from spray drift are provided in Table A.3-12.

The risk to aquatic invertebrates resulting from spray drift is summarized in Table A.3-12. The LOC is not exceeded for freshwater invertebrates exposed to drift at the highest application rates from field (ground boom) or aerial sprayers on an acute basis (RQs \leq 1). However, freshwater invertebrates are at acute risk from airblast spray drift, regardless of whether the HC $_5$ or most sensitive laboratory-derived endpoint is considered (RQs up to 6.0). Freshwater invertebrates are also at chronic risk from thiamethoxam spray drift, regardless of application method or whether the HC $_5$ or most sensitive laboratory-derived endpoint is considered (RQs up to 632). Mitigation in the form of spray buffer zones is proposed for freshwater habitats and is presented in Appendix VIII.

Runoff risk assessment

Aquatic organisms can also be exposed to thiamethoxam as a result of runoff into a body of water. The Pesticide in Water Calculator (PWC) model was used to predict estimated environmental concentrations (EECs) resulting from runoff of thiamethoxam following application. Details on modelling inputs and assumptions are provided in Appendix VI. The models were run for a variety of scenarios to ensure that runoff potential was assessed for a) representative application rates for each of the major application methods, and b) major crop uses across the country (Table 3). Representative seed treatment uses ranged from sweet corn (7.6 g a.i./ha) to peas (150 g a.i./ha). In-furrow uses ranged from potato (140 g a.i./ha) to bell pepper (150 g a.i./ha). Representative foliar spray applications ranged from potato (2 × 26.25 g a.i./ha) to apple (2 × 96.25 g a.i./ha).

Table 3 Thiamethoxam use scenarios selected for surface water modelling

Application Method	Crops selected
Seed treatment	• succulent beans (50 g a.i./ha)
	• succulent peas (150 g a.i./ha)
	• barley (36.3 g a.i./ha)
	• soybean (64 g a.i./ha)
	• canola (32.3 g a.i./ha)
	• spring and winter wheat (52.5 g a.i./ha)
	• potato (117.1 g a.i./ha)
	• corn (118.3 g a.i./ha)
	• sweet corn (7.6 g a.i./ha)
In-furrow drench or surface	• potato (140 g a.i./ha)
band drench plus irrigation	• bell pepper (150 g a.i./ha)
	• lowbush blueberry (140 g a.i./ha)
Foliar Spray	• apple (2 × 96.3 g a.i./ha)
	• potato (2 × 26.3 g a.i./ha)
	• soybean (3 × 25.4 g a.i./ha)
	• bell pepper (2 × 70 g a.i./ha)
	• blueberry (2 × 70 g a.i./ha)
Transplant water	• bell pepper (117 g a.i./ha)

The Level 1 thiamethoxam EECs in a 1-ha receiving waterbody (80 cm deep) predicted by PWC for these crops are presented in Appendix VI. The pore water EECs in a 0.8-m wetland were also generated. Table A.6-3 provides EECs for all selected crops using runoff extraction parameters recommended in Young and Fry (2017) and using a modelling scenario that assumes that, at the time of application, the pesticide is present in soil only at the depth the seed is planted. It alsoprovides alternate EECs for corn, sweet corn and soybean seed treatments generated using a modelling scenario that assumes the pesticide concentration in soil at the time of application linearly increases with depth from the soil surface to the seeding depth. This latter approach takes into consideration the potential impact of dust generated during planting using pneumatic sowing equipment on water EECs. The values reported by PWC are 90th percentile of the concentrations determined at a number of time-frames including the peak (or daily maximum), 96-h, 21-d, 60-d and 90-d averages.

Acute and chronic RQ values were calculated using the EEC for the appropriate time frame which most closely matched the exposure time used to generate the endpoint. For comparison against acute invertebrate endpoints based on data with 24–96-h and 7-d sub-chronic studies, peak EECs were used to derive RQs. Peak EECs were chosen over 96-h EECs as the duration for many of the acute studies considered was < 96-h. There are minimal differences between the peak EECs and the 96-h EECs due to thiamethoxam's persistence in the environment (Table A.6-3), and therefore this choice does not affect risk conclusions. For comparison against chronic invertebrate endpoints based on data with 21 – 40-d NOEC or EC₁₀/EC₂₀ endpoints, 21-day EECs were used to derive the RQs. For comparison against chronic invertebrate endpoints based on pore water exposures, 21-day pore water EECs were used to derive the RQs. The acute and chronic RQ values for aquatic invertebrates are reported in Table A.3-13 (Appendix III). In cases where EECs were modelled for different geographic regions, risk was assessed for each region.

Acute risk

Risk quotients for freshwater invertebrates based on EECs from acute exposure of thiamethoxam in runoff and the acute HC_5 of 9.0 μg a.i./L marginally exceeded the LOC for in-furrow and soil drench uses on potato and blueberry in Atlantic Canada (RQs = 1.1). Risk quotients for thiamethoxam runoff from treated fields did not exceed the LOC for other uses and regions (RQs \leq 0.9).

Thiamethoxam runoff from treated agricultural fields may therefore pose a minimal acute risk to freshwater invertebrates for soil applications (in-furrow and soil drench uses) in potato and blueberry crops in Atlantic Canada. Acute risk from thiamethoxam runoff is not expected from all other modelled uses.

Chronic risk

Freshwater aquatic invertebrates are highly sensitive to chronic thiamethoxam exposure. Risk quotients from exposure to thiamethoxam runoff from treated fields exceeded the LOC based on the chronic 21-d average EECs and chronic HC_5 of 0.026 μg a.i./L for all modelled foliar, transplant water, in-furrow and soil drench use patterns in all regions (RQs = 7.3–346; Table A.3-13). Chronic exposure from seed treatment runoff also exceeded the LOC in at least one region for all modelled uses (RQs range from <0.1 to 129). Chronic risk was also assessed with the most sensitive single-species endpoint available (28-d EC_{10} immobility = 0.43 μg a.i./L for *C. dipterum*). With the exception of foliar use on apples, risk quotients still exceed the PMRA's LOC for all foliar, transplant water, in-furrow and soil drench use patterns (RQs up to 21). Based on the most sensitive single species from laboratory studies, chronic risk exceeded the LOC for seed treatment uses in at least one modelled region for winter wheat, peas and canola, and for corn and soybean modelled with thiamethoxam distribution in soil increasing up to seed planting depth (RQs <0.1–7.8).

Chronic risk from pore water exposure was also assessed for thiamethoxam based on a treated sediment study using C. dilutus (NOEC growth = 120 μ g a.i./L pore water). Thiamethoxam is not expected to pose a chronic risk to aquatic invertebrates from exposure to thiamethoxam in sediment pore water (all RQs <0.1; Table A.3-13).

Further Risk Characterization: Mesocosms

Two acceptable outdoor mesocosm studies were considered for further characterizing the expected level of risk from thiamethoxam inputs to freshwater systems from both spray drift and surface runoff. Both registrant-supplied studies represent potential community-level effects following exposure of thiamethoxam to outdoor artificial ponds. The lowest available NOEC from both studies was used to determine potential risk (Table A.3-8). Based on a NOEC of 0.3 μg a.i./L for reductions in mayfly abundance and emergence, risk from spray drift alone to aquatic habitats exceeds the LOC, with RQs for the highest labelled spray application rates ranging from 6.9–55 (Table A.3-12). Risk from runoff sources to aquatic habitats exceeded the LOC for all modelled foliar applications (with the exception of foliar use on apples in British Columbia), transplant water uses and in-furrow/soil drench plus irrigation uses (RQs = 4.7–30; Table A.3-13). Seed treatment uses that exceeded the LOC in at least one modelled region included winter wheat, peas, beans, canola, corn (modelled with thiamethoxam distributions is soil increasing up to planting depth), and soybean (RQs < 0.1–11).

Further Risk Characterization: Chronic Exposure Level

Chronic runoff EECs used in the refined risk assessment above represent the 90th percentile of the maximum 21-day average EECs over a 50-year period (see Appendix VI for a full description of EEC derivation). The distributions of annual maximum 21-day average EECs for the 50 model years were further characterized to examine the proportion of years where the maximum 21-day average EECs exceeded the LOC. The distribution of the 50 annual maximum 21-day averages for each of the modelled crops and regions are provided in violin plots presented in Figure 1 to Figure 3. The annual maximum 21-day average concentration is plotted along the vertical axis on a logarithmic scale and the width of the plot is proportional to the number of years with similar annual maximum 21-day average concentrations. Three different endpoints are presented on the plots: the chronic HC₅, the lowest single species endpoint and the mesocosm NOEC.

For foliar, in-furrow and transplant water applications, nearly all of the annual maximum 21-d EECs for 50 years exceeded the chronic thiamethoxam HC₅ of 0.026 μg a.i./L by a factor ranging from approximately 10 to 1000 (Figure 1). The exception is for foliar use on apples in British Columbia, where approximately 30% of years exceeded the HC₅. For all other modelled foliar, in-furrow and transplant water applications, the majority of annual maximum 21-d EECs exceeded the most sensitive mesocosm endpoint of 0.30 μg a.i./L and the most sensitive chronic EC₁₀ of 0.43 μg a.i./L.

For seed treatment applications, 50-year distributions were modelled for thiamethoxam concentrations occurring at planting depth (all uses) (Figure 2 and Figure 3) and for thiamethoxam concentrations increasing with soil depth from the surface to planting depth for corn, sweet corn and soybean (Figure 2). For runoff concentrations based on thiamethoxam distributions in the soil at planting depth, there were several crop/region combinations which did not exceed the chronic HC5 either for the entire 50-year period, or for more than 50% of the time. These included sweet corn (all modelled regions) (Figure 2), bean (BC, SK, MB), pea (BC), potato (MB, PEI) (Figure 3). Modelled crops and regions that exceeded the chronic HC5 at least 50% of the time included canola, corn, soybean (all modelled regions) (Figure 2), barley (BC), bean (ON, QC, PEI) and pea (all regions except BC). The only crops modelled for thiamethoxam distributions at planting depth which had the majority of years exceeding the most sensitive mesocosm and single species endpoints were canola (ON, QC) (Figure 2) and pea (PEI) (Figure 3). Of the remaining crops, all except sweet corn (Figure 2), bean (BC, SK, MB), pea (BC) and potato (MB, PEI) (Figure 3) had 21-d EECs approaching (within a factor of 10), or overlapping the mesocosm and most sensitive species endpoints.

For corn, sweet corn and soybean, runoff concentrations based on distributions increasing with depth in the soil were higher than for distributions at planting depth (Figure 2). Annual maximum 21-d EECs for corn and soybean nearly always exceeded the chronic HC₅; however, in general, the majority of annual maximum 21-d EECs for sweet corn were below the HC₅ and other endpoints of concern. Approximately 20 - 40% of corn and soybean distributions exceeded the mesocosm endpoint, with slightly fewer exceeding the most sensitive single species endpoint.

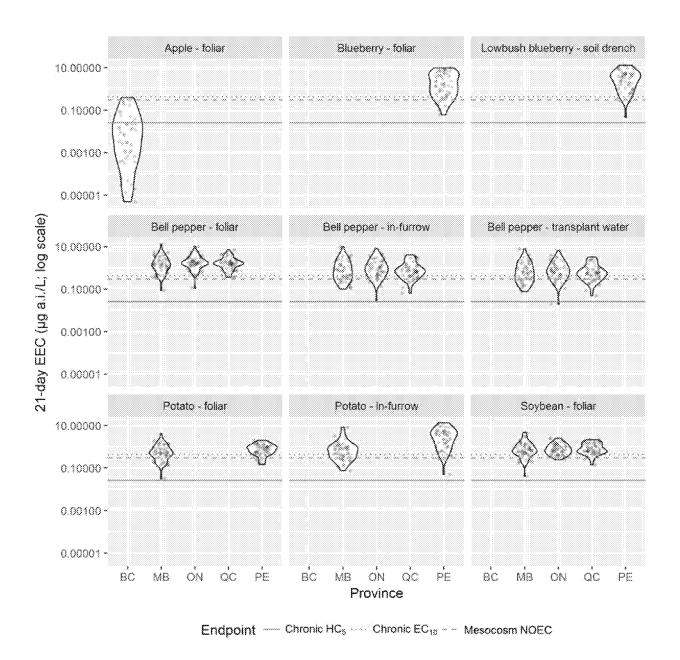


Figure 1 Yearly average 21-day thiamethoxam surface water EECs for modelled foliar, in-furrow, soil drench and transplant water crop uses over a 50-year period compared to chronic endpoints.

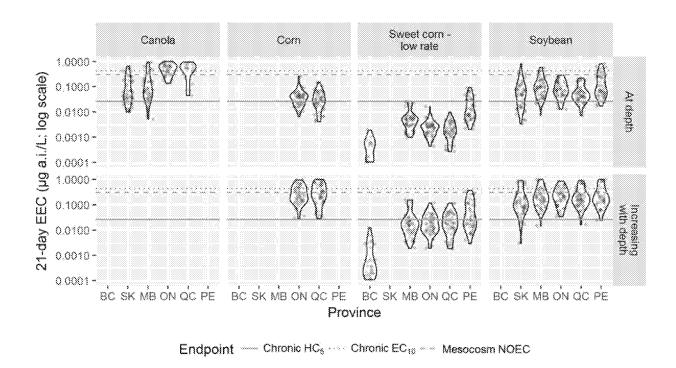


Figure 2 Yearly average 21-day thiamethoxam surface water EECs for modelled seed treatment crop uses over a 50-year period compared to chronic endpoints.

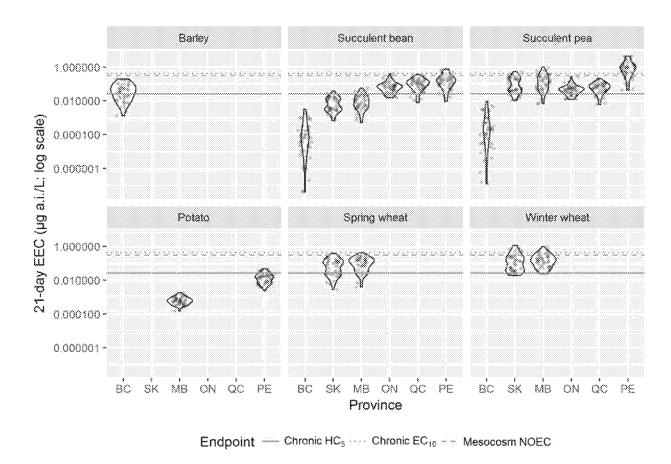


Figure 3 Yearly average 21-day thiamethoxam surface water EECs for modelled seed treatment crop uses over a 50-year period compared to chronic endpoints (continued).

Water monitoring risk assessment

There were sufficient thiamethoxam surface water monitoring data available to consider in the risk assessment for freshwater aquatic invertebrates. No monitoring data for thiamethoxam in estuarine or marine water from Canada were available. This section summarizes available Canadian monitoring data for thiamethoxam in freshwater bodies the PMRA considers to be relevant for use in the risk assessment.

Canadian freshwater monitoring data were available from Prince Edward Island, Nova Scotia, New Brunswick, Ontario, Quebec, Manitoba, Saskatchewan, Alberta and British Columbia. Most sites were located in agricultural areas, but data were also available in urban areas as well as less developed, more pristine sites. The available data for thiamethoxam spanned from 2010 to 2017. Some sites in Quebec and Ontario were sampled over six or seven years; most sites in other locations were sampled over one to three years.

Average concentrations of thiamethoxam can provide an estimate of its presence in water over time. Because the average can be affected by a single value being too high or too low compared to the rest of the values in a data set, median concentrations were also calculated to provide another measure of a middle concentration. The duration of time that concentrations of

thiamethoxam approached or exceeded toxicity endpoints was also considered in the assessment, but exposure estimates for these shorter time periods were not generated. In calculations, the PMRA assigned a value equal to half the limit of detection for samples that showed no detection.

A summary of monitoring data on thiamethoxam in Canadian surface waterbodies is provided in Appendix VII. Table A.7-1 presents data from Prince Edward Island, Nova Scotia and New Brunswick. Table A.7-3 and Table A.7-5 present data from Quebec and Ontario, respectively. Table A.7-7 summarizes data from from Manitoba, Saskatchewan and Alberta and Table A.7-9 presents data from British Columbia. These tables present the number of samples collected at each site, the frequency of detection, the average, median and maximum concentrations as well as how many samples exceed the PMRA's various acute and chronic toxicity endpoints. Risk quotients¹ calculated using measured concentrations and acute and chronic toxicity endpoints are presented in Table A.7-2 for Prince Edward Island, Nova Scotia and New Brunswick, Table A.7-4 for Quebec, Table A.7-6 for Ontario, Table A.7-8 for Manitoba, Saskatchewan and Alberta and Table A.7-10 for British Columbia. Shaded areas in these tables indicate instances where the level of concern is exceeded, meaning that risk quotients equal or exceed a value of 1.

Concentrations of thiamethoxam measured in Canadian waterbodies often exceed chronic toxicity endpoints for freshwater invertebrates throughout the growing season in some agricultural areas, including areas where potatoes, vegetables, corn and soybeans occupy large portions of the watershed. There is also evidence that concentrations in Prairie wetlands, rivers and creeks surrounding fields seeded to a variety of crops often exceed chronic toxicity endpoints at least during some parts of the growing season, particularly in summer and in spring prior to seeding. More details on concentrations detected in these areas follows. Thiamethoxam concentrations sporadically exceed the chronic HC₅ of 0.026 μ g/L in a few waterbodies located in areas where orchards occupy large portions of the watershed. Thiamethoxam concentrations measured in Canadian waterbodies did not exceed the acute toxicity endpoint of 9 μ g/L.

Potatoes

Thiamethoxam can be used on potatoes as a seed treatment, a soil application or a foliar spray. Thiamethoxam concentrations in three waterbodies located in potato-growing areas of Quebec frequently exceeded chronic toxicity endpoints for aquatic invertebrates. Potatoes represented 21% to 47% of the cultivated area of the watershed for the Point-du-Jour Creek, the Chartier Creek and the Blanche River, based on information presented in Giroux 2014 (PMRA# 2544468). Corn, soybeans and cereals are also grown in these watersheds. Corn represents between 21% and 30% of the cultivated area of the three watersheds and cereals represent from 9% to 20%, while soybeans represent 18% of cultivated area in the Point-du-Jour Creek only.

In every year sampled (2010, 2012 and 2017), thiamethoxam concentrations in the three waterbodies exceeded the chronic HC₅ of 0.026 μ g/L between 67% and 100% percent of the time (Table A.7-3). Average and median concentrations exceeded the chronic HC₅ of 0.026 μ g/L for every year sampled. The yearly average and yearly median concentrations ranged from 0.033 μ g/L to 0.41 μ g/L, and from 0.033 μ g/L to 0.26 μ g/L, respectively (Table A.7-3). Risk

Risk quotient = exposure concentration ÷ toxicity endpoint

quotients ranged from 1.3 to 16 for average concentrations of thiamethoxam and from 1.3 and 10 for median concentrations (Table A.7-4).

Thiamethoxam concentrations exceeded the mesocosm NOEC of 0.3 µg/L for community-level effects in 26% to 46% of the samples analyzed in Chartier Creek in 2010, 2012 and 2017. In 2010, thiamethoxam concentrations were higher than the endpoint in seven of the nine samples collected between July 25th and August 22nd. In 2012, the mesocosm NOEC was exceeded in all eight samples collected between late May and late June, and all five samples collected between mid- to late-August. The risk quotient calculated using average concentrations exceeded the level of concern in 2012 (RQ = 1.4; Table A.7-4). In 2017, risk quotients calculated using both average and median concentrations approached the level of concern, at 0.9. In Point-du-Jour Creek, thiamethoxam concentrations approached or were higher than the mesocosm NOEC in two samples collected in Point-du-Jour Creek, in late-May and mid-July 2012, and in one sample in mid-June in 2017. Risk quotients calculated using average and median concentrations in Point-du-Jour Creek using average and median concentrations ranged from 0.1 to 0.6 (Table A.7-4).

Thiamethoxam was seldom detected in waterbodies located in potato-growing areas of Atlantic Canada, and concentrations were below toxicity endpoints (Table A.7-1). While thiamethoxam was rarely detected, clothianidin, a transformation product of thiamethoxam and also a registered neonicotinoid insecticide, was frequently detected in rivers in Atlantic Canada, particularly in Prince Edward Island. The higher detection frequency of clothianidin compared to thiamethoxam could be due to more extensive use of clothianidin than thiamethoxam in this area, to the transformation of thiamethoxam into clothianidin in soil and the subsequent runoff of clothianidin into waterbodies, or to a combination of both.

Mixed vegetables and potatoes

As stated previously, all three methods of application can be used on potatoes (seed treatment, foliar spray, or soil application). Depending on the type of vegetable, thiamethoxam can be used as a seed treatment or as a foliar spray. Waterbodies sampled in vegetable-growing areas of Quebec (Gibeault-Delisle Creek and Norton Creek) had concentrations of thiamethoxam frequently exceeding chronic toxicity endpoints. These waterbodies were sampled two to three times per week from May to August 2013 and 2014. A total of 68% of the watershed upstream of the Gibeault-Delisle Creek sampling site was cultivated, while 46% of the area was cultivated upstream of the Norton Creek site based on information in Giroux 2017 (PMRA# 2821394). In the Gibeault-Delisle Creek watershed, vegetables (mainly carrots, onions, green onions and lettuce) represented 25% of the cultivated area upstream of the sampling site, potatoes represented 21%, and corn and soybeans represented 19% of the area. In Norton Creek, vegetable crops (mainly onions, lettuce, beans, carrots and cucurbits) represented 18% of the cultivated area upstream of the sampling site, potatoes represented approximately 9% of the cultivated area, while corn and soybeans represented approximately 24%.

In 2013 and 2014, thiamethoxam concentrations exceeded the chronic HC₅ of 0.026 μg/L in 53% to 57% of the samples collected in Gibeault-Delisle Creek, and in 11% to 53% of the samples collected in Norton Creek (Table A.7-3). Concentrations of thiamethoxam were higher in Gibeault-Delisle Creek compared to Norton Creek. In Gibeault-Delisle Creek, the yearly average concentrations of thiamethoxam ranged from 0.066 μg/L to 0.27 μg/L, and the yearly median

concentrations ranged from 0.032 μ g/L to 0.034 μ g/L. Associated risk quotients ranged from 2.6 to 10 using the average concentration and from 1.2 to 1.3 using the median concentration (Table A.7-4). In Norton Creek, the yearly average concentration ranged from 0.015 μ g/L to 0.031 μ g/L, while the yearly median concentration ranged from 0.006 μ g/L to 0.028 μ g/L (Table A.7-4). Associated risk quotients for Norton Creek ranged from 0.6 to 1.2 for the average concentration and from 0.2 to 1.1 for the median concentration (Table A.7-4).

In Gibeault-Delisle Creek, thiamethoxam concentrations exceeded the mesocosm NOEC of 0.3 μg/L for community-level effects in three of the 28 samples (11%) of samples collected in 2013 and in one of the 30 samples (3%) collected in 2014 (Table A.7-3). Thiamethoxam concentrations higher than the mesocosm NOEC were observed sporadically in three samples in the 2013 season (in mid-May, late-May and late-June) and in one sample in late-June 2014. The risk quotient based on the mesocosm NOEC and the average concentration in Gibeault-Delisle Creek was 0.9 for the year 2013, which approaches the level of concern of 1 (Table A.7-4). The risk quotient based on the median concentration for 2013 was 0.1, which is below the level of concern.

Water samples were collected in a few watersheds from British Columbia where potatoes and vegetable crops represented large portions of the cultivated area. Potatoes and vegetable crops represent approximately 26% of the cultivated area of the watershed of the Sumas Drainage Canal, while berries and crops such as corn or peas also occupied a significant portion (approximately 16% to 22% each) of the cultivated area (PMRA# 2842169, 2842180). Thiamethoxam was detected at in the Sumas Drainage Canal at concentrations below toxicity endpoints (Table A.7-9).

Due to the mixed uses within the watersheds sampled in Ontario, Quebec and British Columbia, it is not possible, based on the existing monitoring data, to identify which crops are leading to the elevated concentrations of thiamethoxam in some of the waterbodies.

Corn and soybeans

Neonicotinoids are used a seed treatment on corn, soybean and other cereal crops. Thiamethoxam can also be applied as a foliar spray on soybeans. Concentrations of thiamethoxam exceeded the chronic HC₅ of 0.026 μ g/L in several waterbodies located in corn and soybean areas of Quebec and Ontario.

Four rivers located in major corn and soybean areas of Quebec were sampled between 2014 and 2017. Corn and soybeans crops represented between 64% and 83% of the cultivated area of the watersheds, based on information presented in Giroux 2015 (PMRA# 2561884). Other crops in the watershed included cereal crops, which occupied approximately 5% of the cultivated area, and vegetables which represented 0.6% to 11% of the cultivated area.

In each of the four rivers sampled in corn and soybean areas of Quebec between 2012 and 2017, thiamethoxam concentrations exceeded the chronic HC₅ between 27% to 96% of the time (Table A.7-3). For all four years of sampling, the average concentrations of thiamethoxam measured between May and August in the four rivers exceeded the chronic HC₅ of 0.026 μ g/L, with one exception. In 2017, the average concentration of thiamethoxam was 0.023 μ g/L, which is slightly below the chronic HC₅. Risk quotients calculated using average concentrations

approached or exceeded the level of concern for every year sampled, and ranged from 0.9 to 15 (Table A.7-4). Median concentrations of thiamethoxam exceeded the chronic HC₅ during at least one and up to four of the sampling years in the four waterbodies. Risk quotients calculated using median concentrations ranged from 0.5 to 4.6 (Table A.7-4).

Seven other waterbodies in the province of Quebec where the major land uses in the watersheds are mixed crops, as well as corn and soybeans, had concentrations of thiamethoxam exceeding the chronic toxicity endpoints in 11% to 70% of samples collected (Table A.7-3). In these waterbodies, the risk quotients calculated using the average concentrations and the chronic HC₅ ranged from 0.6 to 2.6 (Table A.7-4). The risk quotients calculated using the median concentrations ranged from 0.4 to 1.4.

Similarly to Quebec, several waterbodies located in watersheds in Ontario where row crops such as corn, soybeans and wheat are major components of the watershed showed thiamethoxam concentrations exceeding the chronic HC₅. In at least one year between 2012 and 2017, concentrations of thiamethoxam exceeded the chronic HC₅ of 0.026 µg/L in 25% or more of samples collected in the following eleven waterbodies: Twenty Mile Creek, Lebo Drain, Nottawagasa River, Sturgeon Creek, Sydenham River, Thames River, Big Creek, Garvey Glen, Little Ausable River, North Creek and White Ash Creek (Table A.7-5). In ten of these eleven waterbodies, the average concentration of thiamethoxam exceeded the chronic HC₅ in at least one, and up to five, years sampled between 2012 and 2017. Risk quotients calculated using average concentrations ranged from 0.2 to 14 (Table A.7-6). In Twenty Mile Creek, Lebo Drain, Sturgeon Creek, Sydenham River, Thames River, Big Creek, and North Creek, the median concentration of thiamethoxam exceeded the chronic HC₅ in at least one, and up to five, years sampled between 2012 and 2017. Risk quotients calculated using median concentrations in these waterbodies ranged from 0.3 to 6.7 between 2012 and 2017 (Table A.7-6). In these waterbodies, samples were collected approximately every one to two weeks from April to November. Some other waterbodies located in areas where corn, soybeans and wheat are grown in Ontario had detection frequencies of less than 25%, but the concentrations of thiamethoxam detected were higher than the chronic HC₅ of 0.026 μg/L. For example, thiamethoxam concentrations reported in Decker Creek, McGregor Creek and McKillop Drain were a maximum of 0.45 µg/L, 1.1 µg/L and 0.52 µg/L, respectively, between 2012 and 2014 (Table A.7-5). The limit of detection for samples collected in these waterbodies was 0.09 µg/L, which is more than two times higher than the chronic HC₅. The limit of detection could have a great influence on the detection frequencies; as such, the lower detection frequencies of thiamethoxam in these waterbodies may be at least partly a result of the higher limit of detection.

In British Columbia, thiamethoxam concentrations measured in Hope Slough exceeded the chronic HC_5 in 63% of the eight samples collected in 2015 (Table A.7-9). Corn is a major crop grown in the watersheds for this waterbody. Urban and forest areas are other major land uses in the watershed. The maximum concentration of thiamethoxam measured was 5.5 μ g/L. Risk quotients calculated using average and median concentrations were 30 and 3.4, respectively (Table A.7-10).

Thiamethoxam concentrations higher than the mesocosm NOEC were measured in waterbodies from corn and soybean areas of Ontario, Quebec and British Columbia. In Twenty Mile Creek, Ontario, between 21% and 33% of samples collected between 2013 and 2015 had thiamethoxam

concentrations higher than 0.3 µg/L (Table A.7-5). Samples were collected every two weeks between April and December at this site. The average concentration of thiamethoxam was 0.3 μg/L in 2013, resulting in a risk quotient of 1 (Table A.7-6). The risk quotients did not exceed the level of concern based on average concentrations in other years of sampling. However, the concentrations of thiamethoxam were higher than the mesocosm NOEC in three to five consecutive samples collected between late-May and late-July of the years 2013, 2014 and 2015. In North Creek, Ontario, the average concentration of thiamethoxam was 0.37 µg/L in 2015, resulting in a risk quotient of 1.2 (Table A.7-6). Thiamethoxam concentrations exceeded the mesocosm NOEC in four consecutive samples collected between June 1 and July 10, 2015. In the Saint-Régis River, Quebec, between 8% and 19% of samples collected from 2014 to 2017 had measured thiamethoxam concentrations higher than the mesocosm NOEC (Table A.7-3). The average concentration of thiamethoxam was 0.4 µg/L in 2015, resulting in a risk quotient of 1.3 (Table A.7-6). Between the years 2014 and 2017, thiamethoxam concentrations exceeding or approaching the toxicity endpoint were generally observed in two or three consecutive samples collected over approximately one week's time, between the months of June and August. Thiamethoxam was sporadically detected in several other waterbodies in corn and soybean areas of Quebec and Ontario at concentrations exceeding the mesocosm NOEC (Table A.7-3 and Table A.7-5). In British Columbia, thiamethoxam concentrations reported in Hope Slough exceeded the mesocosm NOEC in 25% of samples collected (Table A.7-9). One sample collected in August and one sample collected in December of 2015 exceeded the mesocosm NOEC. Risk quotients for Hope Slough were 2.6 and 0.3 when using the average and median concentrations, respectively (Table A.7-10). These results indicate that concentrations of thiamethoxam measured in waterbodies from corn and soybean areas of Canada can exceed the chronic endpoint for community-level effects for up to several weeks in some waterbodies.

Researchers have analyzed monitoring data and land use data in watersheds in southwestern Ontario for correlations between surface water monitoring detections and agricultural land uses in the watersheds. Concentrations of thiamethoxam measured in waterbodies from southwestern Ontario have been associated with corn, soybean and cereal grain crops in the areas surrounding the waterways [Struger et al. 2017 (PMRA# 2703534), PMRA# 2818731].

Seed treatments in Prairie Provinces

The primary use of neonicotinoids in the Prairies is as a seed treatment. Monitoring data indicate that concentrations measured in Prairie wetlands, rivers and creeks occasionally exceed chronic toxicity endpoints at different times throughout the season, particularly in the spring and summer.

Monitoring data were available for thiamethoxam in Prairie wetlands. The wetlands sampled were located in agricultural areas, near fields seeded to crops such as canola, barley, oats, wheat, field peas, lentils, soybeans, corn and grasslands. Most wetlands in the available datasets were sampled only once per sampling period, which consisted of spring/pre-seed, summer, or fall. As such, the PMRA did not generate chronic exposure estimates for these waterbodies. The percentage of wetlands with thiamethoxam concentrations exceeding the toxicity endpoints was determined for each sampling period. Risk quotients were calculated using the range of measured concentrations in all wetlands sampled to provide a broad estimate of the potential risks, assuming concentrations measured remained constant over time. There is uncertainty associated with longer-term exposure concentrations in the Prairie wetlands sampled.

Wetlands sampled in the spring prior to seeding exceeded the chronic HC₅ in 1% of the 138 wetlands sampled in 2012, 20% of the 90 wetlands sampled in 2013, and 6% of the 16 wetlands sampled in 2014. Concentrations measured in the spring ranged from below detection limits up to 0.11 µg/L. Risk quotients calculated using the range of concentrations measured spanned from less than 0.1 up to 4.1 (Table A.7-8). Main et al. (2016) reported that the presence of thiamethoxam in wetlands prior to seeding may be a result of the persistence of thiamethoxam residues in the soil and transport to wetlands via snowmelt and particulate matter during spring runoff (PMRA# 2572395).

Wetlands sampled in the summer exceeded the chronic HC₅ in 10% of the 134 wetlands sampled in 2012, 37% of the 144 wetlands sampled in 2013, 22% of the 46 relevant wetlands sampled in 2014 (Table A.7-7). In the summer of 2017, thiamethoxam was detected in nine of the 60 wetlands (15%) sampled in the three Prairie Provinces, at concentrations exceeding the HC₅ (three in Manitoba, four in Saskatchewan, and two in Alberta). Thiamethoxam concentrations measured in the summer ranged from below detection limits up to 1.5 μ g/L (Table A.7-7). Risk quotients calculated using the range of concentrations measured in the summer ranged from less than 0.1 up to 57 (Table A.7-8).

The mesocosm NOEC for possible community-level effects was exceeded in a few wetlands during the summer. Thiamethoxam concentrations were higher than $0.3 \mu g/L$ in 1% of the 134 wetlands sampled in the summer of 2012, in 3% of the 144 wetlands sampled in the summer of 2013, in 5% of the 115 wetlands sampled in the summer of 2014 and in 5% of the 60 wetlands sampled in the summer of 2017 (Table A.7-7). Risk quotients calculated using the mesocosm NOEC ranged from less than 0.1μ to 5 (Table A.7-8).

Thiamethoxam concentrations and detection frequencies in Prairie wetlands were generally lower in the fall compared to spring or summer (Table A.7-7). Some wetlands dried up during the season and thus sampling in the fall could not occur. Thiamethoxam concentrations exceeded the chronic HC_5 of $0.026~\mu g/L$ in one out of the 80 wetlands sampled in the fall of 2012. Thiamethoxam was not detected in any of the 23 wetlands sampled in the fall of 2017. It should be noted that there was widespread drought in the Canadian Prairies in 2017. The highest concentration of thiamethoxam measured in the fall was $0.1~\mu g/L$ in 2012. The risk quotients for wetlands sampled in the fall of 2012 ranged from less than $0.1~\mu$ to 3.8, when comparing concentrations with the chronic HC_5 .

In their research, Main et al. 2014 (PMRA# 2526133) reported that wetlands near canola fields typically had higher maximum neonicotinoid concentrations and higher detection frequencies than wetlands surrounded by grasslands. However, average neonicotinoid concentrations were not statistically different between wetlands near canola fields and those seeded to other crops such as barley, oats, peas, wheat and grassland. Similarly, Main et al. 2016 (PMRA# 2572395) found that wetlands located in oat fields not previously treated with neonicotinoids had similar thiamethoxam concentrations to wetlands found in previously treated canola fields. The authors report that this result may be due to persistence and carry-over of neonicotinoid residues between seasons, where neonicotinoid treated crops such as canola are frequently rotated with untreated crops, such as oats, in alternating years.

Ducks Unlimited Canada (PMRA# 2847073) reported that neonicotinoids were detected more often and at higher concentrations in Prairie wetlands where canola and wheat were the dominant crop types within a 250-metre area surrounding the wetlands. Neonicotinoid concentrations were also reported to vary between wetlands situated in the same field and surrounded by the same crop, possibly due to differences in preferential flow paths of the runoff and the size of contributing areas between the basins.

Based on the available monitoring data for thiamethoxam in Prairie wetlands, there is uncertainty associated with thiamethoxam concentrations over the growing season, as most wetlands were sampled only once per sampling period. Concentrations of thiamethoxam varied between the different sampling periods. However, in the study by Main et al. 2014 (PMRA# 2526133, 2612760), the same wetlands in Saskatchewan were generally sampled more than once, and up to four times, between the spring of 2012 and the spring of 2013. A total of 125 wetlands were sampled both in the spring of 2012 (between April 25 and May 1) and in the summer of 2012 (between June 23 and July 5). Of these, six wetlands (5%) had concentrations exceeding the chronic HC₅ on both occasions. A total of 55 wetlands were sampled during the four sample periods between the spring of 2012 and the spring of 2013. Of these, one wetland (2%) had concentrations of thiamethoxam exceeding the chronic HC₅ for all three consecutive sampling periods (summer 2012, fall 2012 and spring 2013). These results suggest that concentrations in some wetlands may exceed toxicity endpoints for several weeks to months. In addition, the 2017 season was a particularly dry year in the Canadian Prairies and there is uncertainty as to whether concentrations measured represent those that would be present in a more typical season.

Monitoring data for thiamethoxam in flowing waterbodies such as rivers and streams were available in Manitoba, Saskatchewan and Alberta. Thiamethoxam concentrations were generally lower in rivers and streams compared to those measured in Prairie wetlands. Many of the detections were below the chronic HC₅ of 0.026 µg/L. Nonetheless, thiamethoxam concentrations exceeded the chronic HC₅ in a few rivers and creeks. In the Red and Morris Rivers in Manitoba, and in Spirit Creek in Saskatchewan, thiamethoxam concentrations were higher than 0.026 µg/L in 29% to 67% of samples collected in some years (Table A.7-7). Major crops grown in the watersheds of the Red and Morris Rivers include soybeans, wheat, canola, oats, and corn, while canola and wheat represent the major crops grown in the watershed for Spirit Creek. Other sites sampled in the Prairie Provinces showed isolated detections of thiamethoxam above the chronic HC₅ of 0.026 µg./L. The maximum concentration reported in a flowing waterbody in the Prairies was 0.16 µg/L, from a sample collected in Moose Mountain Creek in Saskatchewan. Risk quotients for rivers and creeks ranged from less than 0.1 to 2 when using the average concentrations in a given year and ranged from less than 0.1 to 1.3 when using the median concentrations (Table A.7-8). The number of samples collected at many sites was small and results may not be representative of the entire season. Additionally, most of the monitoring data were from the year 2017, which was a particularly dry in the Canadian Prairies. There is uncertainty as to whether concentrations of thiamethoxam in rivers and streams would exceed toxicity endpoints for aquatic invertebrates when precipitation levels are more typical.

Orchards

Thiamethoxam is used as a foliar spray on orchard crops. Concentrations of thiamethoxam occasionally exceeded the chronic HC₅ of $0.026 \mu g/L$ in a few waterbodies located in areas

where orchards occupy large portions of the cultivated area of a watersheds in Ontario and Ouebec.

Rousse Creek and Déversant-du-Lac Creek are located in Quebec and were sampled in 2010, 2011, 2015 and 2016. Based on crop information presented in Giroux 2017 (PMRA# 2821394), orchards represented approximately 27% and 12.5% of the cultivated area of the watershed upstream of the sampling sites for Rousse Creek and Déversant-du-Lac Creek, respectively. Other crops also represented large portions of watersheds upstream of the sampling sites. In the Rousse Creek watershed, corn and soybeans represented a total of 16% of the cultivated area upstream of the sampling site, while vegetables represented 10%. In the Déversant-du-Lac Creek watersheds, corn and soybeans represented a total of about 65% of the cultivated area upstream of the sampling site, and cereal crops represented approximately 5%.

In Rousse Creek, thiamethoxam concentrations exceeded the chronic HC₅ in 7%, 14% and 23% of samples collected in 2011, 2015 and 2016, respectively (Table A.7-3). The yearly average concentration in Rousse Creek exceeded the chronic HC₅ during one of the four years sampled. The associated risk quotients ranged from less than 0.1 to 1 (Table A.7-4). Thiamethoxam concentrations in water exceeded the chronic HC₅ in 3% to 10% of samples collected in Déversant-du-Lac Creek in 2010, 2011, 2015 and 2016 (Table A.7-3). The yearly average concentration did not exceed the chronic HC₅ in any of the four years sampled. Associated risk quotients ranged from 0.2 to 0.6 based average concentrations and from less than 0.1 to 0.2 based on median concentrations (Table A.7-4). The instances when concentrations of thiamethoxam exceeded the chronic HC₅ were sporadic at both sites, and occurred between the months of May and August.

In Ontario, three waterbodies in areas where orchards occupy large portions of the watersheds showed thiamethoxam concentrations higher than the chronic HC₅ of 0.026 µg/L in some years between 2012 and 2016 (Table A.7-5). In Two Mile Creek, thiamethoxam exceeded the chronic HC₅ in 17% of the samples collected in 2016. Concentrations of thiamethoxam approached or exceeded the HC₅ sporadically in the months of May, June, August and November. In Four Mile Creek, a total of 14% to 33% of the samples had thiamethoxam concentrations higher than the HC₅ during the years 2012 to 2015. Samples were collected every two weeks between the months of April and December. Thiamethoxam concentrations close to or higher than the HC₅ were observed in up to three consecutive samples collected in April, June, July, September, October and November. In Prudhomme Creek, thiamethoxam concentrations exceeded the chronic HC₅ in 8% and 18% of samples collected in 2012 and 2013, respectively. Concentrations near or above the HC₅ were occasionally observed in the months of June, July and September. Risk quotients calculated using the average concentrations ranged from less than 0.1 to 0.5 for Two Mile Creek, from 0.1 to 1.2 for Four Mile Creek, and from 0.1 to 1.2 for Prudhomme Creek (Table A.7-6). Risk quotients calculated using the median concentrations ranged from less than 0.1 to a maximum of 0.6 for the three sites.

Due to the mixed uses within the watersheds sampled in Ontario and Quebec, it is not possible, based on the existing monitoring data, to identify which crops are contributing to the concentrations of thiamethoxam in these waterbodies.

While sampling was also conducted in 2017 in areas of British Columbia where orchards are present in watersheds, thiamethoxam was seldom detected in samples collected between June and September 2017 (Table A.7-9). Neonicotinoid use information for some watersheds in British Columbia (PMRA# 2842180) indicates that growers used neonicotinoids other than thiamethoxam to treat fruit trees in 2017. Therefore, the lack of detections of thiamethoxam in waterbodies where orchards are a major component of watersheds in British Columbia may be due to the absence of use.

Incident reports

Since 26 April 2007, registrants have been required by law to report pesticide incidents to the PMRA that are related to their products. In addition, the general public, medical community, government and non-governmental organizations are able to report pesticide incidents directly to the PMRA. The USEPA's Ecological Incident Information System (EIIS) was also queried for environmental incidents related to thiamethoxam that were available in that database up to February 2018. No incidents involving freshwater invertebrates have been reported in Canada or the United States related to clothianidin use (PMRA# 2852296).

4.5 Uncertainties Identified in the Risk Assessment

The PMRA has identified the following uncertainties in assessing thiamethoxam risk to aquatic invertebrates. These may be addressed in the future with the submission of additional data. However, the PMRA has determined that the risk conclusions presented are sound on the basis of the weight-of-evidence available with the chronic toxicity data, extensive surface water modelling that was conducted, and recent Canadian environmental monitoring data that were available.

4.5.1 Endpoints

The chronic SSD for thiamethoxam was based on a limited dataset of seven species, which is just above the minimum sample size of five for the construction of a species distribution as identified by Belanger et al. (2017) for use in regulatory risk assessment frameworks by global regulatory agencies. The PMRA distribution was statistically sound, meeting the criteria for normality of data. However, a wide confidence interval (CI) of approximately four orders of magnitude in the HC₅ value, indicates that the actual 5% effect level may lie over a wide range of values. The PMRA's HC₅ value of 0.026 μg a.i./L is an order of magnitude lower than the most sensitive chronic endpoint of 0.43 μg a.i./L (28-d EC₁₀ *C. dipterum*); however, it is very similar to a chronic endpoint for the protection of aquatic invertebrates from neonicotinoids of 0.035 μg/L that was recommended by Morrissey et al. (2015). Therefore, despite the wide confidence interval associated with this value, the PMRA has determined that the HC₅ is a valid endpoint for using in a weight-of-evidence approach to assessing chronic risk to the aquatic invertebrate community. Note that in making regulatory conclusions of risk for this proposed decision, the PMRA is using the higher-tiered community-level NOEC of 0.30 μg a.i./L that was based on effects that did not show evidence of recovery during a prolonged exposure period.

4.5.2 Exposure

Similarly to the endpoint selection, the PMRA uses a tiered approach to estimating exposure during a risk assessment which moves from a highly conservative screening level estimation to modelling estimation and finally to real-world monitoring data. Runoff is the primary route of exposure of thiamethoxam to aquatic invertebrates due to its solubility, high potential for movement into surface waters and persistence in waters with limited levels of sunlight penetration. At each step there are some uncertainties that are outlined below.

4.5.3 Modelling

Higher-tiered surface water runoff modelling was conducted for approximately half of the registered outdoor uses of thiamethoxam. Uses were chosen to ensure that runoff potential was assessed for a) representative application rates for each of the major application methods, and b) major crop uses across the country. For the uses of thiamethoxam that were not modelled, the acceptability of continued use of thiamethoxam cannot be demonstrated based on the range of modelled rates for each application method and the exceedance of the level of concern.

4.5.4 Monitoring

While monitoring data provide a real-life picture of the expected exposure concentrations, there were some areas where questions remain.

When considering the water monitoring data, the risk to aquatic invertebrates was assessed for thiamethoxam alone. Neonicotinoids share a common mode of action and have been shown to co-occur in many Canadian waterbodies [Main et al. 2014 (PMRA# 2526133); Main et al. 2015 (PMRA# 2608629); Main et al. 2016 (PMRA# 2572395); Struger et al. 2017 (PMRA# 2703534); Giroux 2014 (PMRA 2544468); Giroux 2015 (PMRA# 2561884); Giroux 2017 (PMRA# 2821394)]. As such, the potential risk from the combined residue is unknown but, the potential risk will be higher in waterbodies containing two or more neonicotinoids than that when the individual neonicotinoids are considered alone.

Thiamethoxam transforms in soils to clothianidin, another neonicotinoid insecticide, which is also toxic to aquatic invertebrates. Detections of clothianidin in water may be a result of the use of thiamethoxam, the use of clothianidin as an insecticide, or a combination of the two. The potential contribution of clothianidin from the transformation of thiamethoxam is not possible to estimate at this time.

Regarding acute exposure, monitoring data likely underestimate short-term exposure to thiamethoxam, as most sampling regimes are unlikely to capture peak concentrations.

Not all regions across Canada are represented equally in a variety of ways. Sampling regimes differ between datasets in different regions; some waterbodies were only sampled a few times during the season resulting in some uncertainty as to the duration of exposure in these areas and some areas of Canada lack water monitoring. In areas where thiamethoxam is used but monitoring data are lacking, there is no reason to believe that detection patterns would differ compared to those observed in areas where monitoring data are available.

Relating thiamethoxam concentrations in water to use on a specific crop is difficult in watersheds where multiple thiamethoxam-treated crops are common. Similarly, it is difficult to relate thiamethoxam concentrations in water to a specific application method in watersheds where the crops grown can be treated using multiple methods (for example, potatoes can be treated using foliar spray, soil application or seed treatment, certain vegetable crops can be treated using either a seed treatment or a foliar spray, while other vegetables crops can be treated using either a seed treatment or a soil application).

In some cases there is limited site information, such as some temporary wetlands, therefore, the relevancy for an aquatic invertebrates risk assessment was difficult to determine. In the absence of additional information, these were considered relevant water bodies in this assessment.

The weather patterns across Canada in 2017 were unusually dry in some areas, especially in the Prairies. This dry year may have affected the concentrations detected in these areas.

Samples showing no detections can be difficult to interpret, particularly when the limit of detection is high, and when use information in the vicinity of sampling areas is not available. The non-detects could be due to factors such as the non-transport of the chemical from the site of application, the lack of use of the chemical in the area studied, or the lack of sensitivity of the analytical method.

4.6 Risk Assessment Conclusions

Surface water modelling of thiamethoxam uses showed widespread exceedences of the level of concern for chronic effects to aquatic invertebrates. The modelling was region-specific, and it encompassed a wide range of crops and application methods across Canada. Recent water monitoring data show that thiamethoxam is being detected in Canadian surface waters at concentrations that exceed the level of concern for chronic adverse effects on aquatic invertebrates. Concentrations that may impact individual species and invertebrate communities occurred from days to weeks in some waterbodies associated with agricultural uses of thiamethoxam. This assessment is based on the exposure of thiamethoxam alone to aquatic invertebrates, whereas neonicotinoids have been shown to co-occur in the environment and share a common mode of action. Thiamethoxam can also transform to clothianidin in soils, which is then available for runoff to surface waters. Thus, the impact of exposure to multiple neonicotinoids will be higher than for exposure to thiamethoxam alone.

Therefore, based on the available information the PMRA is unable to conclude that the risks to aquatic invertebrates are acceptable from outdoor agricultural uses of thiamethoxam. The risks to aquatic invertebrates associated with greenhouse uses of thiamethoxam are acceptable, provided wastewater mitigation instructions on product labels are followed. The PMRA acknowledges that

research on neonicotinoids is ongoing and scientific studies are published regularly. Relevant information that became available after the initiation of the PMRA's publication process and any information submitted during the consultation period will be considered by the PMRA before making a final decision.

4.7 Risk Mitigation for Aquatic Invertebrates

4.7.1 Use Restrictions

Given the risks that have been identified and considering the available information, effective risk mitigation through a use-reduction strategy would be difficult to achieve for several reasons. In mixed-use areas of agriculture, it would be difficult to identify inputs from specific crops or application methods causing the elevated concentrations seen in water. In addition, it is not possible to accurately predict how much use reduction would be necessary to achieve acceptable concentrations of thiamethoxam in the environment and, therefore, any use-reduction strategy would require extensive and comprehensive water monitoring information to confirm that risk reduction targets are being achieved. It is also not possible to estimate how long a reduction in environmental concentrations would take. In addition, in sectors where thiamethoxam is approved for use but not currently used extensively, intensification of uses in the future may lead to additional risks of concern. Given the above, cancellation of all outdoor agricultural and ornamental uses for thiamethoxam is being proposed. At this time, cancellation of indoor use in greenhouses is not proposed, provided appropriate wastewater mitigation measures are followed.

4.7.2 Spray Buffer Zones

During the phase out period, updated spray buffer zones based on the risks identified in this assessment will be required for the protection of freshwater habitats. Spray buffer zones for terrestrial habitats are also required as per existing conditions of use. Spray buffer zones were determined based on existing directions for use on product labels, including a spray quality of ASAE Fine for field and aerial sprayers. The complete proposed spray buffer zone table and drift mitigation instructions for thiamethoxam products are provided in Appendix VIII.

Agriculture and Agri-Food Canada's Multi-stakeholder Mitigation Working Group submitted information on recommended drift mitigation strategies which included:

- promoting the use of best management practices for minimizing spray drift,
- promoting the adoption of the PMRA's on-line spray buffer zone calculator tool, and
- increasing label restrictions for foliar spray applications to minimize spray drift.

As for all pest control products, during the phase out period for agricultural uses of thiamethoxam, the PMRA will continue to encourage the adoption of best management practices for spray drift management. Required drift mitigation measures for specific application methods will be identified on product labels. At this time, additional application restrictions to minimize spray drift are not required. With the exception of identified buffer zones of 800 m for aerial use on potatoes, soybeans and dry beans, the on-line spray buffer zone calculator can be used to further mitigate the potential for spray drift based on the use of coarser spray qualities and by accounting for meteorological conditions at the time of application.

4.7.3 Runoff Mitigation

Precautionary label statements are currently on all product labels to reduce the potential for runoff to adjacent aquatic habitats. Despite the current label statement, concentrations of thiamethoxam posing a risk to aquatic invertebrates have been found in Canadian surface waters where clothianidin is used for pest management in agriculture.

Agriculture and Agri-Food Canada's Multi-stakeholder Mitigation Working Group submitted information on the potential use of vegetative filter strips to reduce runoff into adjacent waterbodies. While studies exist on the effectiveness of vegetative filter strips at reducing runoff of pesticides, most of the research has been conducted using pesticides that are much less water soluble than neonicotinoids. Only two studies were conducted using neonicotinoids, namely those by Denning et al. 2004 (PMRA# 2518467) and Hladik et al. 2017 (PMRA# 2866915) and the results of both studies as to the potential effectiveness of vegetative filter strips to reduce surface water runoff of neonicotinoids were inconclusive. In both studies, neonicotinoid concentrations in surface water runoff were variable and they were not significantly different or were higher at sites with vegetative filter strips compared to sites without them. Field dynamics and/or input from nearby neonicotinoid-treated fields that were not a part of the study confounded the results. No quantifiable measure to reduce the runoff of neonicotinoids into waterbodies using vegetative filter strips could be derived from the two studies. Notwithstanding the lack of quantifiable risk reduction, the PMRA will continue to include the standard recommended label statement for the use of vegetative filter strips on thiamethoxam product labels as part of a runoff mitigation strategy.

5.0 Proposed Special Review Decision for Thiamethoxam

The evaluation of available scientific information related to the aspects of concern indicated that the registered products containing thiamethoxam pose environmental risks that have not been shown to be acceptable. Therefore, under the authority of the *Pest Control Products Act* and based on the evaluation of currently available scientific information, Health Canada is proposing to cancel all outdoor uses of thiamethoxam on food and feed crops (use site categories 13 and 14), including seed treatments (use site category 10), and outdoor ornamentals (use site category 27), over three to five years, in accordance with the provisions outlined in Regulatory Directive DIR2018-01, *Policy on Cancellations and Amendments Following Re-evaluation and Special Review*. The PMRA will consider alternate risk management proposals, provided that they can achieve acceptable levels in the environment within the same timeframe.

The risks to aquatic invertebrates associated with greenhouse uses of thiamethoxam (use site categories 5 and 6) are acceptable and continue registration of these greenhouse uses is proposed, provided wastewater mitigation instructions on product labels are followed.

Additional mitigation measures may be required during the phase-out period (Appendix VIII).

The proposed special review decision is open for public consultation for 90 days from the date of this publication. The PMRA is inviting the public to submit comments on the proposed special review decision for thiamethoxam including proposals that may refine the risk assessment and

risk management. Once the PMRA considers the comments and any information that are received during the public consultation period, the Agency will publish a final decision.

6.0 Next Steps

Before making a special review decision on thiamethoxam, the PMRA will consider all comments received from the public in response to this consultation document. A science-based approach will be applied in making a final decision on thiamethoxam. The PMRA will then publish a special review decision document, which will include the decision, the reasons for it, a summary of the comments received on the proposed decision and the PMRA's response to these comments.

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List of Abbreviations

< less than > greater than

≤ less than or equal to≥ greater than or equal to

 $\begin{array}{ll} \mu g & \text{microgram}(s) \\ \mu M & \text{micromolar} \end{array}$

1/n exponent for the Freundlich isotherm

a.i. technical active ingredient

ASAE American Society of Agricultural and Biological Engineers

atm atmosphere(s)

CAS chemical abstracts service

CG crop group

CI confidence interval cm centimeter(s)

DFOP double first order in parallel

 DT_{50} dissipation time 50% (the time required to observe a 50% decline in

concentration)

DT₉₀ dissipation time 90% (the time required to observe a 90% decline in

concentration)

dw dry weight

EC₁₀ effective concentration on 10% of the population EC₂₀ effective concentration on 20% of the population ECCC Environment and Climate Change Canada

EEC estimated environmental concentration

EP end-use product

FA fraction of species affected

g gram(s)
h hour(s)
ha hectare(s)

HC₅ hazardous concentration estimate that is assumed to be protective of 95% of

species in a species sensitivity distribution

Hg mercury

HPLC high performance liquid chromatography

hr(s) hour(s)

IORE Indeterminate Order Rate Equation Model

IUPAC International Union of Pure and Applied Chemistry

K_d soil-water partition coefficient
 K_F Freundlich adsorption coefficient

kg kilogram(s)

 K_{oc} organic-carbon partition coefficient K_{ow} octanol-water partition coefficient

L litre(s)

LC₁₀ lethal concentration on 10% of the population

LC₅₀ median lethal concentration

LOEC lowest observed effect concentration

LOD limit of detection

LOQ limit of quantitation

m metre(s)
mg milligram(s)
min minute(s)
mL millilitre(s)
mm millimitre(s)
MoA Mode of Action
MOE margin of exposure

mRNA messenger ribonucleic acid

MS mass spectrometry

N sample size
NA not applicable
NC not calculated
ND not detected
ng nanogram(s)

NOAEC no observed adverse effect concentration

NOAEL no observed adverse effect level NOEC no observed effect concentration

NOEL no observed effect level

NR not reported N/R not required

OC organic carbon content
OM organic matter content
PCPA Pest Control Product Act
PCP Pest Control Product number

pKa dissociation constant

PMRA Pest Management Regulatory Agency

ppb parts per billion
ppm parts per million
RQ risk quotient
SFO single first order
sp. species (singular)
spp. species (plural)

SSD Species Sensitivity Distribution

Stdev standard deviation

t_{1/2} half-life

TGAI technical grade active ingredient

t_R representative half-life TWA time weighted average

USEPA United States Environmental Protection Agency

UV ultraviolet wt(s) weight(s)

WWTP waste water treatment plant

Appendix I Registered Thiamethoxam Products as of May 2018 that are subject to this Special Review, Excluding Discontinued Products or Products with a Submission for Discontinuation

Registration Number	Marketing Class	Registrant	Product Name	Formulation Type	Guarantee
26665	Technical	Syngenta Canada Inc.	Thiamethoxam Technical	Dust or powder (solid)	99.1%
26637	Commercial		Helix Liquid Seed Treatment	Suspension	Thiamethoxam 10.3%; metalaxyl-M and S isomer 0.39%; fludioxonil 0.13%; difenoconazole 1.24%
26638			Helix Xtra Seed Treatment		Thiamethoxam 20.70%; metalaxyl-M and S isomer 0.39%; fludioxonil 0.13%; difenoconazole 1.25%
27045	-		Cruiser 5FS Seed Treatment		Thiamethoxam 47.6%
27986			Cruiser 350FS Seed Treatment Insecticide		Thiamethoxam 29.9%
28407			Actara 240SC Insecticide		Thiamethoxam 240 g/L
28408			Actara 25WG Insecticide	Wettable granules	Thiamethoxam 25.0%
28821			Cruiser Maxx Beans Seed Treatment	Suspension	Thiamethoxam 22.6%; metalaxyl-M and S isomer 1.70%; fludioxonil 1.12%
29127			Cruiser Maxx Cereals Commercial Seed Treatment	Suspension	Thiamethoxam 2.8%; metalaxyl-M and S isomer 0. 56%; difenoconazole 3.36%
29192			Cruiser Maxx Cereals Seed Treatment		Thiamethoxam 2.8%; metalaxyl-M and S isomer 0. 56%; difenoconazole 3.36%
30388			A18046A Seed Treatment		Thiamethoxam 261 g/L; metalaxyl-M and S isomer 19.7 g/L; fludioxonil 12.9 g/L; azoxystrobin 10.4g/L
30404			Endigo Insecticide		Thiamethoxam 141 g/L; lambda-cyhalothrin 106 g/L
30436			Cruiser Maxx Vibrance Cereals Seed Treatment		Thiamethoxam 30.7g/L; sedaxane 8.0 g/L; metalaxyl-M and S isomer 9.5 g/L; difenoconazole 36.9 g/L
30723			Flagship Insecticide	Wettable	Thiamethoxam 25%
30900			Minecto Duo 40WG	granules	Thiamethoxam 20%; cyantraniliprole 20%

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Registration Number	Marketing Class	Registrant	Product Name	Formulation Type	Guarantee
30901			Mainspring X		Thiamethoxam 20%;
			Insecticide		cyantraniliprole 20%
31024			Cruiser Maxx Potato	Suspension	Thiamethoxam 250 g/L;
			Extreme		fludioxonil 62.5 g/L;
					difenoconazole 123 g/L
31453			Cruiser Vibrance		Thiamethoxam 61.5 g/L
			Quattro		Difenoconazole 36.9 g/L
					Metalaxyl-M and S-Isomer
					9.2 g/L
					Sedaxane 15.4 g/L
					Fludioxonil 7.7 g/L
31454			Helix Vibrance		Thiamethoxam 269 g/L
					Difenoconazole 16 g/L
					Metalaxyl-M and S-Isomer
					5 g/L
					Sedaxane 3.4 g/L
					Fludioxonil 1.7 g/L

Appendix II Registered Commercial Class Uses of Thiamethoxam in Canada as of May 2018 that are subject to this Special Review

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
5	Greenhouse Peppers	Pepper weevil	Wettable granules	Ground application: foliar spray handwand, backpack sprayers	3.5 g a.i./100L {70 g a.i./ha}	12/year - 3 applications per crop cycle	7
6	Greenhouse ornamentals	Aphids, dipteran leafminers, mealybugs, soft scales, thrips whiteflies	Wettable granules	Ground application: foliar spray handwand, backpack sprayers	7.5 - 15 g a.i./100L 75 - 150 g a.i./ha	8/year -2 application per crop cycle	14
6	Greenhouse ornamentals	Aphids, dipteran leafminers, mealybugs, soft scales, fungus gnats, root aphids, whiteflies, thrips	Wettable granules	Soil drench	10 - 15 g a.i./100L {200 - 300 g a.i./ha/crop cycle}	4/year -1 application per crop cycle	Not applicable
10	Barley, wheat	Wireworms, European chafer	Suspension	On farm and /or commercial seed treatment facility: seed treatment equipment	9.98 - 30 g a.i./100 kg seed {barley: 6.7 - 36.3 g a.i./ha} {wheat: 6.7 - 52.5 g a.i./ha}	1	Not applicable
10	Oats	Wireworms	Suspension	On farm and /or commercial seed treatment facility: seed treatment equipment	9.98 - 19.98 g a.i./100 kg seed {5.4 - 22.8 g a.i./ha}	1	Not applicable

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
10	Buckwheat, millet, sorghum, rye, triticale	Wireworm	Suspension	On farm and/ or commercial seed treatment facility: seed treatment equipment	10 - 30 g a.i./100 kg seed {buckwheat 3.9 - 24.3 g a.i./ha}, {millet 0.55 - 6.7 g a.i./ha}, {sorghum 0.9 - 4.5 g a.i./ha}, {rye 6.2 - 20.2 g a.i./ha}, triticale {8.1 - 63.0 g a.i./ha}	1	Not applicable
10	Bean (dry)	Potato leafhopper, seedcorn maggot	Suspension	Commercial seed treatment facility: seed treatment equipment	30 - 50 g a.i. /100 kg seed {9.6 - 41.5g a.i./ha}	1	Not applicable
10	Bean (dry)	wireworm	Suspension	Commercial seed treatment facility: seed treatment equipment	50 g a.i. /100 kg seed {41.5 g a.i./ha}	1	Not applicable
10	Canola, rapeseed, mustard	Flea beetles	Suspension	Commercial seed treatment facility: seed treatment equipment	199.4 - 403.5 g a.i./100 kg seed {canola/rapeseed 8.0 - 32.3 g a.i./ha},{mustard 9.0 - 45.2 g a.i./ha}	1	Not applicable
10	Chickpeas, faba bean, lentils, lupins, dry peas	Wireworm	Suspension	On farm and/ or commercial seed treatment facility: seed treatment equipment	10 - 30 g a.i./100 kg seed {chickpea 9.0 - 46.5 g a.i./ha}, {faba bean 26.6 - 111.6 g a.i./ha}, {lentils 4.5 - 27.0 g a.i./ha}, {lupins 15.7 - 50.4 g a.i./ha}, {dry peas 10.4 -83.1 g a.i./ha}	1	Not applicable
10	Faba bean	Pea leaf weevil	Suspension	On farm and/ or commercial seed treatment facility: seed treatment equipment	30 g a.i./100 kg seed {79.8 – 111.6 g a.i./ha}	1	Not applicable

USC1	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
10	Com (Field, seed, sweet, popcom)	European chafer, wireworm	Suspension	Commercial seed treatment facility: seed treatment equipment	50 g a.i./100 kg seed {field 7.9 – 11.8 g a.i./ha}, {sweet 5.3- 7.6 g a.i./ha}	1	Not applicable
10	Corn (Field, seed, sweet, popcorn)	Seedcorn maggot, corn flea beetle	Suspension	Commercial seed treatment facility: seed treatment equipment	50 - 100 g a.i./100 kg seed {field 7.9 - 23.7 g a.i./ha}, {sweet 5.3-15.1 g a.i./ha}	1	Not applicable
10	Corn (Field, seed, sweet, popcorn	Corn rootworm	Suspension	Commercial seed treatment facility: seed treatment equipment	200 - 500 g a.i./100 kg seed {field 78.75 -118.3 g a.i./ha}, {sweet 52.5-75.6 g a.i./ha}	1	Not applicable
10	Pea (dry)	Pea leaf weevil	Suspension	On farm and/ or commercial seed treatment facility: seed treatment equipment	30 - 50 g a.i./100 kg seed {31.2 - 138.5 a.i./ha}	1	Not applicable
10	Potato	Aphids, Colorado potato beetle, potato leafhopper	Suspension	Seed piece treatment equipment: slurry	1.9 - 5.86 g a.i./100 kg seed {91.2 or 117 g a.i./ha}	1	Not applicable
10	Soybean	Seedcorn maggot	Suspension	Commercial seed treatment facility: seed treatment equipment	30 - 50.8 g a.i./100 kg seed {17.1 - 63.0 g a.i./ha}	1	Not applicable
10	Soybean	Bean leaf beetle, European chafer, soybean aphid, wireworm	Suspension	Commercial seed treatment facility: seed treatment equipment	50.8 g a.i./100 kg seed {28.5 – 64.g a.i./ha}		Not applicable

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
10	Succulent beans, succulent peas	Potato leafhopper, seedcorn maggot	Suspension	Commercial seed treatment facility: seed treatment equipment	30 - 50 g a.i./100 kg seed {beans 22.5 - 50.0 g a.i./ha}, {pea 30-150 g a.i./ha}	1	Not applicable
10	Succulent beans, succulent peas	Wireworm, soybean aphid	Suspension	Commercial seed treatment facility: seed treatment equipment	50 g a.i./100 kg seed {37.5 - 50.0 g a.i./ha}, {pea 50-150 g a.i./ha}	1	Not applicable
10	Succulent peas	Pea leaf weevil	Suspension	Commercial seed treatment facility: seed treatment equipment	30 - 50 g a.i./100 kg seed {30 - 150 g a.i./ha}	1	Not applicable
10	Sunflowers – importation of treated seeds	Wireworm	Suspension	Not applicable - treated prior to import	0.25 mg a.i./seed {4.4 - 27.5 g a.i/ha}	1	Not applicable
10	Sugar beet	Wireworm, sugar beet root maggot	Suspension		30 - 60 g a.i./100,000 seeds {19.5 - 58.7 g a.i./ha}	1	Not applicable
10	Crop Group 9 Cucurbit Vegetables	Cucumber beetle	Suspension	Not applicable – imported seeds only	0.25 - 0.75 mg a.i./seed {cucumber: 4.6 - 20.8 g a.i./ha}, {pumpkin/squash 0.56 - 8.3 g a.i./ha}	1	Not applicable
13,14	Apple, crab apple	Plum curculio, mullein bug	Water dispersible granule	Ground application: Foliar spray - airblast	78.75 g a.i./ha (pre- bloom) 78.75 - 96.25 g a.i./ha (post bloom)	2 (1 pre-bloom and 1 post bloom or 2 post bloom applications)	10
13,14	Apple, crab apple	Spotted tentiform leafminer	Water dispersible granule	Ground application: Foliar spray - airblast	78.55 g a.i./ha (pre and post bloom)	2 (1 pre-bloom and 1 post bloom or 2 post bloom applications)	10

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
13,14	Apple, crab apple	Rosy apple aphid	Water dispersible granule	Ground application: Foliar spray - airblast	40 g a.i./ha	(1 pre-bloom and 1 post bloom or 2 post bloom applications)	10
14	Pear, Oriental pear	Pear psylla, plum curculio	Water dispersible granule	Ground application: Foliar spray - airblast	78.75 - 96.25 g a.i./ha	2 (post bloom only)	10
13,14	Apple, crab apple, pear, Oriental pear	Brown marmorated stink bug	Water dispersible granule	Ground application: Foliar spray - airblast	96.25 g a.i./ha	2 (post bloom only)	10
14	Cherries (sweet and sour)	Aphids	Water dispersible granule	Ground application: Foliar spray - airblast	40 g a.i./ha	2	10
14	Bean (dry) (Phaseolus spp., Lupinus spp., Vigna spp., dry fava beans, dry lablab beans and chickpeas, soybean)	Bean leaf beetle, Soybean aphid	Suspension	Aerial application: Foliar spray - rotary and fixed wing Ground application: Foliar spray conventional ground equipment	25.38 g a.i./ha	3	7
14	Pepper	Pepper weevil	Water dispersible granule	Foliar spray conventional ground equipment	70 g a.i./ha	2	7
14	Celeriac	Tarnished plant bug	Water dispersible granule	Foliar spray conventional ground equipment	52.5 - 70 g a.i./ha	2	Not stated

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
13,14	Potato	Aphids, Colorado potato beetle, potato leafhopper	Suspension	Ground application : In- furrow drench - ground equipment	0.82 - 1.06 g a.i./100m of row 37.9 - 140 g a.i./ha based upon row spacing of 215 cm to 75 cm	1	Not applicable
13,14	Potato	Aphids, Colorado potato beetle, potato leafhopper	Suspension	Foliar spray conventional ground equipment Aerial application: Foliar spray - rotary and fixed wing	26.2 g a.i./ha	2	7
13,14	Potato	Aphids, Colorado potato beetle, potato leafhopper	Water dispersible granule	Foliar spray conventional ground equipment Aerial application: Foliar spray - rotary and fixed wing	26.25 g a.i./ha	2	7
13,14	Potato	Aphids, Colorado potato beetle, flea beetles, potato leafhopper	Wettable granule	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation	88 - 140 g a.i./ha 0.66 - 3.2 g a.i./100m of row	1	Not applicable
14	Crop Group 1B and 1C Root vegetables	Aphids, Aster leafhopper	Water dispersible granule	Foliar spray conventional ground equipment	26.25 g a.i./ha	2	7
14	Crop Group 4 Leafy vegetables	Aphids	Water dispersible granule	Foliar spray conventional ground equipment	26.25 g a.i./ha	2	7
14	Crop Group 4 Leafy vegetables	Tarnished plant bug	Water dispersible granule	Foliar spray conventional ground equipment	52.5 g a.i./ha	1	Not applicable

USC1	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
14	Crop Group 4 Leafy vegetables	Aphids, dipteran leafminers, leafhoppers, cabbage looper, flea beetle, beet armyworm, corn earworm, fall armyworm	Wettable granule		150 g a.i./ha 0.23 - 4.5 g a.i./100m of row	1	Not applicable
14	Crop Group 5 Brassica vegetables	Aphids, dipteran leafminers, flea beetles, cabbage looper, diamondback moth, imported cabbageworm thrips, beet armyworm, corn earworm, fall armyworm, yellowstripped armyworm	Wettable granule	Ground application: in-	150 g a.i./ha 0.23 - 4.5 g a.i./100m of row	1	Not applicable
14	Crop Group 8 Fruiting vegetables	Aphids, Colorado potato beetle, dipteran leafminers, leafhoppers, potato psyllid cabbage looper, flea beetles, thrips, beet armyworm, corn earworm, fall armyworm, tomato fruitworm, yellowstripped armyworm	Wettable granule	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation	48.5- 146.8 g a.i./ha 0.13 - 4.5 g a.i./100m of row	1	Not applicable

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
14	Crop Group 9 Cucurbit vegetables	Aphids, leafminers, leafhoppers, cucumber beetles, flea beetles,thrips	Wettable granule	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation		1	Not applicable
14	Crop Group 4 Leafy vegetables	Aphids, leafhoppers, dipteran leafminers, flea beetle	Suspension	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation	90 - 150 g a.i./ha	1	Not applicable
14	Crop Group 5 Brassica vegetables	Aphids, flea beetle	Suspension	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation	90 - 150 g a.i./ha	1	Not applicable
14	Crop Group 8-09 Fruiting vegetables	Aphids, Colorado potato beetles, leafhoppers, dipteran leafminers, potato psyllids, flea beetle	Suspension	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation	90 - 150 g a.i./ha	1	Not applicable
14	Crop Group 9 Cucurbit vegetables	Aphids, leafhoppers, dipteran leafminers, flea beetle	Suspension	Ground application: in- furrow drench – ground equipment or surface band drench + irrigation	90 - 150 g a.i./ha	1	Not applicable
14	Crop Group 8 Fruiting vegetables	Aphids	Water dispersible granule	Ground application: foliar spray – ground equipment (over the row sprayer)	26.25 g a.i./ha	2	7

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
14	Crop Group 8 Fruiting vegetables	tarnished plant bug, stink bug	Water dispersible granule	Ground application: foliar spray – ground equipment (over the row sprayer)	26.25 - 52.5 g a.i./ha	2	7
14	Crop Group 8 Fruiting vegetables	brown marmorated stink bug	Water dispersible granule	Ground application: foliar spray – ground equipment (over the row sprayer)	52.5 g a.i./ha	2	7
14	Crop Group 8 Fruiting vegetables	Aphids, Tarnished plant bug, stink bugs	Water dispersible granule	Ground application: in- furrow drench- conventional ground equipment	0.85 - 1.1 g a.i./100m of row 48.5 - 146.8 g a.i./ha	1	Not applicable
14	Crop Group 8 Fruiting vegetables	Aphids, Tarnished plant bug, stink bugs	Water dispersible granule	Ground application: transplant water application	91.25 - 117 g a.i./ha at 30 000 plants/ha	1	Not applicable
14	Crop Group 13-07A Cane berries	Black vine weevil obscure root weevil	Water dispersible granule	Ground application: foliar spray – ground equipment (over the row sprayer)	52.5 - 70 g a.i./ha	2	7
14	Crop Group 13-07B Bush berries	Black vine weevil, obscure root weevil	Water dispersible granule	Ground application: foliar spray – ground equipment (over the row sprayer)	52.5 - 70 g a.i./ha	2	7
14	Crop Group 13-07B Bush berries	Brown marmorated stink bug	Water dispersible granule	Ground application: foliar spray – ground equipment (over the row sprayer)	70 g a.i./ha	2	7

USC ¹	Site(s) ²	Pest(s)	Formulation Type	Application Methods and Equipment	Application Single Rate or Rate Range	Maximum Number Applications per Year	Minimum Interval Between Applications (Days)
14	Crop Group 13-07G Low growing Berries	Adult black vine weevil, Cranberry weevil	Water dispersible granule	Ground application: foliar spray – ground equipment (boom sprayer)	52.5 - 70 g a.i./ha	2	7
14	Crop Group 13-07G Low growing Berries	Black vine weevil, strawberry root weevil	Water dispersible granule	Ground application: soil drench - post renovation	140 g a.i./ha	1	Not applicable
27	Outdoor ornamentals	aphids, black vine weevil, dipteran leafminers, lace bugs, leafhoppers, mealybugs, psyllids, soft scales, thrips	Wettable granules	Ground application equipment - Foliar application	7.5 - 15 g a.i./100L 75 - 150 g a.i./ha	1 at high rate or 2 at low rate	14
27	Viburnum	Viburnum leaf beetle	Water dispersible granule	Ground application: foliar spray – ground equipment	70 g a.i./ha	1	Not applicable
27	Outdoor ornamentals	Black vine weevil	Water dispersible granule	Ground application: foliar spray – ground equipment	2.63 - 3.5 g a.i./100L Maximum of 70 g a.i./ha in 2000 L/ha	(2)	7
27	Outdoor ornamentals	Aphids, leafhoppers	Water dispersible granule	Ground application: foliar spray – ground equipment	26.25 g a.i./ha	(2)	7
27	Outdoor ornamentals	Tarnished plant bug	Water dispersible granule	Ground application: foliar spray – ground equipment	52.5 - 70 g a.i./ha	(2)	7
27	Outdoor nurseries and landscapes	Brown marmorated stink bug	Water dispersible granule	Ground application: foliar spray – ground equipment	70 g a.i./ha	(1)	Not applicable

Use Site Category (USC): 5 - Greenhouse Food crops, 6 - Greenhouse Non-food crops, 10 - Seed and Plant Propagation Materials Food and Feed, 13 - Terrestrial Feed Crops, 14 - Terrestrial Food Crops, 27 - Ornamentals Outdoors.

Crop groups are identified as listed on the end use product labels and may not be identical to the crop groups listed on the Health Canada Residue Chemistry Crop Groups website: http://hc-sc.gc.ca/cps-spc/pest/part/protect-proteger/food-nourriture/recg-gcpcr-eng.php

Appendix III Fate, Toxicity, and Risks to the Aquatic Invertebrates

Table A.3-1 Identity of active substance thiamethoxam

Active Substance Thiamethoxam

(Development Code: CGA 293343)

Function Insecticide

Chemical name

3-(2-chloro-1,3-thiazol-5-ylmethyl)-5-methyl-1,3,5-oxadiazinan-4-1. International Union of Pure and Applied Chemistry ylidene(nitro)amine

(IUPAC)

2. Chemical Abstract Services 3-[(2-Chloro-5-thiazolyl)methyl]tetrahydro-5-methyl-N-nitro-4H-

(CAS) 1,3,5-oxadiazin-4-imine

CAS Number 153719-23-4 Molecular Formula $C_6H_{10}CIN_5O_3S$ Molecular Weight 291.7 g/mol

Structural Formula

Position of Radiolabels in **Environmental Studies**

[Guanidine-4-14C] Thiamethoxam* * Also referred to as

[Oxadiazine-4-14C] Thiamethoxam

[Thiazoly1-2-14C] Thiamethoxam

Table A.3-2 Physical and chemical properties of thiamethoxam relevant to the environment

Property	Value	Comments ¹
Solubility in water	4.1 g/L at 25°C	Very soluble in water.
Vapour pressure	2.7×10^{-9} Pa at 20°C	Low volatility.
	6.6 × 10 ⁻⁹ Pa at 25°C	
Henry's law constant	$1.9 \times 10^{-10} \text{Pa} \cdot \text{m}^3 / \text{mole at } 20^{\circ} \text{C}$	Non-volatile from water and moist soil
	(equivalent to 1.9×10^{-15} atm·m ³ / mole)	surface.
	$4.7 \times 10^{-10} \mathrm{Pa\cdot m^3}$ / mole at 25°C	
	(equivalent to 4.7×10^{-15} atm·m ³ / mole)	
Ultraviolet (UV) / visible	No absorption at wavelengths greater	Minimal phototransformation expected
spectrum	than 300 nm.	in the natural environment.
Octanol/water partition	$\log K_{ow} = -0.13 \text{ at } 25^{\circ}C$	Low potential for bioaccumulation.
coefficient (K _{ow})		
Dissociation constant (p K_a)	None within the range of pH 2 to pH 12	No dissociation at environmentally
		relevant pH.

¹ Source: ERC2011-05

Table A.3-3 Octanol-water partition coefficients for thiamethoxam transformation products

Transformation Product	Value	Comments	Reference
CGA 322704	$\log K_{ow} = 0.84$ (estimated)	Low potential for bioaccumulation.	PMRA# 1529715 Tier II Summary (prepared
CGA 355190	$\log K_{ow} = 1.2$ (estimated)	bioaccumulation.	by the registrant)
CGA 355190	$\log K_{ow} = 0.84$ (measured)		PMRA# 1529718

Table A.3-4 Fate and behaviour in the terrestrial environment

Property	Test substance	Value	Comments	Study	
	Abiotic transformation				
Hydrolysis	Thiamethoxam	At 25°C: t½ pH 5: stable t½ pH 7: 559 - 939 days t½ pH 9: 4.1 - 8.0 days	Major transformation products, formed at pH 9, were CGA 355190 and NOA 404617 (for both the guanidine and thiazolyl radiolabels). In the study with the thiazolyl label, NOA 404617 further hydrolyzed to CGA 309335, which was still increasing at the end of the incubation period.	PMRA# 1178192 and 1178193	
	CGA 322704 (Clothianidin)	Hydrolytically stable at 20°C from pH 4 to pH 9.	Results are similar to existing information submitted to support the registration of clothianidin.	PMRA# 1529731	
Phototransformation on soil	Thiamethoxam	DT ₅₀ = 79 - 97 days (continuous irradiation)	There were no major transformation products other than CO ₂ . Several minor products were formed including CGA 322704, CGA 355190, CGA 353968 and CGA 282149 (all of which are also formed in aerobic soil). Other minor components were not identified. Transformation products were similar in the irradiated and dark samples (irradiation increased the rate of transformation, but did not produce any significant new transformation products).	PMRA# 1196656 and 1196657	
Phototransformation in air	Thiamethoxam	Not required – thiamethoxam is not vola	tile		
		Biotransforma	ation ¹		
Biotransformation in aerobic soil	Thiamethoxam	Sandy loam soil: DT ₅₀ = 286 - 346 days Representative half-life: 447 - 507 days Clay loam soil:	Moderately persistent to persistent. No major transformation products were formed in sandy loam soil. CGA 355190 was a major transformation product in clay loam soil, which further transformed to CGA 353968 with a half-life of 459 days (as reported in study; not recalculated by reviewer at this time). Several	PMRA# 1178196, 1178197 and 1178198	

Property	Test substance	Value	Comments	Study
		DT ₅₀ = 91 days Representative half-life: 122 days	minor transformation products were formed in both test soils, including CGA 322704, CGA 353968, CGA 282149 and CGA 309335. Under sterile conditions, the DT ₅₀ ranged from 286 - 686 days (as reported in study; not recalculated by reviewer at this time).	
	Thiamethoxam	DT ₅₀ at 20°C = 143 days (40% FC, high test dose), 74 days (60% FC, high test dose) and 34 days (60% FC, low test dose). DT ₅₀ at 10° C = 233 days (60% FC, high test dose) Representative half-lives: same	Tests systems were incubated at different combinations of temperature and humidity; drier soil conditions and a lower temperature slowed down the degradation. Also, two test concentrations were used; degradation was more rapid with a low concentration. CGA 322704 was a major transformation product. At 20°C, this compound degraded with a DT ₅₀ of 187 - 495 days depending on test conditions. Minor transformation products included CGA 355190, CGA 265307 and CGA 353968.	PMRA# 1529738
	Thiamethoxam	$DT_{50} = 3727 \text{ d}$ Representative half-life: $5.9 \times 10^8 \text{ d}$	Persistent. No major transformation products were formed in loamy sand soil (Gartenacker soil identified as Borstel soil in study report). CGA 322704 (clothianidin), CGA 355190 and CO ₂ were minor transformation products.	PMRA# 1529745
	Thiamethoxam	$DT_{50} = 78 - 158$ days Representative half-life: 110 - 258 days	Moderately persistent. CGA 322704 was a major transformation product. CGA 355190 was a minor transformation product. Tests were also performed with soils maintained in a greenhouse for months/years prior to the experiment. For these, the DT ₅₀ was longer (153 - 274 days).	PMRA# 1529741
	Thiamethoxam	DT ₅₀ = 60.1 days Representative half-life: same	Moderately persistent. CGA 322704 was a major transformation product. CGA 265307 was a minor transformation product.	PMRA# 2446844
	Thiamethoxam	DT ₅₀ = 78.7 days Representative half-life: same	Moderately persistent CGA 322704 was a major transformation product. Minor transformation products were CGA 355190, CGA 265307 and CGA 353968. This study also included tests with treated seeds. The radioactivity quickly moved from the treated seed to the surrounding soil. The DT ₅₀ in soil was 60.6 days; the more rapid dissipation attributed to the uptake by the growing	PMRA# 2446849

Property	Test substance	Value	Comments	Study
			plant.	
	CGA 322704 (Clothianidin)	Between 60 and 80% of the test substance degraded by the end of the study period of 120 days. CGA 265307 was identified as a minor transformation product.		PMRA# 1529745 and 1529746
	CGA 322704 (Clothianidin)	DT ₅₀ = 258 days Representative half-life: 317 days	No transformation products were identified.	PMRA# 1529747
	CGA 355190	DT ₅₀ = 9.16 - 89.7 days Representative half-life: 22.7 - 141 days	Non-persistent to moderately persistent, depending on soil type. CGA 353968 was identified as a major transformation product.	PMRA# 1529748
	NOA 407475	$DT_{50} = 376 - 443 \text{ days}$ Representative half-life: 419 - 461 days	Persistent. NOA 421275 was identified as a minor transformation product.	PMRA# 1529739 and 1529740
Biotransformation in	Thiamethoxam	See biotransformation in anaerobic water/sediment system (one study used soil rather than sediment).		
anaerobic soil	CGA 322704 (Clothianidin)	DT ₅₀ = 11.5 days Representative half-life: 22 days	Flooded soil, water was spiked. Radioactivity rapidly moved from the water to the soil layer. Major transformation products in anaerobic soil were NOA 421275 and one unidentified product. Minor transformation products were CGA 353968, CGA 265307 and several other unidentified components.	PMRA# 1529750
		Mobility ²		
Adsorption / desorption in soil	Thiamethoxam	Ads $K_d = 0.21 - 2.3 \text{ mL/g}$ Ads $K_{oc} = 33 - 177 \text{ mL/g}$	Moderate to very high mobility. Six soils. - GUS ³ of 4.3 to 6.3 depending on the soil type (leacher) - Most of the Cohen criteria ⁴ are met	PMRA# 1178199
		Ads $K_{oc} = 33 - 151 \text{ mL/g}$	Additional information. Values not verified / recalculated since acceptable data were already available. Values are within range of existing information.	PMRA# 1529758, 1196652, 1529769 and 1529770
		Time dependant sorption (incubation time ranging from 30 and 91 days): The K_{∞} increased with time with a factor of 2.4 - 7.6.	Additional information. Supports results of column leaching study with aged soil.	PMRA# 1196652, 1529769 and 1529771
	CGA 322704	Ads $K_d = 0.82 - 6.8 \text{ mL/g}$	Moderate to high mobility. Six soils.	PMRA#

Property	Test substance	Value	Comments	Study
	(Clothianidin)	Ads $K_{oc} = 58 - 273 \text{ mL/g}$		1196669
		Ads $K_{oc} = 62 - 77 \text{ mL/g}$	Additional information. Values not verified / recalculated since acceptable data were already available. Values are within range of existing information.	PMRA# 1529772 and 1529774
		Time dependant sorption (total incubation time of 91 days):	Additional information.	PMRA# 1529759
		The K_{∞} increased with time with a factor of 2.8.		
	CGA 355190	Ads $K_d = 0.45 - 3.3 \text{ mL/g}$ Ads $K_{oc} = 28 - 125 \text{ mL/g}$	High to very high mobility. Six soils.	PMRA# 1196670
	NOA 404617	Ads $K_d = 0.13 - 1.05 \text{ mL/g}$ Ads $K_{oc} = 8 - 43 \text{ mL/g}$	Very high mobility. Six soils.	PMRA# 119667
	NOA 407475	Ads $K_d = 2.5 - 44 \text{ mL/g}$ Ads $K_{oc} = 400 - 1453 \text{ mL/g}$	Low to moderate mobility. Six soils.	PMRA# 1196667
	CGA 353042	Ads $K_d = 1.8 - 24 \text{ mL/g}$ Ads $K_{oc} = 173 - 1413 \text{ mL/g}$	Low to moderate mobility. Six soils.	PMRA# 1196666
	NOA 459602	Adsorption increased with time to reach K_{oc} of 18 - 52 mL/g with incubation time of 71 days.	Additional information. Very high mobility. The registrant has postulated that these compounds are	PMRA# 1529765 and 1529766
	SYN 501406	Adsorption increased with time to reach K_{∞} of 24 - 34 mL/g with incubation time of 57 days.	transformation products of thiamethoxam in soil, as these were observed at low levels in lysimeter studies.	
Column leaching (unaged soil)	Thiamethoxam	Up to 59% of radioactivity recovered in leachate (amounts varied with soil type). Radioactivity was attributed to thiamethoxam.	Additional information. This compound was classified as moderately mobile in soil, based on the Relative Mobility Factor (RMF = leaching distance of test substance / leaching distance of reference substance). No transformation products were found in the soil or in the leachate.	PMRA# 1529777
Column leaching (aged soil)	Thiamethoxam	At the end of the aging period (30 days), the majority of the soil radioactivity was attributed to thiamethoxam; low amounts of CGA 282149, CGA 322704 and CGA 355190 were observed and less than	Thiamethoxam is less mobile in soil after ageing.	PMRA# 1178249

Property	Test substance	Value	Comments	Study
		2% of the applied radioactivity was recovered in volatile traps. The estimated DT_{50} for thiamethoxam was 124 - 320 days. The majority of the radioactivity remained in the soil after leaching and was mostly found in the 0-6 cm soil layer. Soil radioactivity was primarily thiamethoxam. Radioactivity in the leachate was 0 - 26 % of the applied amount. $K_d = 2.01 - 197.53 \text{ mL/g}$		
	Thiamethoxam	At the end of the ageing period (56 days), soil radioactivity was primarily attributed to thiamethoxam and CGA 322704 (55 - 63 % and 18 - 25 % of the applied amount, respectively); volatiles represented more than 30% of the applied radioactivity. The estimated DT ₅₀ for thiamethoxam was 65 -94 days. Most of the radioactivity remained in the soil after leaching. Thiamethoxam reached a depth of 30 cm (length of column), with highest amounts found at a depth of 12 - 24 cm. CGA 322704 was not found below 18 cm. Radioactivity in the leachate was 1.7 - 3.4 % of the applied amount.	Additional information.	PMRA# 1529778
Volatilization	Thiamethoxam WG 25	2.2% of thiamethoxam volatilized within 3h of application to soil surface. After 6 and 24 hours, volatilization was less than 1%.	Additional information. The volatilization was determined indirectly by measuring the residual radioactivity in the soil.	PMRA# 1529779
	Thiamethoxam	Estimated half-life from the atmospheric oxidation by hydroxyl radicals: 0.5 - 2.5 hours	Additional information. Estimated using the procedure described in Atkinson, R. 1998. Environ. Toxicol. Chem. 7: 435-442.	PMRA# 1529799
	CGA 322704 (Clothianidin)	Estimated half-life from the atmospheric oxidation by hydroxyl	Additional information. Estimated using the procedure described in Atkinson, R.	PMRA# 1529800

Property	Test substance	Value	Comments	Study
		radicals: 0.94 hours	1998. Environ. Toxicol. Chem. 7: 435-442, as developed in the Atmospheric Oxidation Program v1.8.	
		Field studi	es	
Field dissipation in site relevant to Canadian conditions: Alberta, Saskatchewan, Manitoba, Ontario	Helix Seed Treatment	Treated canola seeds at a rate of 500 g a.i./100kg seed: $DT_{50} = 161 \ days \ in \ Ontario. \ There \\ was no clear pattern of dissipation at the Saskatchewan site and a DT_{50} \\ was not \ determined. \ While \\ dissipation \ was \ observed \ in \ Alberta \\ and \ Manitoba, \ rate \ calculations \ were \\ not \ conclusive \ (high \ variability \ in \\ concentrations \ and \ the DT_{50} \ varied \\ markedly \ depending \ on \ the \ model).$	Moderately persistent to persistent in some sites. Major transformation products were CGA 355190 and CGA 322704 (clothianidin). These were detected at all sites in the 0-10 cm soil layer. Thiamethoxam generally remained in the top 10 cm of soil, with occasional detections in the 10-25 cm layer.	PMRA# 1178359
Field dissipation in site relevant to Canadian conditions: Manitoba	Actara 25 WG (25.1% a.i.)	Two broadcast applications at 26.3 g a.i./ha on bare ground: While some degradation is apparent in the first 100 days of the study, rate calculations were not conclusive because of an increase in measured concentrations the following spring.	Persistent. No major transformation products were formed. Minor transformation products were detected a low levels, generally below the limit of quantification. CGA 355190 was most often detected. Other minor transformation products include CGA 322704 (clothianidin), CGA 309335, CGA 353968, CGA 353042 and NOA 404617.	PMRA# 860996, 860997, 860998, 860999 and 1074854
	Actara 240 SC (240 g a.i./L)	One broadcast application at 118 g a.i./ha on bare ground: Rate calculations not conclusive (low initial concentrations and no dissipation pattern)	Transformation products were mostly observed in the 0-10 cm soil layer. No residues of thiamethoxam or its transformation products were found below 25 cm depth. Residues of thiamethoxam are expected to carry-over. Up to $\sim 85\%$ of the applied amount was remaining in the soil at the end of the growing season.	
Field dissipation in site relevant to Canadian conditions: Ontario	Actara 25 WG (25.1% a.i.)	Two broadcast applications at 26.3 g a.i./ha on bare ground: $DT_{50} = 49.8$ days	Slightly to moderately persistent. No major transformation products were formed. CGA 322704 (clothianidin) was observed in measurable	
	Actara 240 SC (240 g a.i./L)	One broadcast application at 118 g a.i./ha on bare ground: DT ₅₀ = 18.7 days	amounts at many sampling events. Other minor transformation products include CGA 353042, CGA 353968 and NOA 407475. Transformation products were mostly observed in the 0-10 cm soil layer. No residues of thiamethoxam or its transformation products were found below 25 cm depth. Residues of thiamethoxam are expected to carry-over. Up	

Property	Test substance	Value	Comments	Study
			to \sim 34% of the applied amount was remaining in the soil at the end of the growing season.	
Field dissipation in site relevant to Canadian conditions: PEI	Actara 25 WG (25.1% a.i.)	Two broadcast applications at 26.3 g a.i./ha on bare ground: $DT_{50} = 18.3 \text{ days}$	Slightly persistent. No major transformation products were formed. CGA 322704 (clothianidin) was observed in measurable	
	Actara 240 SC (240 g a.i./L)	One broadcast application at 118 g a.i./ha on bare ground: DT ₅₀ = 32.4 days	amounts at many sampling events. Other minor transformation products include CGA 353968 and NOA 407475.	
		D150 32.1 days	Transformation products were mostly observed in the 0-10 cm soil layer. No residues of thiamethoxam or its transformation products were found below 25 cm depth. Residues of thiamethoxam are expected to carry-over. Up to $\sim 22\%$ of the applied amount was remaining in the soil at the end of the growing season.	
Field dissipation in site relevant to Canadian conditions: Michigan	Actara 25 WG (25.5% a.i.)	Two broadcast applications at 112 g a.i./ha on bare ground: $DT_{50} = 26.9 \text{ d}$	Slightly persistent. CGA 322704 (clothianidin) was a major transformation product.	PMRA# 861000, 861001,
	Actara 4L (39.8% a.i.)	One in-furrow application at 157 g a.i./ha (941 g a.i./ha within the furrow): $DT_{50} = 26.8 \text{ d}$	Quantifiable levels of thiamethoxam were not observed beyond 30 cm (broadcast) and 90 cm (in-furrow). Quantifiable levels of CGA 322704 were not observed beyond 15 cm (broadcast) and 30 cm (in-furrow). These compounds were detected up to depths of 76 cm (broadcast) and 120 cm (in-furrow).	861002, 861003 and 861004
Field dissipation in site relevant to Canadian conditions: Switzerland ⁵	Thiamethoxam WG 25 formulation	One broadcast application at rate of 207 g a.i./ha on bare ground: DT ₅₀ = 52.9 d	Moderately persistent. Radiolabeled material was used. No major transformation products were formed. Minor transformation products were CGA 322704 (found in greater amounts, observed up to 20 cm depth), CGA 265307 and CGA 355190 (observed in 0-10 cm soil layer). Quantifiable amounts of thiamethoxam were found up to a depth of 50 cm. Amounts below the level of quantification	PMRA# 1529782
Field dissipation in site relevant to Canadian conditions: Switzerland ⁵	Thiamethoxam WG 25 formulation	One broadcast application at 200 g a.i./ha on bare ground: $DT_{50} = 6.84$	were detected up to a depth of 60 cm. Additional information. Short report of 5 pages summarizing analytical results. Quantifiable levels of thiamethoxam were not observed beyond a depth of 10 cm. Quantifiable levels of CGA 322704 were observed once in the 0-10 soil layer but not at	PMRA# 1529793

Property	Test substance	Value	Comments	Study
			any other sampling event.	
Field dissipation in site relevant to Canadian conditions: Switzerland ⁵	A9700B (350 g a.i./L)	Barley seed treatment at 70 g a.i./100 kg seed (equivalent to 150.5 g a.i./ha): $DT_{50} = 61.1 \text{ days}$	Moderately persistent. CGA 322704 was the only transformation product. Quantifiable levels of thiamethoxam and CGA 322704 were observed up to a depth of 30 cm.	PMRA# 2446857
Field dissipation in site relevant to Canadian conditions: Switzerland ⁵	A9584C 25 WG	One broadcast application at 200 g a.i./ha on bare ground: $DT_{50} = 12.1 \text{ days}$	Non-persistent. Transformation products were CGA 322704, CGA 355190 and NOA 407475. Quantifiable levels of thiamethoxam and CGA 322704 were observed up to a depth of 45 cm. NOA 407475 was not found beyond a depth of 30 cm.	PMRA# 2446861
Field dissipation in site relevant to Canadian conditions: France ⁵	Thiamethoxam WG 25 formulation	One broadcast application at 200 g a.i./ha on bare ground: $DT_{50} = 14.5 - 205 \text{ days}$	Additional information. Short reports of 13 - 15 pages summarizing analytical results. Quantifiable levels of thiamethoxam and CGA 322704 were found up to 30 cm in some study sites. Crops had recently been sown at the time the pesticide was applied (corn, soybean or grass depending on the plot). Crop uptake was not assessed.	PMRA# 1529794, 1529795, 1529796, 1529797, 1529783 and 1529784
Field dissipation in site relevant to Canadian conditions: France ⁵	A9700B (350 g a.i./L)	Barley seed treatment at 70 g a.i./100 kg seed (equivalent to 148.4 g a.i./ha): $DT_{50} = 22.4$ days	Slightly persistent. Transformation products were CGA 322704 and CGA 355190. Quantifiable levels of thiamethoxam and CGA 322704 were observed up to a depth of 30 cm. CGA 355190 was not found beyond 10 cm.	PMRA# 2446859
Field dissipation in site relevant to Canadian conditions: Spain ⁵	Actara 25 WG	One broadcast application at 200 g a.i./ha on bare ground: $DT_{50} = 36.1 \text{ days}$	Slightly persistent. On bare ground, quantifiable levels of thiamethoxam were not observed beyond a depth of 20 cm. In the cropped plot, quantifiable levels of thiamethoxam were observed up to a depth of 30 cm (both in treated row and between rows). Quantifiable levels of CGA 322704 were not observed at any sampling event in either plot.	PMRA# 1529789
Field dissipation in other site: California	Platinum 75SG (75% a.i.) Actara 25 WG (25% a.i.)	In-furrow application at 328 g a.i./ha followed 31 and 38 days later by two broadcast sprays at 106 g a.i./ha (bare soil): DT ₅₀ = 16.3 days (after last spray	Non-persistent (cropped) to slightly persistent (bare soil). In the bare soil plot, CGA 322704 and CGA 355190 were major transformation products. Only CGA 322704 was a major transformation product in the cropped plot. NOA 404617, CGA 353042 and NOA 407475 were minor	PMRA# 2446854

Property	Test substance	Value	Comments	Study
		application) In-furrow application at 328 g a.i./ha followed 31 and 38 days later by two broadcast sprays at 106 g a.i./ha (cropped with spinach): DT ₅₀ = 5.51 days (after last spray application)	transformation products in the bare soil plot. These, in addition to CGA 355190 were minor transformation products in the cropped plot. Thiamethoxam was detected up to 36 inches in both the bare soil and cropped plots. CGA 322704 and CGA 355190 were also detected in deeper soil layers.	
Multi-year accumulation study: Switzerland	A9584A or A9584C (25% a.i.)	years as a foliar spray (four applications or peas. Soil was analyzed for thiametho 355190 (last three years of the study) and Concentrations of thiamethoxam in the application of the year, and then dissip residue concentration observed in the observed in the last year of the study, I thiamethoxam accumulates in soil with Thiamethoxam concentrations further concentrations of 0.017 and 0.005 mg/ respectively. No quantifiable residues Concentrations of clothianidin fluctuat of thiamethoxam each year, but the discrop cycle, contrary to what was gener concentrations of clothianidin were 0.0 cm and 20-30 cm soil depths, respectively the level of quantification. Concentrations of CGA355190 were gnowledged.	decreased with deeper soil layers, with maximum /kg dry soil in the 10-20 and 20-30 cm soil layers,	PMRA# 2446853
Field lysimeter	Thiamethoxam WP 25 formulation	treatment was repeated for a second year at a depth of 130 cm. Crops were harvest The amount of total radioactive residues 33%, 2.6-3.0% and 16-21% of the applied radioactivity was attributed to lot The majority of the total radioactive residues.	in soil, leachate and treated crop represented approximately a radioactivity, respectively. Approximately 63% of the	PMRA# 1529775

Property	Test substance	Value Comments	Study
		observed in all layers from 0 to 40 cm and represented 5.5-6.7% of the applied amount. Thiamethoxam, CGA 322704 and unidentified residues were found in the leachate.	
	Thiamethoxam WS 70 seed treatment formulation	Crops sown in the first year: spring barley seeds treated at a rate of 35 g a.i./100 kg seed (equivalent to 52.5 g a.i./ha). After the harvest of barley, planted winter wheat seeds treated at a rate of 63 g a.i./ha. Second year: Planted winter rape seeds treated at a rate of 420 g a.i./100 kg seed, equivalent to 21 g a.i./ha, in one of the two lysimeters plots. Crops were harvested at maturity for analysis. Lysimeters were placed at a depth of 120 cm.	PMRA# 1529776
		The amount of total radioactive residues in soil, leachate and treated crop represented 50-57%, 3.7-4.2% and 1.4-1.6% of the applied radioactivity, respectively. Approximately 38-44% of the applied radioactivity was attributed to losses due to mineralization.	
		The majority of the total radioactive residues in soil were in the 0-40 cm layers. Overall, thiamethoxam and CGA 322704 (clothianidin) in soil represented 3.4-3.8% and 20-25% of the applied radioactivity, respectively.	
		Thiamethoxam, CGA 322704, NOA 459602 and SYN 501406 were found in the leachate.	
Small Scale Prospective Groundwater Monitoring - Michigan ⁶	Platinum 2SC	One in-furrow spray application of the test substance at 193 g a.i./ha when planting cucumber seeds, followed by one ground spray application (without incorporation) of a potassium bromide tracer at 101 kg/ha. Monitoring was carried out for a period of 59 months after treatment (MAT). Surface soil (0-6 inch), soil pore water (suction lysimeters at 3, 6, 9 and 15 feet below ground surface) and groundwater samples (wells at 20-30 and 30-35 feet below ground surface) were collected.	PMRA# 1108402 (progress report) and 1751758
		Rapid movement of the bromide tracer was observed (aquifer recharge at approx. 6 MAT), confirming permeability of the soil. Also, tracer concentrations peaked and then declined back to background levels in lysimeters and wells (i.e. showing movement through the vardose zone and into the groundwater where it continued to decline).	(final report)
		In lysimeters: Thiamethoxam peaked at 14 MAT (max: 3.5 ppb, observed at 9 feet) and declined thereafter. CGA 322704 peaked at 38 MAT (max: 0.57 ppb, observed at 9 feet) and declined thereafter. CGA 355190 was found sporadically (max: 0.078 ppb, observed at 9 feet).	
		In groundwater - shallow wells: Thiamethoxam was first observed at 27 MAT, peaked at around 43 MAT (0.16 ppb) and declined thereafter. NOA 459602 was first observed at 12 MAT, peaked at 13-27 MAT (max: 0.089 ppb) and declined thereafter. SYN 501406 was first observed at 12 MAT, peaked at 21-38 MAT (max: 0.13 ppb) and declined thereafter.	
		In groundwater - deep wells: Thiamethoxam residues were not found. There were only two detections of NOA 459602 at 28-29 MAT (max: 0.063 ppb). SYN 501406 was first detected at 28 MAT and peaked at around 33MAT (max: 0.096 ppb) and declined thereafter.	
		CGA 322704, CGA 355190, CGA 353042, NOA 404617, and NOA 407475 were not found in groundwater.	

- 1 Classification of the relative persistence of pesticide in soils is based on Goring et al. (1975).
- 2 Classification of soil mobility potential is based on McCall et al. (1981)
- 3 GUS = Groundwater Ubiquity Score, based on Gustafson (1989)
- 4 Described in Cohen et al. (1984)
- The relevance of European test sites to Canadian ecoregions was evaluated using ENASGIPSV230_Arc10.2. All European sites from studies shown in this table were found to be relevant to Canada. Other European studies were in an ecoregion not found in North America (Baltic mixed forest) and are not shown in this table: Riepsdorf, Germany [PMRA# 1529785]; Middelfart, Denmark [PMRA# 1529787] and Bjärred, Sweden [PMRA# 1529788].
- Another small scale prospective groundwater monitoring study was performed in Georgia [PMRA# 1751760]. This study was not reviewed, as site is not relevant to Canada.

Table A.3-5 Fate and behaviour in the aquatic environment

Property	Test substance	Value	Comments	Study			
	Abiotic transformation						
Hydrolysis	Thiamethoxam	At 25°C: t½ pH 5: stable t½ pH 7: 559 - 939 days t½ pH 9: 4.1 - 8.0 days	Major transformation products, formed at pH 9, were CGA 355190 and NOA 404617 (for both the guanidine and thiazolyl radiolabels). In the study with the thiazolyl label, NOA 404617 further hydrolyzed to CGA 309335, which was still increasing at the end of the incubation period.	PMRA# 1178192 and 1178193			
Phototransformation in water	Thiamethoxam	$DT_{50} = 2.3 - 3.1$ days (continuous irradiation)	Major transformation products were CGA 353042 (guanidine label) and carbonyl sulfide (volatile product from thiazolyl label). Identified minor transformation products were CGA 355190, CGA 322704, NOA 407475, CGA 353968 and methyl urea. Other minor products were not identified.	PMRA# 1196653 and 1196654			
	Thiamethoxam	DT ₅₀ from 0.76 - 0.84 days in summer to 3.3 -7.8 days in winter in natural sunlight at 40°N - 50°N (annual mean of 1.2 - 1.6 days)	Additional information. Not fully reviewed. No information on transformation products.	PMRA# 152973			
	CGA 322704 (clothianidin)	DT ₅₀ from 7.2 hours in summer to 8.5 days in winter in natural sunlight at 52°N	No major transformation products were formed. Identified minor transformation products were CGA 353968 and NOA 404617. The estimated environmental half-life was not verified by the reviewer since existing data for the phototransformation of clothianidin were consistent with results from this study.	PMRA# 1529737			

Property	Test substance	Value	Comments	Study	
B iotransformation ¹					
Biotransformation in aerobic water	Thiamethoxam	Pond water at 25°C: DT ₅₀ = 9.7 - 24 days Representative half-life: 9.7 - 24 days	Non-persistent to slightly persistent in water. Major transformation products were CGA 355190 and NOA 404617. Minor transformation products were CGA 353968 and one unidentified product. The DT ₅₀ was 12 - 16 days under sterile conditions and the same transformation products as in viable samples were formed. This suggests that transformation was from hydrolysis, which is possible given slightly basic conditions during the study (pH 8.22 to pH 8.67). Major products formed in viable samples were also observed in hydrolysis study.	PMRA# 1196651 and 1196660	
Biotransformation in aerobic water-sediment system	Thiamethoxam	Pond water - loam sediment system at 25°C: DT ₅₀ = 7.2 - 15.0 days (water), 8.3 - 16.3 (whole system) Representative half-life: 9.1 - 15.0 days (water), 8.3 -16.3 (whole system)	Non-persistent to slightly persistent in whole system. NOA 407475, a major transformation product for both the guanidine and thiazolyl labels, was detected primarily in the sediment. CGA 355190 was a major transformation product with the thiazolyl label, but a minor transformation product with the guanidine label. NOA 404617 was a minor transformation product for both labels. CGA 355190 and NOA 404617 are thought to have been formed from the hydrolysis of the parent (pH 8.22 - pH 8.67 in water). Under sterile conditions, the DT ₅₀ was 28 - 35 days (water phase) and 29 - 38 days (whole system). CGA 355190 and NOA 404617 were major transformation products (hydrolysis). Only low levels of NOA 407475 were formed.	PMRA# 1196651 and 1196660	

Property	Test substance	Value	Comments	Study
	Thiamethoxam	River water - sediment system at 20°C: $DT_{50} = 11.9 - 12$ days (water), 35 - 42.8 days (whole system) Representative half-life: 35.9 - 45.5 days (water), 42.8 - 59.4 (whole system) Pond water - sediment system at 20°C: $DT_{50} = 8.3 - 10.6$ days (water), 26.2 - 31.7 days (whole system) Representative half-life: 23.7 - 23.8 days (water), 31.7 - 40.4 (whole system)	Slightly to moderately persistent in whole system. NOA 407475, a major transformation product for both labels, was formed in the sediment. CGA 355190 was a minor transformation product for both labels observed in both the water and sediment phases. A mean sediment/water distribution coefficient was estimated as $K_d = 2.1 - 2.7 \text{ mL/g}$.	PMRA# 1529752 and 1529753
	Thiamethoxam	River - sandy loam sediment system at 20°C: DT ₅₀ = 16.8 - 20.5 days (water), 51.5 - 60.8 days (whole system) Representative half-life: 35.6 - 42.1 days (water), 143 - 194 (whole system)	Slightly to moderately persistent in whole system. In water, NOA 404617 and CGA 355190 were minor transformation products. In the sediment, NOA 407475 was a major transformation product; NOA 404617 and CGA 355190 were identified as minor transformation products.	PMRA# 2529331
	CGA 322704 (clothianidin)	River water - sediment system at 20°C : DT ₅₀ = 23.1 days (water), 45.2 days (whole system) Representative half-life: 34.4 days (water), 45.2 (whole system) Pond water - sediment system at 20°C : DT ₅₀ = 10.9 days (water), 25.1 days (whole system) Representative half-life: 16.5 days (water), 25.1 (whole system)	Slightly persistent in whole system. NOA 407475 was a major transformation product in the sediment. No major transformation products were formed in the water phase. Minor transformation products were not identified. Other information provided in study but not verified by reviewer: CGA 322407 DT ₅₀ in the sediment = 67.9 d (river) and 63.1 d (pond) NOA 421275 DT ₅₀ in the sediment = 248 d (river) and 102 d (pond)	PMRA# 1529754

Property	Test substance	Value	Comments	Study
Biotransformation in anaerobic water- sediment system	Thiamethoxam	Sandy loam soil flooded with water at 25° C: DT ₅₀ = 15.9 - 18 days (water), 29 - 70.5 days (soil), 27.2 - 28.1 days (whole system) Representative half-life: 18 - 19.5 days (water), 29 - 70.5 days (soil), 27.2 - 28.1 days (whole system)	Slightly persistent in whole system. NOA 407475 was the only major transformation product in the soil layer. No major transformation products were formed in the water phase. Minor transformation products in the soil were CGA 322704, CGA 355190, NOA 404617, and CGA 353968. Minor transformation products in the water phase were NOA 407475, CGA 322704, CGA 355190 and NOA 404617.	PMRA# 1196658 and 1196659
	Thiamethoxam	River - silt loam sediment system at 20°C: DT ₅₀ = 27.5 - 28.1 days (water), 81.8 - 85.1 days (whole system) Representative half-life: 51.1 - 57.1 days (water), 81.8 - 85.1 (whole system)	Moderately persistent in whole system. In water, CGA 355190 was a major transformation product; NOA 407475 and NOA 404617 were minor transformation products. In the sediment, NOA 407475 and CGA 355190 were major transformation products; NOA 404617 was identified as minor transformation product.	PMRA# 2529332
Biotransformation in anaerobic water at low temperature	Thiamethoxam	Pond water at 5°C: DT ₅₀ = 12.6 days	Not recalculated (degradation at low temperature not currently a requirement and is not used for modelling). Slightly persistent. Major transformation products were CGA 355190 and NOA 404617. NOA 407475 was identified as a minor transformation product. The DT ₅₀ was also 12.6 days under sterile conditions and the same transformation products as in viable samples were formed. Hydrolysis is the likely route of dissipation in both sterile and viable samples given basic conditions (pH 9.09 to pH 9.95). Also, hydrolysis at pH 9 is rapid, which may explain why degradation was not slower at lower temperatures.	PMRA# 1196650

Property	Test substance	Value	Comments	Study
Biotransformation in anaerobic water-sediment system at low temperature	Thiamethoxam	Pond water - loam sediment system at 5°C: DT ₅₀ = 39.8 days (water), 53.3 days (sediment), 43.9 days (whole system)	Not recalculated (degradation at low temperature not currently a requirement and is not used for modelling). Slightly persistent in the whole system. NOA 407475 was the only major transformation product formed in the sediment. No major transformation products were formed in the water phase. Minor transformation products in sediment were CGA 355190, NOA 404617, and CGA 282149. Minor transformation products in the water phase were NOA 407475, CGA 355190, NOA 404617 and CGA 282149. Under sterile conditions, the DT ₅₀ was 126 and 204 days for the water phase and the whole system, respectively. Major transformation products were NOA 404617 and CGA 355190 (both found mostly in water).	PMRA# 1196650

¹ Classification of the relative persistence of pesticides in water is based on McEwen and Stephenson, 1979.

Table A.3-6 Information on the fate of thiamethoxam from the scientific literature

Type of information	Cited information	Comments	Reference		
	Physical and chemical properties				
Water solubility	4100 mg/L	Original source: pesticide properties database	As cited in Bonmatin et		
Log K _{oc}	-0.13	(http://sitem.herts.ac.uk/aeru/ppdb/en/index.htm)	al. (2015)		
$pK_{\rm a}$	No dissociation				
		Abiotic transformation			
Hydrolysis	Stable at pH1 to pH7	Original source: de Urzedo et al. 2007. Photolytic degradation of the insecticide thiamethoxam in aqueous medium monitored by direct infusion electrospray ionization mass spectrometry. Int J Mass Spectron 42: 1319-1325 [picked up by our literature search]	As cited in Simon- Delso et al. (2015)		
	Quickly hydrolyzed at pH9 and 20°C	Original source: European Commission 2006. Review report for the active substance thiamethoxam. Accessible at: http://ec.europa.eu/food/plant/protection/evaluation/newactive/thiamethoxam_en.pdf			
Aqueous photolysis	$DT_{50} = 2.7 \text{ days}$	Original source: pesticide properties database (http://sitem.herts.ac.uk/aeru/ppdb/en/index.htm)	As cited in Bonmatin et al. (2015)		

Type of information	Cited information	Comments	Reference
	Susceptible to direct photolysis	Original source: Peña et al. 2011. Persistence of two neonicotinoid insecticides in wastewater, and in aqueous solutions of surfactants and dissolved organic matter. Chemosphere, 84(4), 464-470 [picked up by our literature search] A cursory examination of the above article provided more context: Aqueous solutions (MilliQ water) containing thiamethoxam were placed outdoors and exposed to sunlight for 10 h a day. The UV spectrum of thiamethoxam showed a high intensity absorption band at 250–255 nm, extending >290 nm, which means that the insecticide absorbs in the tropospheric range of sunlight, being thus susceptible to direct photolysis. A DT ₅₀ of 18.7 hours is reported by the authors. There was no degradation in dark controls.	
	Almost completely degraded (ca. 96%) under UV radiation in about 10 min	Original source: de Urzedo et al. 2007. [see above]	As cited in Simon- Delso et al. (2015)
		Biotransformation	
Biotransformation in soil	$DT_{50} = 7 - 335 \text{ days}$	Original source: Goulson D. 2013. An overview of the environmental risks posed by neonicotinoid insecticides. J Appl Ecol 50(4):977-987 [picked up in our literature search]. Reported by Goulson (misreported in Bonmatin et al.): 7-353 days. Most values reported by Goulson were drawn from the Australian (APVMA) review of thiamethoxam and Cruiser 350 FS. The 7-day value is likely that calculated by the registrant based on data PMRA# 1529793.	As cited in Bonmatin et al. (2015)
	DT ₅₀ = 46 - 75 days (submerged soil), 91 - 94 days (field moisture capacity) and 201 - 301 days (dry soil)	Original source: Gupta et al. 2008. Soil dissipation and leaching behaviour of a neonicotinoid insecticide thiamethoxam. Bull Environ Contam Toxicol 80:431-437 [picked up in our literature search] Notes from cursory examination of article: Analytical grade thiamethoxam was applied to soil with varying moisture levels at concentrations of 0.01 and 0.1 µg/L. Dissipation is reported to be biphasic; the SFO half-life was 16.1 - 115.5 days and 60.2 - 376.3 days for the first and second phase, respectively, when considering all test concentrations and moisture regimes. Rates were faster at the low test concentration. Ranges are within currently available data for thiamethoxam.	
Biotransformation in water-sediment	$DT_{50} = 40 \text{ days}$	Original source: pesticide properties database (http://sitem.herts.ac.uk/aeru/ppdb/en/index.htm)	As cited in Bonmatin et al. (2015)

Type of information	Cited information	Comments	Reference				
		Mobility					
Groundwater ubiquity score	3.82	Original source: pesticide properties database (http://sitem.herts.ac.uk/aeru/ppdb/en/index.htm)	As cited in Bonmatin et al. (2015)				
Soil column leaching	65 cm of rainfall resulted in the leaching of 66-79% of the applied thiamethoxam and no residues were detected in the soil	Original source: Gupta et al. 2008. Soil dissipation and leaching behaviour of a neonicotinoid insecticide thiamethoxam. Bull Environ Contam Toxicol 80:431-437 [picked up in our literature search] Notes from cursory examination of article: Analytical grade thiamethoxam and two thiamethoxam formulations (Actara and Cruiser) were applied to column soil from India, with little difference in leaching behaviour, although slightly higher amount was recovered in leachate of analytical grade than formulation treatment.					
Sorption	Detection of contamination of groundwater is only a matter of time	Original source: Kurwadkar et al. 2013. Time dependent sorption behavior of dinotefuran, imidacloprid and thiamethoxam. Journal of Environmental Science & Health - Part B, 48: 237-242 [picked up in our literature search] A Notes from cursory examination of article: The time-dependant sorption of thiamethoxam (and other neonicotinoids) was studied ion the lab using soil from a vineyard, sampling interval varied of 0, 2, 4, 8, 12, 24, 60 and 96 hours. Sorption increased with time, but remained low.					
		Field studies					
Field lysimeter	(different location each 75% thiamethoxam, at thiamethoxam, at a rate impregnated polyacryla Platinum 75SG per 75 at thiamethoxam, at a 7-daground surface.	treatments were made on potato in Wisconsin. Trials were carried out for two years year). Treatments were: (1) one in-furrow application of Platinum 75SC, containing a rate of 140 g a.i./ha; (2) seed treatment with Cruiser 5FS, containing 47.6% of 112 g a.i./ha at planting density of 1793 kg seed/ha; (3) thiamethoxammide horticultural granules at 16 kg (of granule?)/ha (with a ratio of 0.834g of granule?); and (4) Two foliar applications of Actara 25WG, containing 25% ay interval and a rate of 105 g a.i./ha/season. Lysimeters were placed at 75 cm below	Huseth and Groves (2014)				
	impregnated polyacryla approximately 17.5 µg/ approximately 12 µg/L for seed treatment and Low levels of thiameth	Residues in leachates were higher at the end of the growing season. The highest residues resulted from impregnated polyacrylate granules. Based on graphical data, thiamethoxam residues in leachate reached up to approximately 17.5 µg/L for impregnated polyacrylate granules (observed 154 days after planting), approximately 12 µg/L for in-furrow application (observed 123 days after planting), approximately 11 µg/L for seed treatment and foliar applications (observed 123 days after planting). Low levels of thiamethoxam residues were also found in leachate from control plots; these were attributed to the contamination of wells from which irrigation water was drawn.					

Table A.3-7 Thiamethoxam and its transformation products formed in the environment

Description	Structure	Matrix: Process (details)
Parent molecule:		
Thiamethoxam	0X	NA
Transformation products (ordered alphanumerically by code	e name):	
CGA 265307 IUPAC Name: N-(2-chloro-thiazol-5-ylmethyl)-N'-nitro-guanidine CAS Name: N-[(2-chloro-5-thiazolyl)methyl]-N'- nitroguanidine CAS Number: 135018-15-4 Molecular formula: C ₅ H ₅ ClN ₅ O ₂ S Molar mass: 235.65	H ³ h H	Soil: Aerobic (minor) Field dissipation (minor) Aerobic and anaerobic (minor, study with CGA 322704) Water: NA Plant: Metabolism (major)
CGA 282149 IUPAC Name: <i>N</i> -nitro-(3-methyl-[1,3,5]-oxadiazinan-4-ylidene)-amine CAS Name: 3,6-dihydro-3-methyl- <i>N</i> -nitro-2H-1,3,5-oxadiazin-4-amine CAS Number: 153719-38-1 Molecular formula: C ₄ H ₈ N ₄ 0 ₃ Molar mass: 160,03	O,N N N N	Soil: Phototransformation (minor) Aerobic (minor) Water: Anaerobic water-sediment at low temperature (minor in sediment and water) Plant: NA
CGA 309335 IUPAC Name: 2-chlorothiazoly-5-lmethyl-amine CAS Name: 2-chloro-5 thiazolemethanamine CAS Number: 120740-08-1 Molecular formula: C ₄ H ₅ ClNS Molar mass: 148.61	H ₂ N S CI	Soil: Hydrolysis (major at pH 9) Aerobic (minor) Field dissipation (minor) Water: Hydrolysis (major at pH 9) Plant: NA
CGA 322704 (Clothianidin) IUPAC Name: 1-(2-chloro-thiazol-5-ylmethyl)-3-methyl- N-nitroguanidine CAS NAME: (E)-N-[(2-chloro-5-thiazolyl)methyl]-N'- methyl-N''-nitroguanidine CAS Number: 205510-53-8 Molecular formula: C6H8ClN5O2S Molar mass: 249.68	HN H CS N CI	Soil: Phototransformation (minor)

Description	Structure	Matrix	: Process (details)
CGA 353042	NH	Soil:	Field dissipation (minor)
IUPAC Name: 3-methyl-1,3,5]oxadiazinan-4-	L. A. J.	Water:	Phototransformation (major)
ylideneamine	H,C-N / NH	Plant:	Metabolism (major)
CAS Name: 3,6-dihydro-3-methyl-2H-1,3,5-oxadiazin-4-			
amine	0		
CAS Number: not issued			
Molecular formula: C ₄ H ₈ N ₃ O			
Molar mass: 115.14			
CGA 353968	Q Q	Soil:	Phototransformation (minor)
IUPAC Name: 1-(2-chloro-thiazol-5-ylmethyl)-3-methyl-	l		Aerobic (minor)
urea	HN		Anaerobic water-soil (minor in soil)
CAS Name: N-[(2-chloro-5-thiazolyl)methyl]-N'-methyl-			Field dissipation (minor)
urea			Aerobic (major, study with CGA 355190)
CAS Number: not issued			Anaerobic (minor, study with CGA 322704)
Molecular formula: C ₆ H ₈ ClN ₃ OS		Water:	` ′
Molar mass: 205.67			Aerobic water (minor)
			Phototransformation (minor, study with CGA
			322704)
		Plant:	Metabolism (minor)
CGA 355190	l Y	Soil:	Hydrolysis (major at pH 9)
IUPAC Name: 3-(2-chloro-thiazol-5-ylmethyl)-5-	\.,\\\.,\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\		Phototransformation (minor)
methyl[1,3,5]oxadiazinan-4-one	7 7 N //		Aerobic (major)
CAS Name: 3-[(2-chloro-5-thiazolyl)methyl]tetrahydro-5-			Anaerobic water-soil (minor in soil and water)
methyl-4H-1,3,5-oxadiazin-4-one	\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \		Field dissipation (major)
CAS Number: not issued			Leaching (PGW)
Molecular formula: C ₈ H ₁₀ ClN ₃ O ₂ S		Water:	Hydrolysis (major at pH 9)
Molar mass: 247.17			Phototransformation (minor)
			Aerobic water (major)
			Aerobic water-sediment (major)
			Anaerobic water-sediment (major in sediment and
		Di	water)
		Plant:	Metabolism (minor)

Description	Structure	Matrix: Process (details)
NOA 404617 IUPAC Name: 1-(2-chloro-thiazol-5-ylmethyl)-3- nitrourea CAS Name: N-[(2-chloro-5-thiazolyl)methyl]-N'-nitro- urea CAS Number: not issued Molecular formula: C ₅ H ₅ ClN ₄ O ₃ S Molar mass: 236.63		Soil: Hydrolysis (major at pH 9) Anaerobic water-soil (minor in soil and water) Field dissipation (minor) Water: Hydrolysis (major at pH 9) Aerobic water (major) Aerobic water-sediment (minor) Anaerobic water-sediment (minor in sediment and water)
		Phototransformation (minor, study with CGA 322704) Plant: NA
NOA 405217 IUPAC Name: N-nitro-N'-methyl-guanidine CAS Name: N-nitro-N'-methyl-guanidine CAS Number: not issued Molecular formula: C ₂ H ₆ N ₄ O ₂ Molar mass: 118.10	M NH2	Soil: NA Water: NA Plant: Metabolism (minor)
NOA 407475 IUPAC Name: 3-(2-chloro-thiazol-5-ylmethyl)-5- methyl[1,3,5]oxadiazinan-4-ylideneamine CAS Name: 3-[(2-chloro-5-thiazolyl)methyl] tetrahydro- 5-methyl- <i>N</i> -nitro-4H-1,3,5-oxadiazin-4-imine CAS Number: not issued Molecular formula: C ₈ H ₁₁ ClN ₄ OS Molar mass: 246.72	H ₃ C _N S	Soil: Anaerobic water-soil (major in soil, minor in water) Field dissipation (minor) Water: Phototransformation (minor) Aerobic water-sediment (major) Anaerobic water-sediment (major in sediment, minor in water) Aerobic (major in sediment, study with CGA 322704) Plant: Metabolism (major)
NOA 421275 IUPAC Name: N-(2-chloro-thiazol-5-ylmethyl)-N'- methyl-guanidine CAS Name: N-[(2-chloro-thiazol-5-ylmethyl)]-N'-methyl- guanidine CAS Number: not issued Molecular formula: C ₆ H ₉ ClN ₄ S Molar mass: 204.68	a Name	Soil: Aerobic (minor, study with NOA 407475 Anaerobic (major, study with CGA 322704) Water: NA Plant: Metabolism (major)

Description	Structure	Matrix	: Process (details)
NOA 459602	NO ₂	Soil:	Leaching (Field lysimeter, PGW)
IUPAC Name: 5-(5-methyl-4-nitroimino-	N e	Water:	NA
[1,3,5]oxadiazinan-3-ylmethyl)thiazole-2-	\ [*] \/~\/ [*] \/\$0,#	Plant:	NA
sulfonate			
CAS Name: 5-[(5-methyl-4-nitroimino-			
[1,3,5]oxadiazinan)-3-ylmethyl)]thiazole-2-sulfonate	₹.8		
CAS Number: not issued			
Molecular formula: $C_8H_{11}N_5O_6S_2$			
Molar mass: 337.32			
NOA 501406 / SYN 501406 a	/***~ \$a	Soil:	Leaching (Field lysimeter, PGW)
IUPAC Name:	\$ "_N#	Water:	NA
5-(N'-Methyl-N''-nitro-guanidinomethyl)-thiazole-2-		Plant:	NA
sulfonate	SO ₃ H N T		
CAS Name: 5-(N'-Methyl-N''-nitroguanidinomethyl)-	NO.		
thiazole-2-sulfonate	*		
CAS Number: not issued			
Molecular formula: $C_6H_9N_5O_5S_2$			
Molar mass: 295.29			
Carbonyl Sulfide	O=C=S	Soil:	NA
CAS Number: 463-58-1		Water:	Phototransformation (major)
		Plant:	NA
Methylurea	R	Soil:	NA
Molecular formula: C ₂ H ₆ N ₂ O		Water:	Phototransformation (minor)
Molar mass: 74.08		Plant:	Metabolism (minor)
	HW N		

Italic font was used when transformation process was observed in a study carried out with a thiamethoxam transformation product rather than thiamethoxam itself.

The following transformation products are thought to be common to both thiamethoxam and clothianidin: CGA 265307 = TZNG, CGA 353968 = TZMU, NOA 405217 = MNG and NOA 421275 = TMG.

^a NOA 501406 and SYN 501406 are believed to be the same compound; both names are used in documentation provided by the registrant (e.g., Tier II summaries, PMRA# 1529715).

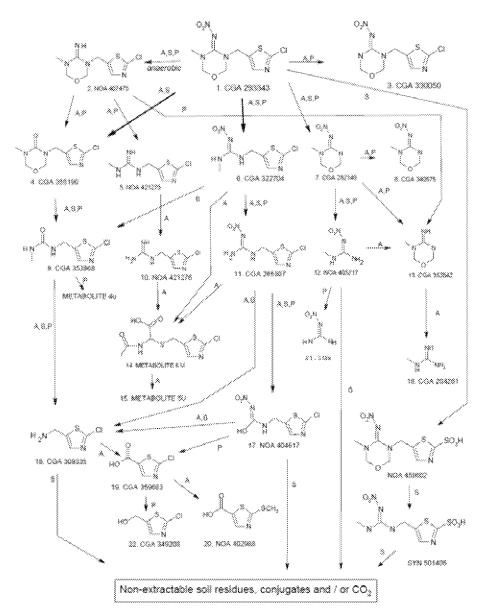


Figure A.3-1 Proposed transformation pathway for thiamethoxam in soil (S), plants (P) and animals (A)

Source: Tier II Summary prepared by registrant (PMRA# 1529715)

P = photolysis

B = biological degradation

Figure A.3-2 Proposed transformation pathway for thiamethoxam in the aquatic environment

Source: Tier II Summary prepared by registrant (PMRA# 1529715)

Table A.3-8 Effects of thiamethoxam and formulated products containing thiamethoxam alone on aquatic invertebrates

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
		<u> </u>		Acute		ı	1
Freshwater inver	tebrates						
Crustaceans - Cla	adocera						
Daphnia magna	Acute 48-h	Thiamethoxam (98.6%)	$48-h EC_{50} >$ $106 000$ $(15\% mortality/immobilization)$	Practically non-toxic	No ¹		1196664
	Acute 48-h	Formulation (WG25%)	48-h EC ₅₀ > 25 000	Practically non-toxic	No ¹	Practically non-toxic based on amount of EP (48-h EC ₅₀ > 100 000 µg EP/L)	EC 2006
	Acute 48-h	Formulation (WG70%)	48-h EC ₅₀ = 27 300	Slightly toxic	Yes	48 -h EC ₅₀ = 39 000 μ g EP/L	EC 2006
	Acute 48-h	Formulation SC 240 (A9795B) (21.5% thiamethoxam)	48-h EC ₅₀ > 106 000 (0% mortality/ immobilization)	Practically non-toxic	No ¹		2712668
	Acute 48-h	Formulation Actara 75WG (A-9549C) (74.8% thiamethoxam)	$48-h EC_{50} >$ $100~000$ (0% mortality/ immobilization)	Practically non-toxic	No ¹		2712669
	Acute 48-h	Formulation A 9584 C (25.4% thiamethoxam)	48-h EC ₅₀ > 25 400 (5% mortality/immobilization)	Practically non-toxic	No ¹	Practically non-toxic based on amount of EP (EC ₅₀ >100 000 μg/L EP)	2712675
	Acute 48-h	Formulation FS 600 (A 9765 C; 605.4 g a.i./L thiamethoxam)	48-h EC ₅₀ > $46 100$ (30% mortality/immobilization)	Practically non-toxic	No ¹	Practically non-toxic based on amount of EP (EC ₅₀ >100 000 μg/L EP)	2712676
	Acute 48-h	Formulation A 9700 B (35.8% thiamethoxam)	$48-h EC_{50} > 35$ 000 (0% mortality/immobilization)	Practically non-toxic	No ¹	Practically non-toxic based on amount of EP (EC ₅₀ >100 000 μg/L EP)	2712678
	Acute 48-h	Formulation	$48-h EC_{50} = 27.3$	Highly toxic	Yes		2712665 (Li et al.

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
		WG25%	(20.4–36.1) (mortality/ immobilization)				2013)
	Acute 48-h	Thiamethoxam (≥ 98.6% purity)	48-h LC ₅₀ > 80 000 (0% mortality)	Not toxic up to highest concentration tested.	No ¹	EC ₅₀ Not available (immobilization not recorded)	2842540 (Raby et al. 2018)
Daphnia pulex	Acute 24-h	Thiamethoxam (98.6%)	24-h EC ₅₀ > $100~000$ (20% mortality/immobilization)	Practically non-toxic	No ¹		2712696
Ceriodaphnia dubia	Acute 48-h	Thiamethoxam (≥ 98.6% purity)	48-h LC ₅₀ > 80 000 (0% mortality)	Not toxic up to highest concentration tested.	No ¹	EC ₅₀ Not available (immobilization not recorded)	2842540 (Raby et al. 2018)
Thamnocephalus platyurus	Acute 24-h	Thiamethoxam (98.6%)	24-h EC ₅₀ > 100 000 (0% mortality/immobilization)	Practically non-toxic	No ¹		2712696
Crustaceans - Co	pepoda				J		
Copepoda sp.	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ > 100 000 (0% immobilization)	Practically non-toxic	No ¹	48-h LC ₅₀ > 100 000 μg a.i./L (CI NA)	2712684
Crustaceans - Ost	tracoda		· · · · · · · · · · · · · · · · · · ·		1		
Cyprididae sp.	Acute 48-h	Thiamethoxam (98.6%)	$48-h EC_{50} = 180$ (150–220) (immobilization)	Highly toxic	Yes		2712699
Crustaceans – An	nphipoda	<u> </u>			J		
Hyalella azteca	Sub-chronic 7-d	Thiamethoxam (≥ 95% purity)	$7-d LC_{50} = 215$ (192–240)	Highly toxic	Yes		2753706 (ECCC 2017)
	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 391 (312.1–469.9) (immobilization)	Highly toxic	Yes		2842540 (Raby et al. 2018)

Organism	Exposure	Test Substance	Endpoint value (μg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
			96-h LC ₅₀ = 801.0 (518.7–1083.3)	Highly toxic	No ²		
Gammarus sp.	Acute 48-h	Thiamethoxam (98.6%)	$48-h EC_{50} = 2800$ (1700–4100) (immobilization)	Moderately toxic	Yes	24-h EC ₅₀ = 15 000 (10 000–23 000) μg a.i./L	2712697
Gammarus kischineffensis	Acute 96-h	Formulation Actara 240SC	96-h EC ₅₀ = 3751 (3506–8332) (mortality/ immobilization)	Moderately toxic	Yes	Reported LC ₅₀ , includes mortality + immobility (can therefore be considered as EC ₅₀). 48-h EC ₅₀ = 23 510 (18 840– 27 730) μ g a.i./L	2712706 (Uğurlu et al. 2015)
Crangonyx pseudogracilis	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 420$ (200–870) (immobilization)	Highly toxic	Yes	48-h LC ₅₀ = 20 000 (7280–96 000) μg a.i./L	2712684
	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ = 1010 (310–3350) (mortality/ immobilization)	Moderately toxic	Yes		2712685
Crustaceans -Iso	poda					-	
Asellus aquaticus	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 84$ (44–160) (immobilization)	Highly toxic	Yes	48-h LC ₅₀ = 2300 (820–7320) μg a.i./L	2712684
	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ > 320 (0% mortality/ immobilization)	Not toxic up to highest concentration tested.	No ²		2712685
Caecidotea sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 4775.4 (2976.3– 6574.6) (immobilization)	Moderately toxic	Yes		2842540 (Raby et al. 2018)
			96-h LC ₅₀ > 35 600 (0% mortality)	Not toxic up to highest concentration tested.	No ²		
Crustaceans –De	copoda						
Procambarus	Acute 96-h	Thiamethoxam	$96\text{-h EC}_{50} = 2310$	Moderately	Yes		2712681

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
clarkii		(98.4% purity)	(1630–3280) (mortality/ immobilization)	toxic			
			96-h LC ₅₀ = 2300- 2600 (CI NA)	Moderately toxic	No	EC ₅₀ is the more appropriate endpoint from this study.	
	Acute 96-h (juvenile)	Thiamethoxam (99.5%)	96-h EC ₅₀ = 967 (879–1045) (mortality/ immobilization)	Highly toxic	Yes	Reported LC ₅₀ includes mortality + immobility (can therefore be considered as EC ₅₀).	2712686 (Barbee and Stout 2009)
Rotifera		<u> </u>	·				
Brachionus calyciflorus	Acute 24-h	Thiamethoxam (98.6%)	24-h EC ₅₀ > 100 000 (6.7% mortality/immobilization)	Practically non-toxic	No ¹		2712696
Molluscs							
Lymnea stagnalis	Acute 48-h	Thiamethoxam (98.6%)	48-h EC ₅₀ > 100 000 (10% immobilization)	Practically non-toxic	No ¹		2712699
	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ > 100 000 (0% immobilization)	Practically non-toxic	No ¹	48-h LC ₅₀ > 100 000 μg a.i./L (CI NA)	2712684
Radix peregra	Acute 48-h	Thiamethoxam (98.6%)	48-h EC ₅₀ > 100 000 (0% immobilization)	Practically non-toxic	No¹		2712699
Lampsilis fasciola	Acute 48-h	Thiamethoxam (>95%)	48-h LC ₅₀ > 691 (2.7% mortality)	Not toxic up to highest concentration tested.	Yes		2712688 (Prosser et al. 2016)
Planorbella pilsbryi	Sub-chronic 7-d	Thiamethoxam (>95%)	$7-d LC_{50} = 6195.0$ (2907.8 - 9482.2)	Moderately toxic	Yes	7-d LC ₁₀ mortality = 347.4 (104.4–590.4)	2712688 (Prosser et al. 2016)
Annelids							
Erpobdellidae sp.	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ > 100 000 (37.5% immobilization)	Practically non-toxic	No ¹	48-h LC ₅₀ > 100 000 μg a.i./L (CI NA)	2712684
Lumbriculus sp.	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 7700$ (CI NA)	Moderately toxic	Yes	48-h LC ₅₀ > 32 000 μg a.i./L (CI NA)	2712684

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
			(immobilization)				
Planariidae sp.	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ > 100 000 (0% immobilization)	Practically non-toxic	No ¹	48-h LC ₅₀ > 100 000 μg a.i./L (CI NA)	2712684
Insects - Diptera	l		i i i i i i i i i i i i i i i i i i i			J	
Chironomus riparius	Acute 48-h	Thiamethoxam (97.4%)	$48-h LC_{50} = 35$	Very highly toxic	Yes		1196663
	Acute 48-h	Thiamethoxam	$48-h EC_{50} = 22$	Very highly toxic	Yes		USEPA 2011
	Acute 48-h	Thiamethoxam (99.6%)	$48-h EC_{50} = 86.4$ (74.4–100) (mortality/ immobilization)	Very highly toxic	Yes	Reported LC ₅₀ includes mortality + immobility (can therefore be considered as EC ₅₀).	2720027 (Saraiva et al. 2017)
	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 45$ (CI NA) (immobilization)	Very highly toxic	Yes	48-h LC ₅₀ = 260 (130– 520) μg a.i./L	2712684
	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ = 103 (10–160) (mortality/ immobilization)	Very highly toxic	Yes		2712685
	Acute 48-h	Thiamethoxam WG25 (A9584C; 25.3% thiamethoxam)	$48-h EC_{50} = 38.6$ (12.5–119.5) (immobilization)	Very highly toxic	Yes		2712702
	Acute 48-h	Thiamethoxam FS (A9765N; 617 g a.i./L thiamethoxam)	$48-h EC_{50} = 57.6$ (immobilization)	Very highly toxic	Yes		2712703
	Acute 48-h	Thiamethoxam SC (A9795B; 253 g a.i./L thiamethoxam)	$48-h EC_{50} = 72.9$ (immobilization)	Very highly toxic	Yes		2712704
	Acute 48-h	Thiamethoxam FS (A9765R; 49.5%	48-h EC ₅₀ = 101 (immobilization)	Highly toxic	Yes		2712708

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
		thiamethoxam)					
Chironomus dilutus	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 36.8 (29.4 - 44.3) (immobilization)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			$96-h LC_{50} = 61.9$ (45.4–78.4)	Very highly toxic	No ²		
	Acute 96-h	Thiamethoxam (98.8% purity)	$96-h LC_{50} = 55.3$ (44.0–69.6)	Very highly toxic	Yes		2818524 (Maloney et al. 2017)
Chrionomus tepperi	Acute 24-h	Clothianidin (TI 435, 200 g a.i./L SC)	$24-h LC_{50} = 5.19$ (3.95-6.83)	Very highly toxic	NA	Qualitative endpoint. Cannot be used quantitatively in a risk assessment.	2712705 (Stevens et al. 2005)
Chaoborus sp.	Acute 48-h	Thiamethoxam (98.6% purity)	$48-h LC_{50} = 5500$ ($4400-6600$) (immobilization)	Moderately toxic	Yes		2712699
Chaoborus crystallinus	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 7300$ (5400–10 000) (immobilization)	Moderately toxic	Yes	48-h LC ₅₀ = 11 000 (7900 – 17 000) μg a.i./L	2712684
Aedes sp.	Acute 48-h	Thiamethoxam (≥ 98.6% purity)	48-h LC ₅₀ = 67.4 (42.2–92.5)	Very highly toxic	Yes	EC ₅₀ Not available (immobilization not recorded)	2842540 (Raby et al. 2018)
Aedes aegypti	Acute 24-h	Thiamethoxam (purity NA)	$24-h LC_{50} = 183$ (162-205)	Highly toxic	Yes		2712689 (Riaz et al. 2013)
	Acute 72-h	Thiamethoxam (99.5% purity)	$72-h LC_{50} = 90$ (29–190)	Very highly toxic	NA	Qualitative endpoint. Cannot be used quantitatively in a risk assessment.	2841145 (Ahmed and Matsumura 2012)
	Acute 72-h	Thiamethoxam (99.1% purity)	72 -h $LC_{50} = 298$ (CI NA)	Highly toxic	Yes		2841146 (Uragayala et al. 2015)
Anopheles stephensi	Acute 72-h	Thiamethoxam (99.1% purity)	72-h LC ₅₀ = 52 (CI NA)	Very highly toxic	Yes		2841146 (Uragayala et al. 2015)
			$72-h LC_{50} = 64$ (CI NA)	Very highly toxic	Yes		
Culex quinqefasciatus	Acute 72-h	Thiamethoxam (99.1% purity)	72-h LC ₅₀ = 343 (CI NA)	Highly toxic	Yes		2841146 (Uragayala et al. 2015)

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
Insects - Epheme	roptera				1		
Cloeon sp.	Acute 48-h	Thiamethoxam (98.6% Purity)	48-h EC ₅₀ = 14 (11–17) (immobilization and behavioural changes)	Very highly toxic	Yes		2296375
	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 44.1 (31.2–62.4) (immobilization)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			96-h LC ₅₀ = 4633.6 (1835.8–7431.3)	Moderately toxic	No ²		
Cloeon dipterum	Acute 96-h	Thiamethoxam WG25 (25% thiamethoxam)	96-h EC ₅₀ = 20 (15–26) (immobilization)	Very highly toxic	Yes		2712707 (Van den Brink et al. 2016)
			$96-h LC_{50} = 52$ (CI NA)	Very highly toxic	No ²		
	i I	Thiamethoxam (98.7% purity)	48-h EC ₅₀ = 21 (CI NA) (immobilization)	Very highly toxic	Yes		2712684
			$48-h LC_{50} = 53$ $(38-73)$	Very highly toxic	No ²		
	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 34$ (24–47) (mortality/immobilization)	Very highly toxic	Yes		2712685
Neocloeon triangulifer	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	$96-h EC_{50} = 5.5$ (3.9–7.8)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			$96-h LC_{50} = 5.5$ $(3.9-7.8)$	Very highly toxic	No ²		
Hexagenia sp.	Acute 96-h	Thiamethoxam (≥ 95% purity)	$96-h EC_{50} = 630$ $(140-2900)$	Very highly toxic	Yes	EC ₅₀ based on number of surviving animals after 96 h found inside artificial burrows. Endpoints based on nominal concentrations.	2861091 (Bartlett et al. 2018)
	Acute 96-h	Thiamethoxam (≥ 95% purity)	96-h LC ₅₀ > 10 000	Not toxic up to highest	No ²		

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
				concentration tested.			
	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 35.8 (14.1–57.4) (immobilization)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			96-h LC ₅₀ > 30 800 (0% mortality)	Slightly toxic	No ²		
Caenis sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	$EC_{50} < 23.3$ (100% immobilization)	Highly toxic	Yes ³		2842540 (Raby et al. 2018)
			$96-h LC_{50} = 381.9$ $(185.0-578.8)$	Highly toxic	No ²		
Ephemerella sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	EC ₅₀ < 59 (100% immobilization)	Highly toxic	Yes ³		2842540 (Raby et al. 2018)
			$96-h LC_{50} = 334.9$ (135.9-533.9)	Highly toxic	No ²		
Isonychia bicolor	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	EC ₅₀ < 445.0 (100% immobilization)	Highly toxic	Yes ³		2842540 (Raby et al. 2018)
			96-h LC ₅₀ > 7120 (30% mortality)	Not toxic up to highest concentration tested	No ²		
McCaffertium sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	$96-h EC_{50} = 81.7$ (58.0-115.0)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			96-h LC ₅₀ $>$ 920	Highly toxic	No ²		
Insects - Odonata							
Coenagrionidae	Acute 48-h	Thiamethoxam (98.7% purity)	48-h EC ₅₀ = 980 (CI NA) (immobilization)	Highly toxic	Yes		2712684
			$48-h LC_{50} = 1600$ $(820-2900)$	Moderately toxic	No ²		

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
Insects - Plecopte	ra						
Agnetina, Paragnetina sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	EC ₅₀ < 445 (100% immobilization)	Highly toxic	Yes ³		2842540 (Raby et al. 2018)
	nsects - Hemiptera		96-h LC ₅₀ > 7120.0	Not toxic up to highest concentration tested	No ²		
Insects - Hemipte	ra						
Trichocorixa sp.		Thiamethoxam (≥ 98.6% purity)	$48-h EC_{50} = 56.3$ (34.3 - 68.6) (immobilization)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			48-h LC ₅₀ = 1473.1 (176.3–2769.9)	Moderately toxic	No ²		
Insects - Trichopt	era	•	,		,		
Cheumatopsyche sp.	Acute 96-h	cute 96-h Thiamethoxam (≥ 98.6% purity)	$48-h EC_{50} = 118.5$ (108.8–218.0) (immobilization)	Highly toxic	Yes		2842540 (Raby et al. 2018)
			48-h LC ₅₀ = 170.1 (78.6–261.6)	Highly toxic	No ²		
Micrasema sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	$48-h EC_{50} = 18.5$ (13.1–26.2) (immobilization)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			$48-h LC_{50} = 32.8$ $(26.4-39.2)$	Very highly toxic	No ²		
Insects - Coleopte	ra						
Dytiscidae sp. (adults)	Acute 48-h	Thiamethoxam (98.7% purity)	$48-h EC_{50} = 47$ (22–94) (mortality/ immobilization)	Very highly toxic	Yes		2712685
Gyrinus sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 14.0 $(7.6-20.4)$ (immobilization)	Very highly toxic	Yes		2842540 (Raby et al. 2018)
			96-h LC ₅₀ = 31.0 $(21.9-43.8)$	Very highly toxic	No ²		

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
Stenelmis sp.	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 148 (109.6–186.4) (immobilization)	Highly toxic	Yes		2842540 (Raby et al. 2018)
			$96-h LC_{50} = 148$ (109.6–186.4)	Highly toxic	No ²		
Oligochaetes							
Lumbriculus variegatus	Acute 96-h	Thiamethoxam (≥ 98.6% purity)	96-h EC ₅₀ = 2035.1 (1699.7–2370.6) (immobilization)	Moderately toxic	Yes		2842540 (Raby et al. 2018)
			96-h LC ₅₀ = 3438.2 (3025.5–3850.9)	Moderately toxic	No ²		
Marine inverteb	rates						
Crustaceans - De	ecapoda						
Americamysis bahia	Acute 96-h	e 96-h Thiamethoxam (99.2%)	96-h EC ₅₀ = 4500 (3800 - 5300) (mortality/ swimming behaviour)	Moderately toxic	NA	Incorrectly reported in ERC2007-01 as EC ₅₀ = 5400 μg a.i./L	1196685
			$96-h LC_{50} = 6800$ $(5400-8400)$	Moderately toxic	NA		
Molluscs			······				
Crassostrea virginica	Acute 96-h	Thiamethoxam (99.2%)	96-h EC ₅₀ > 119 000	Practically non-toxic	NA	NOEC = 7400 μg a.i./L	1196674
			•	Chronic			
Freshwater inver	rtebrates						
Crustaceans - Cl	adocera						
Daphnia magna	Chronic 21-d	Thiamethoxam (98.6% purity)	21-d NOEC reprod	uction = 50 000	Yes	Previously assessed by PMRA as NOEC = 100 000 μg/L (ERC2011-05). However, PMRA concurs with USEPA 2011 reported NOEC reproduction = 50 000	1196696

Organism	Exposure	Test Substance	Endpoint value Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)	
					μg/L (MRID 447149-24).		
Crustaceans - Am	phipoda						
Hyalella azteca	28-d Chronic	Thiamethoxam (purity not reported)	28-d NOEC growth = 62.5	Yes	NOEC determined by PMRA from raw data. 28- d EC ₁₀ growth = 71 (35 - 140) µg a.i./L; 28-d EC ₅₀ growth = 200 (160–240) µg a.i/L.	2753706 (ECCC 2017)	
			28-d NOEC mortality = 125	No ²	28-d LC ₁₀ = 160 (120 - 220) µg a.i/L; 28-d LC ₅₀ = 220 (200–240) µg a.i/L.		
Molluscs							
Planorbella Chronic 28-d oilsbryi (ELS)	Thiamethoxam (>95% purity)	28-d EC ₁₀ growth: 21.3 (-30.3–72.9)	Yes	28-d EC ₅₀ growth: 52.1 (- 35.2 –139.4) μg a.i/L.	2712688 (Prosser et al. 2016)		
		Thiamethoxam (>95% purity)	28-d EC ₁₀ biomass: 21.4 (-30.9–73.6)	No ²	28-d EC ₅₀ biomass: 51.3 (-34.9–137.6) µg a.i/L		
Insects - Epheme	roptera						
Cloeon dipterum	Chronic 28-d	Thiamethoxam WG25 (25%	28-d EC ₁₀ immobilization = 0.43 (0.13–1.4)	Yes	28-d EC ₅₀ immobilization = 0.68 (0.38–1.2) μg a.i/L.	2712707 (Van den Brink et al. 2016)	
		thiamethoxam)	$28-d LC_{10} = 0.81 (0.75-0.88)$	No ²	28-d LC ₅₀ = 0.94 (0.88 – 10) μ g a.i/L		
Insects - Diptera							
Chironomus riparius	Chronic 30-d	Thiamethoxam (98.6% purity)	NOEC emergence = 5	Yes	Treated water portion of study. EC ₅₀ emergence/development = 11 µg a.i/L. Previously reported endpoint ERC2007-01.	1196701	
	Chronic 10-d	Thiamethoxam (99.6% purity)	10-d NOEC growth rate = 10.5	No ²		2720027 (Saraiva et al. 2017)	
	Chronic 28-d	Thiamethoxam (99.6% purity)	28-d NOEC emergence = 6.5	Yes			
Chironomus dilutus	14-d Chronic	Thiamethoxam (98.9% purity)	$14\text{-d LC}_{50} = 23.6 (20.4-26.9)$	No ²		2712687 (Cavallero et al. 2017)	
	40-d Life- cycle bioassay	Thiamethoxam (98.9% purity)	40-d EC ₂₀ emergence = 0.48 (0.05–2.76)	Yes	40-d EC ₅₀ emergence = 4.13 (3.53 - 4.76)		

Organism	Exposure	Test Substance	Endpoint value Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
			14-d EC ₂₀ biomass = 10.2 (7.38–14.6)	No ²	14-d EC ₅₀ biomass = 21.39 (17.38 - 28.65)	
			40-d EC ₂₀ sex ratio = 0.31 (0.12–0.75)	No	40-d EC ₅₀ sex ratio = 3.6 (CI NA). This is the lowest endpoint for this species, but emergence will be used for the risk assessment rather than sex ratio.	
	28-d Chronic	Thiamethoxam (98.9% purity)	28-d EC ₂₀ emergence = 4.62 (0.85–6.70)	No ⁴	28-d EC ₅₀ emergence = 8.91 (5.79–12.37) μg a.i./L	2873503 (Maloney et al. 2018)
Chaoborus sp.	Chronic 34-d	Thiamethioxam (98.6% purity)	34-d NOEC emergence = 60	Yes	Endpoints determined by PMRA based on TWA concentrations (Days 0 – 14). 34-d EC ₅₀ emergence = 260 (110 –1160) µg a.i/L NOEC development ≥ 440 µg a.i./L (highest concentration with sufficient survival)	2712701
Studies using tre	eated sediments:					1
Endpoints based	on pore water co	ncentrations:				
Chironomus dilutus	10-d Chronic	Thiamethoxam (99.8% purity)	10-d NOEC growth rate = 120	NA	10-d EC ₅₀ growth > 640 μg a.i/L. Mean measured pore water concentrations.	2712693
			10-d NOEC survival = 360	NA	10-d LC ₅₀ survival = 510 (360–640) μg a.i/L	
Endpoints based	on sediment conc	centrations (µg a.i./	kg dw):			
Chironomus dilutus	10-d Chronic	Thiamethoxam (99.8% purity)	10-d NOEL growth rate = 600 μg a.i./kg dw	NA	10-d EC ₅₀ growth > 2600 μg a.i./kg dw. Mean mesured sediment concentrations.	2712693
			10-d NOEL survival = 1300 μg a.i./kg dw	NA	10-d LC ₅₀ survival = 2000 (1900–2100) μg a.i./kg dw	
Chironomus riparius	Chronic 30-d	Thiamethoxam (98.6% purity)	30-d NOEL emergence/development = 43 μg a.i./kg dw	NA	Treated sediment portion of study. EC ₅₀ emergence/development =	1196701

Organism	Exposure	Test Substance		egree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
			, the state of the			99 µg a.i./kg dw. Previously reported endpoint ERC2007-01. Based on nominal concentrations.	
Microcosm or me	esocosm tests						
Natural species assemblage	93-d Chronic	Thiamethoxam 25 WG (A9584C; 25% thiamethoxam)	93-d NOEC community	= 9.4	NA	Single application to outdoor mesocosms. NOEC based on a significant reduction in chironomid emergence at the 34 µg a.i./L treatment on Day 15. Emergence was comparable with the controls on all other sampling occasions. There was an insufficient abundance of Ephemeropterans to assess effects on this sensitive group of insects. NOEC determined by PMRA as TWA concentration due to loss of test material over time in mesocosms.	2712709, 2712710
Natural species assemblage Marine invertebr	35-d Chronic	Thiamethoxam 25 WG (A9584C; 25% thiamethoxam)	35-d NOEC larval abundance/emergence = 0	0.3	NA	Multiple applications to outdoor mesocosms. NOEC based on significant reductions in larval mayfly abundance and emergence at 1.0 µg a.i/L treatments. No evidence of recovery in the study.	2681280

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Data used in SSD	Comments	Reference PMRA# (Publication)
Crustaceans - Dec	apoda						
Americamysis	28-d Life-	Thiamethoxam	28-d NOEC survival =	= 560	NA	LOEC survival = 1100 μg	2712712
bahia	cycle bioassay	(99.8% purity)				a.i./L	

NA: Not applicable, an SSD was not constructed for these taxa

- 1 Unbound endpoint was not included as a more sensitive endpoint is available for this species or a similar taxa from another study (as per EFSA 2013 guidance)
- 2 A more sensitive endpoint is available from the same study
- 3 Unbound endpoint was included as it represents the most sensitive endpoint for this unique species (as per EFSA 2013 guidance) 4 28-d EC20 for Chironomus dilutus was not included in a geomean with the 40-d EC20 for this same species as the difference in toxicity is thought to be due to the longer exposure period in the latter study.

The studies by Cavallaro et al. (2017) and Maloney et al. (2018) were conducted in the same laboratory using the same protocols.

Table A.3-9 Effects of major transformation products of thiamethoxam on aquatic invertebrates

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Comments	Reference PMRA# (Publication)
			Acute			
Freshwater invert	ebrates					
Crustaceans - Cla	docera					
Daphnia magna	Acute 48-h	CGA 322704 [clothianidin] (99.8% purity)	$48-h EC_{50} > 100 000$ (0% mortality/ immobilization)	Practically non- toxic		2712674
		CGA 355190 (99 ± 2% purity)	$48-h EC_{50} > 100 000$ (0% mortality/ immobilization)	Practically non-toxic		2712679
		NOA 407475 (99.9% purity)	48-h EC ₅₀ = 82 900 (68 400– 102 300) (mortality/ immobilization)	Slightly toxic	Slightly toxic 27	
		NOA 459602 (99 ± 2% purity)	$48-h EC_{50} > 120 000$ (0% mortality/ immobilization)	Practically non- toxic		2712677
		CGA 282149 (CA2343; 96.7% purity)	$48-h EC_{50} > 100 000$ (0% mortality/ immobilization)	Practically non- toxic		2712670
Crustaceans -Isop	oda					
Asellus aquaticus	Acute 48-h	CGA 322704 [clothianidin] (99% purity)	$48-h EC_{50} = 67 (43-105)$ (mortality/immobilization)	Highly toxic		2712685
Insects - Ephemer	optera					
Cloeon dipterum	Acute 48-h	CGA 322704 [clothianidin] (99% purity)	48-h EC ₅₀ = 12 (8–16) (mortality/immobilization)	Very highly toxic		2712685
Insects - Coleopte	ra			<u> </u>		
Dytiscidae sp. (adults)	Acute 48-h	CGA 322704 (99% purity)	$48-h EC_{50} = 7 (2-14)$ (mortality/ immobilization)	Very highly toxic		2712685
Insects - Diptera				***************************************		
Chironomus	Acute 48-h	CGA 355190 (98.7% purity)	48-h EC ₅₀ = 4100 (2600–6400)	Moderately toxic		1529851

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Degree of toxicity	Comments	Reference PMRA (Publication)
riparius			(mortality/immobilization)	•		
*		NOA 404617 (99.7% purity)	$48-h EC_{50} > 105 000$ Pra	actically non-		1529853
			(10% immobilization) tox	xic		
		CGA 282149 (CA2343; 99.3%	$48-h EC_{50} > 100 000$ Pra	actically non-		2712691
		purity)	(10% immobilization) tox	xic		
		NOA 421275 (98% purity)	48-h EC ₅₀ > 100 000 Pra	actically non-		2712692
		, , ,	(10% immobilization) tox	xic		
		CGA 322704 [clothianidin]	$48-hEC_{50} = 14 (4-29)$ Hi	ghly toxic		2712685
		(99% purity)	(mortality/immobilization)	•		
			Chronic			
Freshwater inverte	brates					
Crustaceans - Clad						
Daphnia magna	Chronic 21-d	CGA 282149 (CA2343; 96.7%	21-d NOEC length = 56000			2712680
		purity)				
Insects - Diptera						
Chironomus	Chronic 26-d	CGA 353042 (94% purity)	26-d NOEC emergence = 56 400 (highe	st concentration	NOEC at highest	1529852
riparius			tested)		concentration tested.	
	Chronic 24-d	NOA 459602 (99 ± 2% purity)	24-d NOEC emergence/development = :	50 000	EC_{50} emergence = 56 000	2712682
					μg/L; EC ₅₀ development	
					NA	
	Chronic 28-d	CGA 322704 [clothianidin]	28-d NOEC emergence/sex ratio = 0.55	Recoveries were low and	2712700	
		(98% purity)			reported endpoints were	
					based on nominal	
					concentrations.	
					Significant effects on	
					emergence rate and sex	
					ratio were observed at 2	
					μg a.i./L nominal. NOEC	
					determined by PMRA	
					based on mean measured	
					concentrations from Day	
					0 and 7 at 0.67 mg a.i./L	
					nominal treatment. 28-d	
					EC_{50} emergence = 1.2 μg	
					a.i./L nominal.	
	Chronic 28-d	SYN 501406 (NOA 501406;	28-d NOEC emergence = 1100			2712683
		98% purity)				
Studies using treate	ed sediments:	1 2/			1	
	n sediment concentr	ations:				
Chironomus	Chronic 28-d	NOA 407475 (99.9% purity)	28-d NOEC emergence/development = 4	410 μg/kg dw	NOEC at highest	1529854
riparius		((highest concentration tested)	10 6	concentration tested;	
*					PMRA DER and	

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L) Degree o toxicity	1 (111111111111111111111111111111111111	Reference PMRA# (Publication)
				endpoint as 410 µg/L when it should be 410 µg/kg sediment (mean measured).	
	Chronic 28-d	CGA 322704 [clothianidin] (99% purity)	28-d NOEC emergence = 15 μg/kg dw	EC ₅₀ emergence = 25 μg/kg dw. Recoveries were low and endpoints were based on nominal exposure concentrations. The endpoints cannot be used quantitatively in a risk assessment, but may be used as weight of evidence only.	2712695

Table A.3-10 Summary of screening level risk of thiamethoxam to aquatic invertebrates exposed at a range of seasonal application rates

Organism	Exposure	Species	Endpoint eported (μg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	EEC² (μg a.i./L)	RQ	LOC Exceeded
Freshwater org	anisms						
Invertebrates Acute	Acute	37 invertebrate species	$HC_5 = 9.0$	9.0	0.563 (minimum seed treatment rate)	0.06	No
				18.8 (maximum seed treatment rate)	2.1	Yes	
				22.2 (maximum foliar rate)	2.5	Yes	
	Chronic	7 invertebrate species	$HC_5 = 0.026$	0.026	0.563 (minimum seed treatment rate)	22	Yes
					18.8 (maximum seed treatment rate)	723	Yes
					22.2 (maximum foliar rate)	854	Yes
Most sensitive single	Acute	Mayfly Neocloeon triangulifer	96-h $EC_{50} = 5.5$	2.8	0.563 (minimum seed treatment rate)	0.2	No
invertebrate species (for					18.8 (maximum seed treatment rate)	6.8	Yes
comparison					22.2 (maximum	8.1	Yes

Organism	Exposure	Species	Endpoint eported (µg a.i./L)	Endpoint for RA¹ (µg a.i./L)	EEC² (μg a.i./L)	RQ	LOC Exceeded
against SSD HC5					foliar rate)		
values).	Chronic	Mayfly Cloeon dipterum	28-d EC ₁₀ immobilization = 0.43	0.43	0.563 (minimum seed treatment rate)	1.3	Yes
	_			18.8 (maximum seed treatment rate)	44	Yes	
					22.2 (maximum foliar rate)	52	Yes
Marine/Estuarir	neorganisms						
Mysid shrimp	Acute	Mysidopsis bahia	$96-h EC_{50} = 4500$	2250	0.563 (minimum seed treatment rate)	< 0.01	No
					18.8 (maximum seed treatment rate)	< 0.01	No
					22.2 (maximum foliar rate)	0.01	No
	Chronic		28-d NOEC survival = 560	560	0.563 (minimum seed treatment rate)	< 0.01	No
					18.8 (maximum seed treatment rate)	0.03	No
					22.2 (maximum foliar rate)	0.04	No

¹ Endpoints used in the acute exposure risk assessment (RA) are derived by dividing the EC50 or LC50 from the appropriate laboratory study by a factor of two (2) for aquatic invertebrates. The HC5 is the 5th percentile of the species sensitivity distribution for 24 – 96-h and 7-d sub-chronic LC50 or EC50 endpoints (acute exposures), or for 14 – 40-d NOEC or EC10 endpoints (chronic exposures).

² EEC based on an 80 cm water depth.

Bolded values indicates an exceedence of the level of concern (RQ = 1).

Table A.3-11 Summary of screening level risk of major thiamethoxam transformation products to aquatic invertebrates exposed at the highest seasonal cumulative rate for all crops (foliar application rate of 178.1 g a.i./ha)

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	EEC² (μg a.i./L)	RQ	LOC Exceeded
			Acute				
Freshwater invertel	rates						
Crustaceans - Clade	ocera						
Daphnia magna	Acute 48-h	CGA 322704 (99.8% purity)	$48-h EC_{50} > 100 000$	50 000	19.0	< 0.01	No

Organism	Exposure	Test Substance	Endpoint value (μg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	EEC² (μg a.i./L)	RQ	LOC Exceeded
	Acute 48-h	CGA 355190 (99 ± 2% purity)	48-h EC ₅₀ > 100 000	50 000	18.8	< 0.01	No
	Acute 48-h	NOA 407475 (99.9% purity)	48-h EC ₅₀ = 82 900 (68 400 – 102 300)	41 450	18.8	< 0.01	No
	Acute 48-h	NOA 459602 (99 ± 2% purity)	48-h EC ₅₀ > 120 000	60 000	25.7	< 0.01	No
	Acute 48-h	CGA 282149 (CA2343; 96.7% purity)	48-h EC ₅₀ > 100 000	50 000	12.2	< 0.01	No
Amphipods/Isopod	S						
Asellus aquaticus	Acute 48-h	CGA 322704 (99% purity)	$48-h EC_{50} = 67 (43-105)$	33.5	19.0	0.57	No
Insects - Ephemero	pterans	1	- F		1		
Cloeon dipterum	Acute 48-h	CGA 322704 (99% purity)	$48-h EC_{50} = 12 (8-16)$	6	19.0	3.2	Yes
Insects - Coleopter	ans						
Dytiscidae	Acute 48-h	CGA 322704 (99% purity)	$48-h EC_{50} = 7 (2-14)$	3.5	19.0	5.4	Yes
Insects – Diptera			L			<u> </u>	
Chironomus riparius	Acute 48-h	CGA 355190 (98.7% purity)	48-h EC ₅₀ = 4100 (2600 – 6400)	2050	18.8	0.01	No
•	Acute 48-h	NOA 404617 (99.7% purity)	48-h EC ₅₀ > 105 000	52 500	18.0	< 0.01	No
	Acute 48-h	CGA 282149 (CA2343; 99.3% purity)	48-h EC ₅₀ > 100 000	50 000	12.2	< 0.01	No
	Acute 48-h	NOA 421275 (98% purity)	48-h LC ₅₀ > 100 000	50 000	15.6	< 0.01	No
	Acute 48-h	CGA 322704 (99% purity)	$48\text{-hEC}_{50} = 14 (4 - 29)$	7	19.0	2.7	Yes
			Chronic				
Freshwater inverte	brates						
Crustaceans – Clad	locera						
Daphnia magna	21-d Chronic	CGA 282149 (CA2343; 96.7% purity)	21-d NOEC length = 56 000	56 000	12.2	< 0.01	No
Insects - Diptera							
Chironomus riparius larvae	26-d Chronic	CGA 353042 (94% purity)	26-d NOEC emergence = 56 400	56 400	8.8	< 0.01	No
	24-d Chronic	NOA 459602 (99 ± 2% purity)	24-d NOEC emergence/ development = 50 000	50 000	25.7	< 0.01	No

Organism	Exposure	Test Substance	Endpoint value (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	EEC² (μg a.i./L)	RQ	LOC Exceeded
	28-d	CGA 322704 (98% purity)	28-d NOEC emergence/sex	0.55	19.0	34.6	Yes
	Chronic		ratio = 0.55				
	28-d	SYN 501406 (NOA 501406; 98%	28-d NOEC emergence =	1100	22.5	0.02	No
	Chronic	purity)	1100				

¹ Endpoints used in the acute exposure risk assessment (RA) are derived by dividing the EC50 or LC50 from the appropriate laboratory study by a factor of two (2) for aquatic invertebrates.

Table A.3-12 Refined risk assessment of thiamethoxam for aquatic invertebrates from predicted levels of spray drift

Organism	Exposure	Species	Endpoint eported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	EEC² (μg a.i./L)	RQ	LOC Exceeded
Freshwater orga	ınisms						
Invertebrates	Acute	37 invertebrate species	$HC_5 = 9.0$	9.0	2.06 (field sprayer)	0.2	No
					16.4 (airblast	1.8	Yes
					sprayer)		
					2.22 (aerial sprayer)	0.2	No
	Chronic	7 invertebrate species	$HC_5 = 0.026$	0.026	2.06 (field sprayer)	79	Yes
					16.4 (airblast	632	Yes
					sprayer)		
					2.22 (aerial sprayer)	85	Yes
Most sensitive	Acute	Mayfly	$96-h EC_{50} = 5.5$	2.8	2.06 (field sprayer)	0.8	No
single		Neocloeon triangulifer			16.4 (airblast	6.0	Yes
nvertebrate					sprayer)		
species (for					2.22 (aerial sprayer)	0.8	No
comparison	Chronic	Mayfly	28-d EC ₁₀ immobilization =	0.43	2.06 (field sprayer)	4.8	Yes
ngainst SSD HC5		Cloeon dipterum	0.43		16.4 (airblast	38	Yes
alues).					sprayer)		
					2.22 (aerial sprayer)	5.2	Yes
Microcosm or m	esocosm tests						
nvertebrates	Chronic	Emergent insects	35-d NOEC = 0.30 (reductions	0.30	2.06 (field sprayer)	6,9	Yes
		-	in mayfly abundance and		16.3 (airblast	55	Yes
			emergence)		sprayer)		
					2.22 (aerial sprayer)	7.4	Yes

¹ Endpoints used in the acute exposure risk assessment (RA) are derived by dividing the EC50 or LC50 from the appropriate laboratory study by a factor of two (2) for aquatic invertebrates. The HC5 is the 5th percentile of the species sensitivity distribution for 24 – 96-h and 7-d sub-chronic LC50 or EC50 endpoints (acute exposures), or for 21 – 40-d NOEC or EC10/EC20 endpoints (chronic exposures).

² EECs based on an 80 cm water depth. EECs for transformation products based on highest thiamethoxam screening-level EEC for foliar spray rate of 2 × 96.25 or 178 g a.i./ha thiamethoxam (maximum cumulative rate). EECs for individual transformation products adjusted for the molecular-weight ratio relative to thiamethoxam. For example, EEC in 80 cm for CGA 322704 = 22.2 μg a.i./L thiamethoxam × (249.7 g/mol CGA 322704 / 291.7 g/mol thiamethoxam) = 19.0 μg/L CGA 322704.

² EECs based on an 80 cm water depth. EECs based on maximum cumulative use rates for each application method: Field sprayer = 1×150 g a.i./ha (outdoor ornamentals), EEC in 80 cm = $18.8 \mu g$ a.i./L; airblast = 2×96.25 g a.i./ha with 10-d application interval and 80^{th} percentile $t_{1/2} = 42.8$ d (e.g. pome fruit), EEC in 80 cm = $22.2 \mu g$ a.i./L; aerial sprayer = 3×25.38 g a.i./ha with 7-d application interval and 80^{th} percentile $t_{1/2} = 42.8$ d (dry beans), EEC in 80 cm = $8.54 \mu g$ a.i./L. EECs were then adjusted for expected spray drift deposit 1 m downwind: Field sprayer = 11% (ASAE Fine spray quality); aerial sprayer = 26% (ASAE Fine spray quality); airblast = 74% (early season). Bolded values indicates an exceedence of the level of concern (RQ = 1).

Table A.3-13 Refined risk assessment of thiamethoxam for aquatic invertebrates from predicted levels of pesticide runoff

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC³ (µg a.i./L)	RQ	LOC exceeded		
Freshwater org	ganisms												
Invertebrates	Acute	37 invertebrate species	$HC_5 = 9.0$	9.0	Foliar	Apple	2 × 96.25 g a.i./ha at a 10-d interval	BC	0.23	0.0	No		
		•				Potato	2 × 26.25 g a.i./ha at a 7-	Prairie- MB	1.8	0.2	No		
							d interval	Atlantic	1.7	0.2	No		
						Soybean	3 × 25.38 g a.i./ha at a 7-	Prairie- MB	2.4	0.3	No		
							d interval	ON	2	0.2	No		
								QC	2	0.2	No		
						Bell pepper	2 × 70 g a.i./ha at a 7-	Prairie- MB	4.8	0.5	No		
						d interval	ON	4.9	0.5	No			
								QC	4.6	0.5	No		
						Blueberry	2 × 70 g a.i./ha at a 7- d interval	Atlantic	8	0.9	No		
					Transplan t water	Bell pepper	1 × 117 g a.i./ha	Prairie- MB	2.9	0.3	No		
						t water			ON	2.8	0.3	No	
											QC	2.9	0.3
					In-	Potato	1 × 140 g	Prairie	2.6	0.3	No		
					furrow/so il drench		a.i./ha	Atlantic	10	1.1	Yes		
					plus	Bell	1 × 150 g	Prairie	3.7	0.4	No		
					irrigation	pepper	a.i./ha	ON	3.5	0.4	No		
							QC	3.7	0.4	No			
						Blueberry (lowbush)	1 × 140 g a.i./ha	Atlantic	10	1.1	Yes		

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
					Seed treatment	Barley	1 × 36.3 g a.i./ha	BC	0.128	0.0	No
						Winter wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.416	0.0	No
								Prairie- SK	0.472	0.1	No
						Spring wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.208	0.0	No
								Prairie- SK	0.184	0.0	No
						Peas	1 × 150 g	BC	0.0032	0.0	No
							a.i./ha	Prairie- MB	0.36	0.0	No
								Prairie- SK	0.304	0.0	No
								ON	0.096	0.0	No
								QC	0.104	0.0	No
								Atlantic	2.64	0.3	No
						Beans	1 × 50 g	BC	0.0008	0.0	No
							a.i./ha	Prairie- MB	0.0216	0.0	No
								Prairie- SK	0.0176	0.0	No
								ON	0.144	0.0	No
								QC	0.184	0.0	No
								Atlantic	0.448	0.0	No
						Canola	1 × 32.3 g	MB	0.512	0.1	No
							a.i./ha	SK	0.368	0.0	No
								ON	2.4	0.3	No
								QC	3.6	0.4	No
						Potato	1 × 117.12 g a.i./ha	Prairie- MB	0.0008	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
								Atlantic	0.0304	0.0	No
						Corn ⁴	1 × 118.3 g	ON	0.0776	0.0	No
							a.i./ha	QC	0.0632	0.0	No
						Corn ⁵	1 × 118.3 g	ON	0.792	0.1	No
							a.i./ha	QC	0.744	0.1	No
						Sweet	1 × 7.6 g	BC	0.0008	0.0	No
						com ⁴	a.i./ha	MB	0.0168	0.0	No
								ON	0.0048	0.0	No
								QC	0.004	0.0	No
								Atlantic	0.0528	0.0	No
						Sweet	1 × 7.6 g	BC	0.0032	0.0	No
						com ⁵	a.i./ha	MB	0.104	0.0	No
								ON	0.0512	0.0	No
								QC	0.0472	0.0	No
								Atlantic	0.208	0.0	No
						Soybean ⁴	1 × 64 g a.i./ha	Prairie- MB	0.36	0.0	No
								Prairie- SK	0.28	0.0	No
								ON	0.2	0.0	No
								QC	0.168	0.0	No
								Atlantic	0.44	0.0	No
						Soybean ⁵	1 × 64 g a.i./ha	Prairie- MB	0.784	0.1	No
								Prairie- SK	0.68	0.1	No
								ON	0.736	0.1	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC's (µg a.i./L)	RQ	LOC exceeded
								QC	0.68	0.1	No
								Atlantic	1.76	0.2	No
	Chronic	7 invertebrate species	$HC_5 = 0.026$	0.026	Foliar	Apple	2 × 96.25 g a.i./ha at a 10-d interval	BC	0.19	7.3	Yes
		•				Potato	2 × 26.25 g a.i./ha at a 7-	Prairie- MB	1.4	53.8	Yes
							d interval	Atlantic	1.5	57.7	Yes
						Soybean	3 × 25.38 g a.i./ha at a 7-	Prairie- MB	2	76.9	Yes
							d interval	ON	1.6	61.5	Yes
								QC	1.7	65.4	Yes
						Bell pepper	2 × 70 g a.i./ha at a 7-	Prairie- MB	4	153.8	Yes
							d interval	ON	4.6	176.9	Yes
								QC	4.1	157.7	Yes
						Blueberry	2 × 70 g a.i./ha at a 7- d interval	Atlantic	7.1	273.1	Yes
					Transplan	Bell	1 × 117 g	Prairie-	2.5	96.2	Yes
					t water	pepper	a.i./ha	MB	2.4	02.2	X 7
	İ							ON	2.4	92.3	Yes
					T	D	1 140	QC C	2.4	92.3	Yes
					In- furrow/so	Potato	1 × 140 g a.i./ha	Prairie	2.3	88.5	Yes
					il drench	Bell		Atlantic Prairie	3.2	346.2 123.1	Yes Yes
					plus irrigation	pepper	1 × 150 g a.i./ha	ON			
					nngauon	1 11		QC	3.1	119.2 115.4	Yes Yes
						Dlyohom	1 × 140 g	Atlantic	9	346.2	Yes
						Blueberry (lowbush)	a.i./ha	Auanuc	9	340.2	res
					Seed treatment	Barley	1 × 36.3 g a.i./ha	BC	0.112	4.3	Yes

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC³ (µg a.i./L)	RQ	LOC exceeded
						Winter wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.36	13.8	Yes
								Prairie- SK	0.44	16.9	Yes
						Spring wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.192	7.4	Yes
								Prairie- SK	0.168	6.5	Yes
						Peas	1 × 150 g	BC	0.0024	0.1	No
							a.i./ha	Prairie- MB	0.336	12.9	Yes
								Prairie- SK	0.256	9.8	Yes
								ON	0.088	3.4	Yes
								QC	0.096	3.7	Yes
								Atlantic	2.4	92.3	Yes
						Beans	1 × 50 g	BC	0.0008	0.0	No
							a.i./ha	Prairie- MB	0.02	0.8	No
								Prairie- SK	0.0152	0.6	No
								ON	0.128	4.9	Yes
								QC	0.176	6.8	Yes
								Atlantic	0.408	15.7	Yes
						Canola	1 × 32.3 g	MB	0.464	17.8	Yes
							a.i./ha	SK	0.32	12.3	Yes
								ON	2.08	80.0	Yes
								QC	3.36	129.2	Yes
						Potato	1 × 117.12 g a.i./ha	Prairie- MB	0.0008	0.0	No
								Atlantic	0.0264	1.0	Yes
						Corn ⁴	1 × 118.3 g	ON	0.0688	2.6	Yes

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
							a.i./ha	QC	0.06	2.3	Yes
						Corn ⁵	1 × 118.3 g	ON	0.728	28.0	Yes
							a.i./ha	QC	0.688	26.5	Yes
						Sweet	1 × 7.6 g	BC	0.0008	0.0	No
						com ⁴	a.i./ha	MB	0.016	0.6	No
								ON	0.0048	0.2	No
								QC	0.004	0.2	No
								Atlantic	0.048	1.8	Yes
						Sweet	1 × 7.6 g	BC	0.0024	0.1	No
						corn ⁵	a.i./ha	MB	0.096	3.7	Yes
								ON	0.0472	1.8	Yes
								QC	0.044	1.7	Yes
								Atlantic	0.192	7.4	Yes
						Soybean ⁴	1 × 64 g a.i./ha	Prairie- MB	0.312	12.0	Yes
								Prairie- SK	0.232	8.9	Yes
								ON	0.176	6.8	Yes
								QC	0.136	5.2	Yes
								Atlantic	0.4	15.4	Yes
						Soybean ⁵	1 × 64 g a.i./ha	Prairie- MB	0.68	26.2	Yes
								Prairie- SK	0.576	22.2	Yes
								ON	0.656	25.2	Yes
								QC	0.56	21.5	Yes
								Atlantic	1.6	61.5	Yes
Most sensitive single invertebrate	Acute	Mayfly Neocloeon triangulifer	96-h EC ₅₀ = 5.5	2.8	Foliar	Apple	2 × 96.25 g a.i./ha at a 10-d interval	BC	0.23	0.1	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC' (µg a.i./L)	RQ	LOC exceeded
species (for comparison						Potato	2 × 26.25 g a.i./ha at a 7-	Prairie- MB	1.8	0.6	No
against SSD							d interval	Atlantic	1.7	0.6	No
HC ₅ values).						Soybean	3 × 25.38 g a.i./ha at a 7-	Prairie- MB	2.4	0.9	No
							d interval	ON	2	0.7	No
								QC	2	0.7	No
						Bell pepper	2 × 70 g a.i./ha at a 7-	Prairie- MB	4.8	1.7	Yes
							d interval	ON	4.9	1.8	Yes
								QC	4.6	1.6	Yes
						Blueberry	2 × 70 g a.i./ha at a 7- d interval	Atlantic	8	2.9	Yes
					Transplan t water	Bell pepper	1 × 117 g a.i./ha	Prairie- MB	2.9	1.0	Yes
								ON	2.8	1.0	Yes
								QC	2.9	1.0	Yes
					In-	Potato	1 × 140 g	Prairie	2.6	0.9	No
					furrow/so il drench		a.i./ha	Atlantic	10	3.6	Yes
					plus	Bell	1 × 150 g	Prairie	3.7	1.3	Yes
					irrigation	pepper	a.i./ha	ON	3.5	1.3	Yes
								QC	3.7	1.3	Yes
						Blueberry (lowbush)	1 × 140 g a.i./ha	Atlantic	10	3.6	Yes
					Seed treatment	Barley	1 × 36.3 g a.i./ha	BC	0.128	0.0	No
						Winter wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.416	0.1	No
								Prairie- SK	0.472	0.2	No
						Spring	1 × 52.5 g	Prairie-	0.208	0.1	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i/L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (µg a.i./L)	RQ	LOC exceeded
						wheat	a.i./ha	MB			
								Prairie- SK	0.184	0.1	No
						Peas	1 × 150 g	BC	0.0032	0.0	No
							a.i./ha	Prairie- MB	0.36	0.1	No
								Prairie- SK	0.304	0.1	No
								ON	0.096	0.0	No
								QC	0.104	0.0	No
								Atlantic	2.64	0.9	No
						Beans	1 × 50 g	BC	0.0008	0.0	No
							a.i./ha	Prairie- MB	0.0216	0.0	No
								Prairie- SK	0.0176	0.0	No
								ON	0.144	0.1	No
								QC	0.184	0.1	No
								Atlantic	0.448	0.2	No
						Canola	1 × 32.3 g	MB	0.512	0.2	No
							a.i./ha	SK	0.368	0.1	No
								ON	2.4	0.9	No
								QC	3.6	1.3	Yes
						Potato	1 × 117.12 g a.i./ha	Prairie- MB	0.0008	0.0	No
								Atlantic	0.0304	0.0	No
						Corn ⁴	1 × 118.3 g	ON	0.0776	0.0	No
							a.i./ha	QC	0.0632	0.0	No
						Corn ⁵	1 × 118.3 g a.i./ha	ON	0.792	0.3	No
							a.1./11d	QC	0.744	0.3	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
						Sweet	1 × 7.6 g	BC	0.0008	0.0	No
						Sweet com ⁵	a.i./ha	MB	0.0168	0.0	No
								ON	0.0048	0.0	No
								QC	0.004	0.0	No
								Atlantic	0.0528	0.0	No
							1 × 7.6 g a.i./ha	BC	0.0032	0.0	No
								MB	0.104	0.0	No
								ON	0.0512	0.0	No
								QC	0.0472	0.0	No
								Atlantic	0.208	0.1	No
						Soybean ⁴	1 × 64 g a.i./ha	Prairie- MB	0.36	0.1	No
								Prairie- SK	0.28	0.1	No
								ON	0.2	0.1	No
								QC	0.168	0.1	No
								Atlantic	0.44	0.2	No
						Soybean ⁵	1 × 64 g a.i./ha	Prairie- MB	0.784	0.3	No
								Prairie- SK	0.68	0.2	No
								ON	0.736	0.3	No
								QC	0.68	0.2	No
								Atlantic	1.76	0,6	No
	Chronic	Mayfly (Cloeon dipterum)	28-d EC ₁₀ immobilizatio n = 0.43	0.43	Foliar	Apple	2 × 96.25 g a.i./ha at a 10-d interval	BC	0.19	0.4	No
		ong vor sinn)				Potato	2 × 26.25 g a.i./ha at a 7-	Prairie- MB	1.4	3.3	Yes
							d interval	Atlantic	1.5	3.5	Yes
						Soybean	$3 \times 25.38 \text{ g}$	Prairie-	2	4.7	Yes

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (µg a.i./L)	RQ	LOC exceeded
							a.i./ha at a 7 d	MB			
							interval	ON	1.6	3.7	Yes
								QC	1.7	4.0	Yes
						Bell pepper	2 × 70 g a.i./ha at a 7-	Prairie- MB	4	9.3	Yes
							d interval	ON	4.6	10.7	Yes
								QC	4.1	9.5	Yes
						a.i./ha at a d interval	2 × 70 g a.i./ha at a 7- d interval	Atlantic	7.1	16.5	Yes
					Transplan t water		1 × 117 g a.i./ha	Prairie- MB	2.5	5.8	Yes
								ON	2.4	5.6	Yes
							1140	QC	2.4	5.6	Yes
					In-	Potato	Potato 1 × 140 g a.i./ha	Prairie	2.3	5.3	Yes
					furrow/so il drench			Atlantic	9	20.9	Yes
					plus	Bell		Prairie	3.2	7.4	Yes
					irrigation	pepper	a.i./ha	ON	3.1	7.2	Yes
								QC	3	7.0	Yes
						Blueberry (lowbush)	1 × 140 g a.i./ha	Atlantic	9	20.9	Yes
					Seed treatment	Barley	1 × 36.3 g a.i./ha	BC	0.112	0.3	No
						Winter wheat Spring wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.36	0.8	No
								Prairie- SK	0.44	1.0	Yes
							1 × 52.5 g a.i./ha	Prairie- MB	0.192	0.4	No
							SAAT AND	Prairie- SK	0.168	0.4	No
						Peas	1 × 150 g	BC	0.0024	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
							a.i./ha	Prairie- MB	0.336	0.8	No
							1 × 50 g a.i./ha	Prairie- SK	0.256	0.6	No
								ON	0.088	0.2	No
								QC	0.096	0.2	No
								Atlantic	2.4	5.6	Yes
						Beans		BC	0.0008	0.0	No
								Prairie- MB	0.02	0.0	No
								Prairie- SK	0.0152	0.0	No
								ON	0.128	0.3	No
								QC	0.176	0.4	No
								Atlantic	0.408	0.9	No
						Canola	1 × 32.3 g	MB	0.464	1.1	Yes
							a.i./ha	SK	0.32	0.7	No
								ON	2.08	4.8	Yes
								QC	3.36	7.8	Yes
						Potato	1 × 117.12 g a.i./ha	Prairie- MB	0.0008	0.0	No
								Atlantic	0.0264	0.1	No
						Corn ⁴	1 × 118.3 g	ON	0.0688	0.2	No
							a.i./ha	QC	0.06	0.1	No
						Corn ⁵	1 × 118.3 g a.i./ha 1 × 7.6 g a.i./ha	ON	0.736	1.7	Yes
								QC	0.688	1.6	Yes
						Sweet		BC	0.0008	0.0	No
						corn ⁴		MB	0.016	0.0	No
								ON	0.0048	0.0	No
								QC	0.004	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC³ (µg a.i./L)	RQ	LOC exceeded
								Atlantic	0.048	0.1	No
						Sweet	1 × 7.6 g	BC	0.0024	0.0	No
						corn ⁵	a.i./ha $1 \times 64 \text{ g}$ a.i./ha	MB	0.096	0.2	No
								ON	0.0472	0.1	No
								QC	0.044	0.1	No
								Atlantic	0.192	0.4	No
						Soybean ⁴		Prairie- MB	0.312	0.7	No
								Prairie- SK	0.232	0.5	No
								ON	0.176	0.4	No
								QC	0.136	0.3	No
								Atlantic	0.4	0.9	No
						Soybean ⁵	1 × 64 g a.i./ha	Prairie- MB	0.68	1.6	Yes
								Prairie- SK	0.576	1.3	Yes
								ON	0.656	1.5	Yes
								QC	0.56	1.3	Yes
								Atlantic	1.6	3.7	Yes
Studies using t	reated sedim	ents								·	
Chironomid	Chronic	Chironomus dilutus	10-d NOEC growth rate = 120 (pore	120	Foliar	Apple	a.i./ha at a 10-d interval to 2 × 26.25 g a.i./ha at a 7- d interval	BC	0.055	0.0	No
			water concentrations			Potato		Prairie- MB	0.64	0.0	No
)					Atlantic	0.67	0.0	No
						Soybean		Prairie- MB	0.96	0.0	No
								ON	0.65	0.0	No
								QC	0.7	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC' (µg a.i./L)	RQ	LOC exceeded
						Bell pepper	2 × 70 g a.i./ha at a 7-	Prairie- MB	1.7	0.0	No
							d interval	ON	2.1	0.0	No
								QC	2.1	0.0	No
						Blueberry	2 × 70 g a.i./ha at a 7- d interval	Atlantic	2.7	0.0	No
					Transplan t water	Bell pepper	1 × 117 g a.i./ha	Prairie- MB	0.78	0.0	No
								ON	0.85	0.0	No
								QC	0.64	0.0	No
					In-	rrow/so drench is Bell	1 × 140 g	Prairie	0.82	0.0	No
							a.i./ha	Atlantic	3.6	0.0	No
					plus			Prairie	1	0.0	No
					irrigation		pepper a.i./ha	ON	1.1	0.0	No
								QC	0.82	0.0	No
						Blueberry (lowbush)	1 × 140 g a.i./ha	Atlantic	4.2	0.0	No
					Seed treatment	Barley	1 × 36.3 g a.i./ha	BC	0.0464	0.0	No
						Winter wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.208	0.0	No
								Prairie- SK	0.192	0.0	No
						Spring wheat Peas	1 × 52.5 g a.i./ha	Prairie- MB	0.0776	0.0	No
								Prairie- SK	0.068	0.0	No
							1 × 150 g	BC	0.0008	0.0	No
							a.i./ha	Prairie- MB	0.144	0.0	No
								Prairie- SK	0.104	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
								ON	0.0288	0.0	No
								QC	0.0408	0.0	No
								Atlantic	0.96	0.0	No
						Beans	1 × 50 g	BC	0.00024	0.0	No
							a.i./ha	Prairie- MB	0.0088	0.0	No
								Prairie- SK	0.0064	0.0	No
								ON	0.0456	0.0	No
								QC	0.0712	0.0	No
								Atlantic	0.16	0.0	No
						Canola	1 × 32.3 g	MB	0.16	0.0	No
							a.i./ha	SK	0.128	0.0	No
								ON	0.704	0.0	No
								QC	1.52	0.0	No
						Potato	1 × 117.12 g a.i./ha	Prairie- MB	0.00032	0.0	No
								Atlantic	0.0104	0.0	No
						Corn ⁴	1 × 118.3 g	ON	0.0256	0.0	No
							a.i./ha	QC	0.0248	0.0	No
						Corn ⁵	1 × 118.3 g	ON	0.264	0.0	No
							a.i./ha	QC	0.296	0.0	No
						Sweet	weet 1 × 7.6 g a.i./ha	BC	0.0008	0.0	No
						com ⁴		MB	0.0064	0.0	No
								ON	0.0016	0.0	No
								QC	0.0016	0.0	No
								Atlantic	0.0176	0.0	No
						Sweet	1 × 7.6 g	BC	0.0008	0.0	No
						com ⁵	a.i./ha	MB	0.04	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
								ON	0.0168	0.0	No
								QC	0.0184	0.0	No
								Atlantic	0.0704	0.0	No
						Soybean ⁴	1 × 64 g a.i./ha	Prairie- MB	0.088	0.0	No
								Prairie- SK	0.0776	0.0	No
								ON	0.0584	0.0	No
								QC	0.0392	0.0	No
								Atlantic	0.144	0.0	No
						Soybean ⁵	1 × 64 g a.i./ha	Prairie- MB	0.192	0.0	No
							COAT ARE	Prairie- SK	0.2	0.0	No
								ON	0.216	0.0	No
								QC	0.16	0.0	No
», «*								Atlantic	0.592	0.0	No
Microcosm or r	Chronic		25 4 NOEC -	0.30	Foliar	A1	2 × 0< 25 ~	BC	0.19	0.6	No
Invertebrates	Cinonic	Emergent insects	35-d NOEC = 0.30 (reductions in	0.30	ronar	Apple	2 × 96.25 g a.i./ha at a 10-d interval	ВС	0.19	0.0	NO
			mayfly abundance and			Potato	2 × 26.25 g a.i./ha at a 7-	Prairie- MB	1.4	4.7	Yes
			emergence)				d interval	Atlantic	1.5	5.0	Yes
						Soybean	3 × 25.38 g a.i./ha at a 7-	Prairie- MB	2	6.7	Yes
							d interval	ON	1.6	5.3	Yes
								QC	1.7	5.7	Yes
						Bell pepper	2 × 70 g a.i./ha at a 7-	Prairie- MB	4	13.3	Yes
							d interval	ON	4.6	15.3	Yes
								QC	4.1	13.7	Yes

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC's (µg a.i./L)	RQ	LOC exceeded
						Blueberry	2 × 70 g a.i./ha at a 7- d interval	Atlantic	7.1	23.7	Yes
					Transplan t water	Bell pepper	1 × 117 g a.i./ha	Prairie- MB	2.5	8.3	Yes
								ON	2.4	8.0	Yes
								QC	2.4	8.0	Yes
					In-	Potato	1 × 140 g	Prairie	2.3	7.7	Yes
					furrow/so		a.i./ha	Atlantic	9	30.0	Yes
					il drench plus	Bell	1 × 150 g	Prairie	3.2	10.7	Yes
					irrigation	pepper	a.i./ha	ON	3.1	10.3	Yes
								QC	3	10.0	Yes
					(lowbush	Blueberry (lowbush)	1 × 140 g a.i./ha	Atlantic	9	30.0	Yes
					Seed treatment	Barley	1 × 36.3 g a.i./ha	BC	0.112	0.4	No
						Winter wheat	inter $1 \times 52.5 \text{ g}$	Prairie- MB	0.36	1.2	Yes
								Prairie- SK	0.44	1.5	Yes
						Spring wheat	1 × 52.5 g a.i./ha	Prairie- MB	0.192	0.6	No
								Prairie- SK	0.168	0.6	No
						Peas	1 × 150 g	BC	0.0024	0.0	No
							a.i./ha	Prairie- MB	0.336	1.1	Yes
								Prairie- SK	0.256	0.9	No
								ON	0.088	0.3	No
								QC	0.096	0.3	No
								Atlantic	2.4	8.0	Yes
						Beans	1 × 50 g	BC	0.0008	0.0	No

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i./L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (µg a.i./L)	RQ	LOC exceeded
							a.i./ha	Prairie- MB	0.02	0.1	No
								Prairie- SK	0.0152	0.1	No
								ON	0.128	0.4	No
								QC	0.176	0.6	No
								Atlantic	0.408	1.4	Yes
						Canola	1 × 32.3 g	MB	0.464	1.5	Yes
							a.i./ha	SK	0.32	1.1	Yes
								ON	2.08	6.9	Yes
								QC	3.36	11.2	Yes
						Potato	1 × 117.12 g a.i./ha 1 × 118.3 g	Prairie- MB	0.0008	0.0	No
								Atlantic	0.0264	0.1	No
						Corn ⁴		ON	0.0688	0.2	No
							a.i./ha	QC	0.06	0.2	No
						Corn ⁵	1 × 118.3 g	ON	0.736	2.5	Yes
							a.i./ha	QC	0.688	2.3	Yes
						Sweet	1 × 7.6 g	BC	0.0008	0.0	No
						com ⁴	a.i./ha	MB	0.016	0.1	No
								ON	0.0048	0.0	No
								QC	0.004	0.0	No
								Atlantic	0.048	0.2	No
						Sweet		BC	0.0024	0.0	No
						com ⁵		MB	0.096	0.3	No
								ON	0.0472	0.2	No
								QC	0.044	0.1	No
								Atlantic	0.192	0.6	No
						Soybean ⁴	1 × 64 g	Prairie-	0.312	1.0	Yes

Organism	Exposure	Species	Endpoint reported (µg a.i./L)	Endpoint for RA ¹ (µg a.i/L)	Use Scenario	Crop	Use rate ²	Region	EEC³ (μg a.i./L)	RQ	LOC exceeded
							a.i./ha	MB			
								Prairie- SK	0.232	0.8	No
								ON	0.176	0.6	No
								QC	0.136	0.5	No
								Atlantic	0.4	1.3	Yes
						Soybean ⁵	1 × 64 g a.i./ha	Prairie- MB	0.68	2.3	Yes
								Prairie- SK	0.576	1.9	Yes
								ON	0.656	2.2	Yes
								QC	0.56	1.9	Yes
								Atlantic	1.6	5.3	Yes

¹ The HCs is the 5th percentile of the species sensitivity distribution for the LCs0 or ECs0 at 50% confidence intervals (acute exposures) or NOEC or EC10/EC20 (chronic exposures).

² Use rate represents the maximum number of applications and rate (g a.i./ha) for a crop.

³ EECs based on an 80 cm water depth. For comparison against acute invertebrate endpoints based on data with 24 – 96-h and 7-d sub-chronic studies, peak EECs were used to derive RQs. For comparison against chronic invertebrate endpoints based on data with 21 - 40-d NOEC or EC10/EC20 endpoints, 21-day EECs were used to derive RQs. For comparison against chronic invertebrate endpoints based on pore water exposures, 21-day pore water EECs were used to derive RQs.

⁴ Use on corn, sweet corn or soybeans modelled using the "at depth" scenario.
⁵ Use on corn, sweet corn or soybeans modelled using the "increasing with depth" scenario. Bolded values indicates an exceedence of the level of concern (RQ = 1).

Appendix IV Species Sensitivity Distribution (SSD)

Background information

The median HC₅ and confidence values were reported for the species sensitivity distributions (SSDs). The hazardous concentration to 5% of species (HC₅) is theoretically protective of 95% of all species at the effect level used in the analysis (e.g., LC₅₀, NOEC, etc). An SSD is conducted for taxonomic groups of interest where sufficient data are available. The software program ETX 2.1 is used to generate SSDs which was developed by RIVM (Rijksinstituut voor Volksgezondheid en Milieu, The Netherlands).

SSD Toxicity Data Analysis for thiamethoxam

Data submitted by the registrant and published literature studies were consulted in the risk assessment process. Only those studies with acceptable quantitative effects endpoints were considered for the SSDs. Additional sorting was done to separate data into taxonomic sub groups while also accounting for appropriate test methods, exposure durations, matrices and other variables. Studies from the published literature were deemed acceptable if they reported the appropriate biologically relevant endpoints and generally followed recognized methods such as the Organisation for Economic Co-operation and Development (OECD) or similar.

Results of SSD analysis for thiamethoxam insecticide: Distributions were determined for the taxonomic groups below. Results are reported in summary Table A.4-1 to Table A.4-3:

• Aquatic species: Freshwater invertebrates. Acute and chronic data sets.

The acute HC_5 is 8.96 μg a.i./L, and the chronic HC_5 is 0.026 μg a.i./L. Based on the available data, the results indicate that the HC_5 for chronic effects is approximately 2.5 orders of magnitude more sensitive than the HC_5 for acute effects for freshwater invertebrate populations.

Table A.4-1 Summary of Species Sensitivity Distribution (SSDs) toxicity data analysis for thiamethoxam insecticide.

Study	SSD results
type/Exposure	Freshwater invertebrates
	HC ₅ : 8.96 μg a.i./L
	CI: 3.3-19
Acute toxicity	FA: 1.9-9.2%
	Number of species used: 37 (24 – 96-h, 7-d subchronic EC ₅₀ /LC ₅₀ s)
	Most sensitive species: Neocloeon triangulifer; 96-h $EC_{50} = 5.5 \mu g \text{ a.i./L}$
	HC ₅ : 0.026 μg a.i./L
	CI: 3×10 ⁻⁵ -0.63
Chronic toxicity	FA: 0.34-25%
	Number of species used: 7 (NOEC/EC _{10/20} s)
	Most sensitive species: Cloeon dipterum; 28-d $EC_{10} = 0.43 \mu g \text{ a.i./L}$

 HC_5 = Hazardous concentration to 5% of species.

CI = lower and upper 90% confidence level of HC₅

FA = fraction of species affected. This value reflects the lower and upper 90% confidence level of the proportion of species expected to be affected at the HC_5 value.

Table A.4-2 Toxicity data used in the Species Sensitivity Distribution (SSD) for acute effects of thiamethoxam on freshwater invertebrates.

Species count	Species name	EC ₅₀ /LC ₅₀ (μg ai/L)
1	Lumbriculus sp.	7700.0
2	Chaoborus crystallinus	7300.0
3	Snail (Planorbella pilsbryi)	6195.0
4	Chaoborus sp.	5500.0
5	Caecidotea sp.	4775.4
6	Gammarus kischineffensis	3751.0
7	Gammarus sp.	2800.0
8	Lumbriculus variegatus	2035.1
9	Procambarus clarkii¹	1491.3
10	Coenagrionidae	980.0
11	Daphnia magna ¹	826.1
12	Wavy-rayed lampmussel (Lampsilis fasciola)	691.0
13	Crangonyx pseudogracilis¹	651.3
14	Isonychia bicolor	445.0
15	Agnetina, Paragnetina sp.	445.0
16	Culex quinqefasciatus	343.0
17	Hyalella azteca ¹	289.9
18	Aedes aegypti ¹	233.5
19	Ostracoda (Cyprididae sp.)	180.0
20	Hexagenia sp.	150.2
21	Stenelmis sp.	148.0
22	Cheumatopsyche sp.	118.5
23	Asellus aquaticus	84.0
24	McCaffertium sp.	81.7
25	Aedes sp.	61.9
26	Ephemerella sp.	59.0
27	Anopheles stephensi (from Nadiad) ¹	57.7
28	Trichocorixa sp.	56.3
29	Chironomus riparius ¹	55.5
30	Dytiscidae	47.0
31	Chironomus dilutus ¹	45.0
32	Cloeon sp. 1	24.8
33	Cloeon dipterum ¹	24.3
34	Caenis sp.	23.3
35	Micrasema sp.	18.5
36	Gyrinus sp.	14.0
37	Neocloeon triangulifer	5.5

¹ Toxicity value based on geometric mean

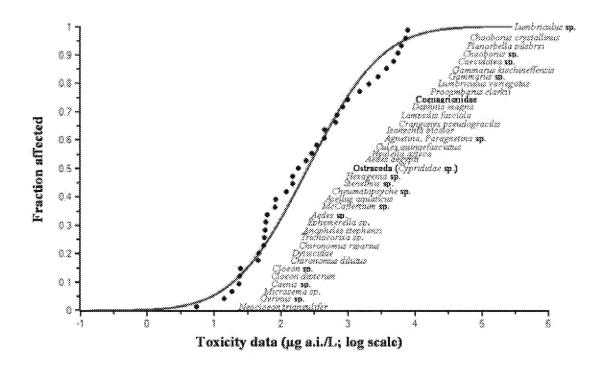


Figure A.4-1 Species Sensitivity Distribution (SSD) for acute toxicity of thiamethoxam to freshwater aquatic invertebrates.

Table A.4-3 Toxicity data used in the Species Sensitivity Distribution (SSD) for chronic effects of thiamethoxam on freshwater invertebrates.

Species count	Species name	NOEC/EC _{10/20} (µg ai/L)
1	Daphnia magna	50000
2	Hyalella azteca	62.5
3	Chaoborus sp.	60
4	Snail (Planorbella pilsbryi)	21.3
5	Chironomus riparius (larvae) ¹	5.7
6	Chironomus dilutus	0.48
7	Cloeon dipterum	0.43

¹ Toxicity value based on geometric mean

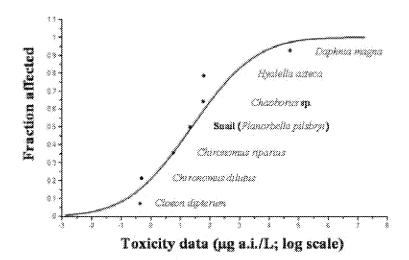


Figure A.4-2 Species Sensitivity Distribution (SSD) for chronic toxicity of thiamethoxam to freshwater aquatic invertebrates.

Comments on data handling for SSDs

Data sorting for use in the SSDs:

- The measurement endpoints used within data subsets are similar (exposure units, toxicity units) and appropriate to the duration category.
- The endpoints included in all data sets are those assumed to ultimately affect survival of the test organisms or populations.
- All short term exposure data are grouped together as "acute" (i.e., 24 hours, 48 hours, 96 hours, etc.) for individual taxonomic groups.
- All data which are considered to be "chronic" are grouped together for individual taxonomic groups (i.e., studies examining the survival or sub-lethal effects from long exposure periods).
- Geometric means of toxicity values are calculated for multiple endpoints for the same species.
- Where more than one measurement endpoint was available for a given study (e.g., both an EC₅₀ and an LC₅₀ are provided, or endpoints from multiple time periods), the more sensitive endpoint is used and not a geometric mean.
- Study results which are insufficient or not compatible for inclusion in either the acute or chronic distribution groups established for the current assessment were not used. This includes for example incompatible effects levels such as EC₂₅, different or unique exposure matrix studies and units, different exposure time/method, etc.

Additional notes on data handling specific to the current active:

- Toxicity data having no effects at the highest test concentration were excluded (e.g., EC₅₀ > X) if there were other results to represent the species (consistent with EFSA (2013) guidance).
- In cases where only one study was available for a species and the resulting endpoint was unbound, i.e., a greater than or less than (</>) toxicity value, the endpoint was used to represent that species (consistent with EFSA (2013) guidance).
- Where both LC₅₀ and EC₅₀ values were available, the more sensitive value was used.
- For chronic effects, NOECs and EC₁₀/EC₂₀ values were considered from studies with a water phase exposure.

Appendix V Estimated Environmental Concentrations from Spray Drift

Table A.5-1 Summary of highest cumulative thiamethoxam use rates according to application method

Ground and ac	rial Use Data						
Crop	Formulation Type	Min single application rate (g ai/ha)	Max single application rate (g ai/ha)	Number of applications	Application interval (days)	Max seasonal rate (g ai/ha)	Max cumulative seasonal rate (g ai/ha) ¹
Ground boom f	oliar spray						
Outdoor ornamentals	Water dispersible granule	75	150	1 at high rate or 2 at low rate	14	150	150
Airblast foliar s	spray		ō		5		
Apple, crab apple	Water dispersible granule	78.75	96.25	2 (1 pre- bloom and 1 post bloom or 2 post bloom applications)	10	192.5	178.1
Aerial applicati						,	~
Bean	Suspension	25.38	25.38	3	7	76.1	68.3
In-furrow dren	<u>ch or irrigation</u>	application		r		r	
Crop group 4 Leafy vegetables, Crop Group 5 Brassica vegetables, Crop group 8 Fruiting vegetables, Crop Group 9: Cucurbit Vegetables	Suspension	90	150	1	NA	150	150
Seed treatment		20	1.50	r		1.50	1 7 = ^
Succulent peas	Suspension	30	150	1	NA	150	150
Sorghum	Suspension	1.6	4.5	1	NA II	4.5	4.5

¹Maximum cumulative seasonal rate = maximum single application rate \times number of applications, adjusted for degredation between applications using the 80^{th} percentile of aerobic aquatic half-lives = 42.8 d and the application interval.

Table A.5-2 Screening level EEC of thiamethoxam and its transformation products in a body of water 80 cm deep after direct application rates of 4.5 g a.i./ha (minimum seed treatment rate), 150 g a.i./ha (maximum seed treatment rate) and 2×96.25 or 178.1 g a.i./ha (maximum cumulative foliar treatment rate)

Compound	Molecular weight	Ratio	4.5 g a.i./ha	150 g a.i./ha	2 × 96.25 g a.i./ha				
Compound	(g/mol)	Ratio	80 cm depth (μg/L)						
CGA 293343 (Thiamethoxam)	291.7	1	0.563	18.8	22.2				
CGA 355190	247.17	0.847	0.477	15.9	18.8				
CGA 322704 (Clothianidin)	249.68	0.856	0.482	16.1	19.0				
CGA 282149	160.03	0.549	0.309	10.3	12.2				
CGA 353042	115.14	0.395	0.222	7.4	8.8				
SYN/NOA 501406	295.29	1.012	0.570	19.0	22.5				
NOA 459602	337.32	1.156	0.651	21.7	25.7				
NOA 407475	246.72	0.846	0.476	15.9	18.8				
NOA 421275*	204.68	0.702	0.395	13.2	15.6				
NOA 404617	236.63	0.811	0.457	15.2	18.0				

^{*:} major transformation product found in both clothianidin and thiamethoxam

Appendix VI Estimated Environmental Concentrations from Water Modelling

1.0 Introduction

The following sections summarize the estimated environmental concentrations (EECs) of thiamethoxam resulting from water modelling for aquatic ecoscenarios.

2.0 Modelling Estimates

2.1 Application Information and Model Inputs

Crops, application rates and timing for various regions were used for modelling ground and aerial foliar applications, ground in-furrow drench or surface band drench plus irrigation, ground transplant water and seed treatments. Regional information on planting and seeding depths for seed treatments was considered. The timing for soybean is assumed from May 1 to June 30 for this use pattern across Canada; the timing for canola treated with clothianidin is used for canola, and timing for sweet corn treated with imidacloprid is used for sweet corn. "Ground transplant water" applications and "ground in-furrow drench or surface band drench plus irrigation" applications were assumed as ground soil applications. The lowest rate for corn of 7.6 g a.i./ha for sweet corn was also modelled. The planting depth for sweet corn in British Columbia is from the Prairie region. All application information is summarized in Table A.6-1.

Table A.6-1 Application rates, timing and other relevant information

Region	Crop	Use pattern	Application method	Seed depth (cm)	Timing
BC	Succulent beans	1×50 g a.i./ha	Seed treatment	4.0	Early April to mid- June
	Succulent peas	1×150 g a.i./ha	Seed treatment	4-5	Early April to mid- June
	Barley	1×36.3 g a.i./ha	Seed treatment	2.0-5.0	April 12 to June 28
	Apple	2×96.25 g a.i./ha at a 10-d interval	Ground foliar	NA	Mid-April to late May
	Sweet corn	1×7.6 g a.i./ha	Seed treatment	3-7.5	May 1 to May 31
Prairie	Succulent beans	1×50 g a.i./ha	Seed treatment	3.8-5	Mid-April to mid- June
	Soybean	1×64 g a.i./ha	Seed treatment	1.90-4.45	Early May –end of June
	Succulent peas	1×150 g a.i./ha	Seed treatment	2.5-5	Mid-April to mid- June
	Spring wheat	1×52.5 g a.i./ha	Seed treatment	2.05-7.5	April 2 to June 21
	Winter wheat	1×52.5 g a.i./ha	Seed treatment	1.5-3.5	August 15 to October 31
	Potato	1×117.12 g a.i./ha	Seed piece treatment	7-15	April 25 to May 31
	Potato	1×140 g a.i./ha	Ground in-furrow drench or surface band drench plus irrigation	NA	April 25 to May 31
	Potato	2×26.25 g a.i./ha at a 7-d interval	Ground and aerial foliar	NA	Early May to early September

Region	Crop	Use pattern	Application method	Seed depth	Timing		
				(cm)			
	Soybean	3×25.38 g a.i./ha at a 7-d interval	Ground and aerial foliar	NA	Early July to Mid- September		
	Bell pepper	1×117 g a.i./ha	Ground transplant water	NA	Early June to early July		
	Bell pepper	2×70 g a.i./ha at a 7-d interval	Ground foliar	NA	Early June to September 22		
	Bell pepper	1×150 g a.i./ha	Ground in-furrow drench or surface band drench plus irrigation	NA	Early June to early July		
	Sweet corn	1×7.6 g a.i./ha	Seed treatment	3-7.5	April 20 to May 31		
	Canola	1×32.3 g a.i./ha	Seed treatment	1.2-5.0	April 17 to June 28		
ON/QC	Bell pepper	1×117 g a.i./ha	Ground transplant water	NA	May 10 to June 15		
	Bell pepper	2×70 g a.i./ha at a 7-d interval	Ground foliar	NA	Early June to late September		
	Bell pepper	1×150 g a.i./ha	Ground in-furrow drench or surface band drench plus irrigation	NA	May 10 to June 15		
	Succulent beans	1×50 g a.i./ha	Seed treatment	2.54-5.08	Early April to end of June		
	Soybean	1×64 g a.i./ha	Seed treatment	2.5-6.4	Early May –end of June		
	Succulent peas	1×150 g a.i./ha	Seed treatment	3.8-7.6	Early April to end of June		
	Com	1×118.3 g a.i./ha	Seed treatment	3.8-6.5	April 14 to June 30		
	Soybeans	3×25.38 g a.i./ha at a 7-d interval	Ground and aerial foliar	NA	Late June to early September		
	Sweet corn	1×7.6 g a.i./ha	Seed treatment	3.8-6.5	April 14 to June 15		
	Canola	1×32.3 g a.i./ha	Seed treatment	0-3	April 1 to June 10		
Atlantic	Succulent beans	1×50 g a.i./ha	Seed treatment	2.5-5	Mid-April to early June		
	Soybean	1×64 g a.i./ha	Seed treatment	2.5-4.0	Early May –end of June		
	Succulent peas	1×150 g a.i./ha	Seed treatment	2-2.5	Mid-April to early June		
	Potato	1×117.12 g a.i./ha	Seed piece treatment	5-15	April 20 to June 15		
	Potato	1×140 g a.i./ha	Ground in-furrow drench or surface band drench plus irrigation	NA	April 20 to June 15		
	Potato	2×26.25 g a.i./ha at a 7	Ground and aerial foliar	NA	Late June to Mid- September		
	Blueberry	2×70 g a.i./ha at a 7-d interval	Ground foliar	NA	Early May to late September		
	Blueberry (lowbush)	1×140 g a.i./ha	Ground soil drench	NA	Early March to October 25		
	Sweet corn	1×7.6 g a.i./ha	Seed treatment	2.5-6	May 1 to June 15		

The main environmental fate parameters used in the models are summarized in Table A.6-2.

Table A.6-2 Major groundwater and surface water model inputs for the ecoscenario assessment of thiamethoxam

Parameter	Value	Comment
Molecular weight (g/mol)	291.7	
Vapour pressure (mm Hg) at 25°C	4.95E-11	
Solubility (mg/L) in water	4100	
Henry's law constant (unitless)	7.77E-14	
Photolysis half-life at 36.1° latitude (day)	4.58	Novartis, North Carolina
Hydrolysis at pH 7	939	Longer of 2 values
K_{∞} (L/kg)	31.14	20 th centile of 6 values
Soil half-life 20°C (day)	402	90 th centile confidence on the mean of
		11 values
Aerobic aquatic half-life 20°C (day)	42.8	80 th centile of 6 values
Anaerobic aquatic half-life 20°C (day)	34.0	80 th centile of 3 values
Application efficiency	0.99, 1.0	ground foliar, seed treatment
Diffusion coefficient in air (cm²/day)	3960	
Heat of Henry (J/mole)	59000	default in PWC

2.2 Aquatic Ecoscenario Assessment

The estimated environmental concentrations (EECs) of thiamethoxam from runoff into a receiving waterbody were simulated using the Pesticide in Water Calculator model (PWC version 1.52) model. The PWC model simulates pesticide runoff from a treated field into an adjacent body of water and the fate of a pesticide within it. Spray drift is not considered for this modelling. The waterbody used in the modelling is a 1-ha wetland with an average depth of 0.8 m and a drainage area of 10 ha. The pore water EECs in a 0.8 m wetland were also generated.

Various initial application dates were modelled (9 to 31 depending on the use patterns and application windows) with eight standard scenarios to cover all use patterns listed in Table A.6-1. For seed treatments where a range of seeding depths were available, the shallowest was selected for modelling. The models were run for 50 years for all scenarios.

For each year of the simulation, PWC calculates peak (or daily maximum) and time-averaged concentrations calculated by averaging the daily concentrations over five time periods (96-hour, 21-day, 60-day and 90-day). The 90th percentiles over each averaging period are reported as the EECs for that period.

The EECs were generated for all selected crops using runoff extraction parameters recommended in Young and Fry (2017). These parameters include a runoff interaction fraction of 0.19, a maximum runoff interaction depth of 8 cm and an exponential decline coefficient of 1.4 cm⁻¹.

Specifically for seed treatments, PWC allows for different modelling approaches to determine pesticide concentrations in water. For the current modelling, two of these scenarios were selected: "at depth" and "increasing with depth". The "at depth" scenario assumes that, at the time of application, the pesticide is present in soil only at the depth the seed is planted. This scenario was used for all the seed treatments selected for modelling. The "increasing with depth" scenario assumes that the pesticide concentration in soil at the time of application linearly increases with depth from the soil surface to the seeding depth. This scenario was used for corn, sweet corn and soybeans, as these are larger seeds which are typically sown using pneumatic

equipment. With this type of seeding method, as the seed penetrates the soil, there is deposition of seeding dust close to the surface and up to the final depth of the seed.

Modelled EECs are presented in Table A.6-3.

The EECs shown in Table A.6-4 were generated for corn, sweet corn and soybean seed treatments using new recommended runoff extraction parameters, as described above, and the application method of Δ (pesticide distributed linearly increase with soil depth from soil surface to a given depth for seed treatment applications).

Table A.6-3 Modelled EECs (µg a.i./L) for thiamethoxam in a waterbody 0.8 m deep, excluding spray drift

Crop	Rate	Region	EEC (ug a.i./L)) in over	lying wate	r	EEC (μg a.i./L) in pore water		
			Peak	96-	21-	60-day	90-	Peak	21-day	
				hour	day		day			
Foliar uses										
Apple	2 × 96.25 g a.i./ha at a 10-d interval	BC	0.23	0.22	0.19	0.14	0.11	0.056	0.055	
Potato	$2 \times 26.25 \text{ g}$	Prairie-MB	1.8	1.7	1.4	1.2	0.97	0.63	0.64	
	a.i./ha at a 7-d interval	Atlantic	1.7	1.7	1.5	1.2	1.2	0.66	0.67	
Soybean	3 × 25.38 g	Prairie-MB	2.4	2.3	2	1.6	1.6	0.96	0.96	
	a.i./ha at a 7-d	ON	2	1.9	1.6	1.3	1.2	0.65	0.65	
	interval	QC	2	1.9	1.7	1.4	1.2	0.7	0.7	
Bell pepper	2 × 70 g a.i./ha	Prairie-MB	4.8	4.6	4	3.1	2.6	1.7	1.7	
	at a 7-d interval	ON	4.9	4.7	4.6	4.1	3.8	2.1	2.1	
		QC	4.6	4.5	4.1	4.2	3.7	2.1	2.1	
Blueberry	2 × 70 g a.i./ha at a 7-d interval	Atlantic	8	7.8	7.1	5.8	4.8	2.8	2.7	
Transplant wa			1							
Bell pepper	1 × 117 g a.i./ha	Prairie-MB	2.9	2.8	2.5	1.9	1.5	0.79	0.78	
1 11		ON	2.8	2.7	2.4	1.9	1.6	0.86	0.85	
		QC	2.9	2.8	2.4	1.7	1.3	0.65	0.64	
In-furrow/soil	drench plus irriga	tion uses	-1		-1					
Potato	1 × 140 g a.i./ha	Prairie	2.6	2.6	2.3	1.8	1.5	0.83	0.82	
		Atlantic	10	9.8	9	7.3	6.2	3.6	3.6	
Bell pepper	1 × 150 g a.i./ha	Prairie	3.7	3.6	3.2	2.4	2	1	1	
1 11		ON	3.5	3.4	3.1	2.4	2	1.1	1.1	
		QC	3.7	3.5	3	2.1	1.7	0.83	0.82	
Blueberry (lowbush)	1 × 140 g a.i./ha	Atlantic	10	9.8	9	7.4	6.7	4.2	4.2	
	t uses modelled us	ing "at depth'	scenar	io						
Barley	1 × 36.3 g	BC	0.16	0.16	0.14	0.12	0.099	0.059	0.058	
	a.i./ha					<u> </u>				
Winter wheat	1 × 52.5 g	Prairie-MB	0.52	0.5	0.45	0.39	0.38	0.26	0.26	
~	a.i./ha	Prairie-SK	0.59	0.56	0.55	0.47	0.41	0.24	0.24	
Spring wheat	1 × 52.5 g	Prairie-MB	0.26	0.26	0.24	0.2	0.17	0.098	0.097	
	a.i./ha	Prairie-SK	0.23	0.22	0.21	0.17	0.15	0.086	0.085	
Peas	1 × 150 g a.i./ha	BC	0.004	0.003	0.003	0.002	0.002	0.001	0.001	
		Prairie-MB	0.45	0.45	0.42	0.035	0.3	0.3	0.18	
		Prairie-SK	0.38	0.36	0.32	0.27	0.23	0.13	0.13	

Crop	Rate	Region	EEC (ug a.i./L)	r	EEC (μg a.i./L) in pore water			
			Peak	96- hour	21- day	60-day	90- day	Peak	21-day
		ON	0.12	0.12	0.11	0.084	0.069	0.037	0.036
		QC	0.13	0.13	0.12	0.1	0.088	0.051	0.051
		Atlantic	3.3	3.3	3	2.4	2.1	1.2	1.2
Beans	1 × 50 g a.i./ha	BC	0.001	0.001	0.001	0.001	0.001	0.0003	0.0003
		Prairie-MB	0.027	0.027	0.025	0.021	0.018	0.011	0.011
		Prairie-SK	0.022	0.021	0.019	0.016	0.013	0.008	0.008
		ON	0.18	0.17	0.16	0.13	0.11	0.058	0.057
		QC	0.23	0.23	0.22	0.18	0.15	0.089	0.089
		Atlantic	0.56	0.55	0.51	0.42	0.35	0.2	0.2
Potato	1 × 117.12 g	Prairie-MB	0.001	0.001	0.001	0.001	0.001	0.0004	0.0004
	a.i./ha	Atlantic	0.038	0.037	0.033	0.027	0.023	0.013	0.013
Corn	1 × 118.3 g	ON	0.097	0.095	0.086	0.07	0.058	0.032	0.032
	a.i./ha	QC	0.079	0.078	0.075	0.064	0.054	0.031	0.031
Sweet corn	1 × 7.6 g a.i./ha	BC	< 0.00	< 0.00	< 0.00	< 0.001	< 0.00	< 0.001	< 0.001
			1	1	1	0.01	1		
		Prairie-MB	0.021	0.021	0.020	0.017	0.014	0.008	0.008
		ON	0.006	0.006	0.006	0.005	0.004	0.002	0.002
		QC	0.005	0.005	0.005	0.004	0.003	0.002	0.002
		Atlantic	0.066	0.065	0.060	0.048	0.040	0.022	0.022
Soybean	1 × 64 g a.i./ha	Prairie-MB	0.45	0.43	0.39	0.28	0.23	0.11	0.11
		Prairie-SK	0.35	0.33	0.29	0.23	0.18	0.098	0.097
		ON	0.25	0.25	0.22	0.17	0.14	0.075	0.073
		QC	0.21	0.2	0.17	0.12	0.097	0.05	0.049
~ .		Atlantic	0.55	0.55	0.5	0.4	0.33	0.19	0.18
Canola	1 × 32.3 g	Prairie-MB	0.64	0.63	0.58	0.47	0.38	0.21	0.20
	a.i./ha	Prairie-SK	0.46	0.45	0.40	0.33	0.28	0.16	0.16
		ON	3.0	2.9	2.6	2.0	1.7	0.89	0.88
		QC	4.5	4.4	4.2	3.5	2.9	2.0	1.9
	nt uses modelled us				T	T . ==		T	
Corn	$1 \times 118.3 \text{ g}$	ON	0.99	0.96	0.92	0.75	0.61	0.34	0.33
	a.i./ha	QC	0.93	0.91	0.86	0.76	0.65	0.37	0.37
Sweet corn	1×7.6 g a.i./ha	BC	0.004	0.003	0.003	0.002	0.002	0.001	0.001
		Prairie-MB	0.13	0.13	0.12	0.10	0.086	0.050	0.050
		ON	0.064	0.062	0.059	0.048	0.039	0.022	0.021
		QC	0.059	0.058	0.055	0.049	0.042	0.024	0.023
	<u> </u>	Atlantic	0.26	0.26	0.24	0.19	0.16	0.089	0.088
Soybean	1 × 64 g a.i./ha	Prairie-MB	0.98	0.94	0.85	0.62	0.5	0.24	0.24
		Prairie-SK	0.85	0.82	0.72	0.57	0.47	0.25	0.25
		ON	0.92	0.9	0.82	0.63	0.51	0.27	0.27
		QC	0.85	0.82	0.7	0.49	0.39	0.2	0.2
		Atlantic	2.2	2.2	2	1.6	1.3	0.75	0.74

Appendix VII Summary of Water Monitoring Analysis

Table A.7-1 Summary statistics for thiamethoxam measured in waterbodies from Prince Edward Island, Nova Scotia and New Brunswick.

- -In calculations, the PMRA assigned a value equal to half the limit of detection to samples that showed no detection.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred once or twice per month between May and October. Sampling at some sites occurred only a few times over a short time period, and values measured may not represent concentrations throughout the growing season.

Waterbody	Land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of sam)	oles) exceeding	g the toxicity o	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
						Prin	ce Edward I	sland						7 7
Clyde River (PMRA# 2745506, 2468268)	Pasture, forest, potatoes, soybeans, other crops	2012 2015	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Clyde River (PMRA# 2845169)	Pasture, forest, potatoes, soybeans, other crops	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Dunk River (PMRA# 2745506, 2468268)	Pasture, forest, potatoes, other crops	2010 2013	0.01	4	0	0	0.005 0.005	0	0.005 0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Dunk River (PMRA# 2845169)	Pasture, forest, potatoes, other crops	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Huntley River (PMRA# 2745506, 2468268)	Pasture, potatoes, soybeans, other crops	2012 2015	0.01	4	3 4	75 100	0.016 0.023	0.008	0.005 0.005	0.02	0 (0%) 2 (50%)	0 (0%)	0 (0%)	0 (0%)
Huntley River (PMRA# 2845169)	Pasture, potatoes, soybeans, other crops	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mill River (PMRA# 2745506, 2468268)	Forest, potatoes, soybeans, other crops	2011 2014	0.01 0.01	4	0 1	0 25	0.005 0.006	0.003	0.005 0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mill River (PMRA# 2845169)	Forest, potatoes, soybeans, other crops	2017	0.01	5	2	40	0.007	0.003	0.005	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Montague River	Forest, potatoes,	2011	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of sami	oles) exceeding	the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
(PMRA# 2745506, 2468268)	soybeans, wheat, other crops	2014	0.01	4	1	25	0.006	0.003	0.005	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Montague River (PMRA# 2845169)	Forest, potatoes, soybeans, wheat, other crops	2017	0.01	5	1	20	0.006	0.002	0.005	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Morell River	Mainly not	2010	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2745506, 2468268)	cultivated (forest, shrubland, pasture)	2013	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Morell River (PMRA# 2845169)	Mainly not cultivated (forest, shrubland, pasture)	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
West River	Mainly not	2010	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2745506, 2468268)	cultivated (forest, shrubland, pasture)	2013	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
West River (PMRA# 2845169)	Mainly not cultivated (forest, shrubland, pasture)	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Wilmot River	Potatoes,	2012	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2745506, 2468268)	soybeans, other crops, pasture	2015	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Wilmot River	Potatoes,	2015	0.00139	6	5	67	0.003	0.002	0.003	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2834289)	soybeans, other crops, pasture	2016	0.00139	3	3	100	0.004	0.0002	0.004	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Wilmot River (PMRA# 2845169)	Potatoes, soybeans, other crops, pasture	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Winter River	Corn, soybeans,	2011	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2745506, 2468268)	cereals, fruit, vegetables	2014	0.01	4	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Winter River (PMRA# 2845169)	Potatoes, barley, wheat, corn	2017	0.01	5	0	0	0.005	0	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
						Ŋ	ew Brunswi	ck						
Big Presqu'île CMP station (PMRA# 2834289)	Potatoes, corn, other crops	2015	0.00139	7	7	100	0.003	0.0006	0.003	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
							Nova Scotia	1						
Cornwallis River	Urban, potatoes,	2015	0.00139	6	6	100	0.006	0.007	0.003	0.021	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2834289)	corn, other crops	2016	0.00139	1	1	100	0.001	NA	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of samples) exceeding the toxicity endpoints			
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
Coleman Brook (PMRA# 2834289)	Forest, shrubland, wheat, corn, other crops	2016	0.00139	1	1	100	0.008	NA	0.008	0.008	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Rand Brook (PMRA# 2834289)	Corn, pasture, other crops, wheat	2016	0.00139	1	1	100	0.002	NA	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Skinner Brook (PMRA# 2834289)	Cranberries, corn, urban, potatoes, other crops	2016	0.00139	1	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Watton Brook (PMRA# 2834289)	Urban, shrubland, pasture and forages	2016	0.00139	1	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)

LOD = limit of detection; N = sample size; Stdev = standard deviation; Chronic HC_5 = the 5^{th} percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC_{10} = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC_5 = the 5^{th} percentile of the species sensitivity distribution for the LC_{50} (the median lethal concentration) at 50% confidence intervals; NA = not applicable

Table A.7-2 Risk quotients for thiamethoxam measured in waterbodies located in Prince Edward Island, Nova Scotia and New Brunswick.

- -Shaded cells indicate the level of concern is exceeded, meaning that the risk quotient is equal to or greater than a value of 1.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred once or twice per month between May and October. Sampling at some sites occurred only a few times over a short time period, and values measured may not represent concentrations throughout the growing season.

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated us	ronic Risk Quotie ing median ² conce onic toxicity endp	ntrations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
					Prince F					
Clyde River (PMRA# 2745506, 2468268)	Pasture, forest, potatoes, soybeans, other crops	2012	4	0.2	< 0.1 < 0.1	< 0.1 < 0.1	0.2	< 0.1 < 0.1	< 0.1 < 0.1	< 0.1
Clyde River (PMRA# 2845169)	Pasture, forest, potatoes, soybeans, other crops	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Dunk River	Pasture, forest,	2010	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	potatoes, other crops	2013	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Dunk River (PMRA# 2845169)	Pasture, forest, potatoes, other crops	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Huntley River	Pasture, potatoes,	2012	4	0.6	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	soybeans, other crops	2015	4	0.9	< 0.1	0.1	0.2	< 0.1	< 0.1	< 0.1
Huntley River (PMRA# 2845169)	Pasture, potatoes, soybeans, other crops	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Mill River	Forest, potatoes,	2011	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	soybeans, other crops	2014	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated us	ronic Risk Quotic ing median ² conc onic toxicity endp	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint	
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Mill River (PMRA# 2845169)	Forest, potatoes, soybeans, other crops	2017	5	0.3	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Montague River (PMRA# 2745506, 2468268)	Forest, potatoes, soybeans, wheat, other crops	2011 2014	4	0.2	< 0.1 < 0.1	< 0.1 < 0.1	0.2	< 0.1 < 0.1	< 0.1 < 0.1	< 0.1 < 0.1
Montague River (PMRA# 2845169)	Forest, potatoes, soybeans, wheat, other crops	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Morell River	Mainly not	2010	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	cultivated (forest, shrubland, pasture)	2013	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Morell River (PMRA# 2845169)	Mainly not cultivated (forest, shrubland, pasture)	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
West River	Mainly not	2010	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	cultivated (forest, shrubland, pasture)	2013	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
West River (PMRA# 2845169)	Mainly not cultivated (forest, shrubland, pasture)	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Wilmot River	Potatoes,	2012	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	soybeans, other crops, pasture	2015	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Wilmot River	Potatoes,	2015	6	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2834289)	soybeans, other crops, pasture	2016	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Wilmot River (PMRA# 2845169)	Potatoes, soybeans, other crops, pasture	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Winter River	Corn, soybeans,	2011	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA# 2745506, 2468268)	cereals, fruit, vegetables	2014	4	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated us	ronic Risk Quotic ing median ² conc onic toxicity endp	entrations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Winter River (PMRA# 2845169)	Potatoes, barley, wheat, corn	2017	5	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
					New	Brunswick				
Big Presqu'île CMP station (PMRA# 2834289)	Potatoes, corn, other crops	2015	7	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
					No	va Scotia				
Cornwallis River (PMRA# 2834289)	Urban, potatoes, corn, other crops	2015 2016	6 1	0.2 0.1	< 0.1 < 0.1	< 0.1 < 0.1	0.1 0.1	< 0.1 < 0.1	< 0.1 < 0.1	< 0.1 < 0.1
Coleman Brook (PMRA# 2834289)	Forest, shrubland, wheat, corn, other crops	2016	1	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
Rand Brook (PMRA# 2834289)	Corn, pasture, other crops, wheat	2016	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Skinner Brook (PMRA# 2834289)	Cranberries, corn, urban, potatoes, other crops	2016	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Watton Brook (PMRA# 2834289)	Urban, shrubland, pasture and forages	2016	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1

N = sample size; Chronic HC₅ = the 5th percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC₁₀ = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC₅ = the 5th percentile of the species sensitivity distribution for the LC₅₀ (the median lethal concentration) at 50% confidence intervals

¹Risk Quotient = concentration ÷ toxicity endpoint

²Average, median and maximum concentrations over the sampling period are reported in Table A.7-1.

³Because monitoring may not capture peak concentrations, maximum concentrations may be underestimated.

Table A.7-3 Summary statistics for thiamethoxam measured in waterbodies from Quebec.

- -In calculations, the PMRA assigned a value equal to half the limit of detection to samples that showed no detection.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred one to three times per week between May and August. Sampling at two sites occurred only once, and values measured at these sites may not represent concentrations throughout the growing season.

Waterbody	Major land use	Year	LOD	N	N	0/0	Average	Stdev	Median	Max	N (% of san	iples) exceed	ing the toxicity	endpoints
(Data source)			(μg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic	Chronic	Mesocosm	Acute
											HC5 of	EC ₁₀ of	NOEC of	HC5 of
											0.026 μg/L	0.43 μg/L	0.3 μg/L	9 μg/L
Chibouet River	Corn, soybeans	2014	0.001	29	29	100	0.037	0.049	0.018	0.24	10 (34%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2561884,		2015	0.001	26	26	100	0.048	0.042	0.029	0.17	16 (62%)	0 (0%)	0 (0%)	0 (0%)
2709791, 2821395)		2016	0.001	30	30	100	0.038	0.025	0.03	0.12	17 (57%)	0 (0%)	0 (0%)	0 (0%)
		2017	0.002	22	22	100	0.02	0.012	0.015	0.057	6 (27%)	0 (0%)	0 (0%)	0 (0%)
Des Hurons River	Corn, soybeans	2014	0.001	30	29	97	0.044	0.067	0.019	0.24	9 (30%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2561884,		2015	0.001	27	27	100	0.044	0.04	0.025	0.15	13 (48%)	0 (0%)	0 (0%)	0 (0%)
2709791, 2821395)		2016	0.001	29	28	97	0.038	0.046	0.013	0.17	11 (38%)	0 (0%)	0 (0%)	0 (0%)
		2017	0.002	23	23	100	0.023	0.022	0.014	0.1	8 (35%)	0 (0%)	0 (0%)	0 (0%)
Saint-Régis River	Corn, soybeans	2014	0.001	29	29	100	0.16	0.15	0.11	0.59	24 (83%)	2 (7%)	5 (17%)	0 (0%)
(PMRA# 2561884,		2015	0.001	27	27	100	0.4	0.92	0.12	4.5	25 (93%)	3 (11%)	5 (19%)	0 (0%)
2709791, 2821395)		2016	0.001	30	28	93	0.18	0.17	0.12	0.69	26 (87%)	3 (10%)	5 (17%)	0 (0%)
		2017	0.002	24	24	100	0.11	0.16	0.065	0.74	23 (96%)	2 (8%)	2 (8%)	0 (0%)
Saint-Zéphirin	Corn, soybeans	2014	0.001	29	28	97	0.034	0.053	0.02	0.27	11 (38%)	0 (0%)	0 (0%)	0 (0%)
River		2015	0.001	27	27	100	0.057	0.079	0.026	0.31	13 (48%)	0 (0%)	1 (4%)	0 (0%)
(PMRA# 2561884,		2016	0.001	30	29	97	0.04	0.034	0.03	0.13	16 (53%)	0 (0%)	0 (0%)	0 (0%)
2709791, 2821395)		2017	0.002	23	23	100	0.045	0.051	0.028	0.21	12 (52%)	0 (0%)	0 (0%)	0 (0%)
Blanche River	Potatoes, corn,	2012	0.002	24	24	100	0.059	0.015	0.059	0.089	24 (100%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2544468,	cereals	2017	0.002	30	30	100	0.075	0.019	0.078	0.1	29 (97%)	0 (0%)	0 (0%)	0 (0%)
2821395)														
Chartier Creek	Potatoes, corn,	2010	0.001	27	27	100	0.2	0.2	0.12	0.9	27 (100%)	2 (7%)	7 (26%)	0 (0%)
(PMRA# 2523837,	cereals	2012	0.002	28	28	100	0.41	0.42	0.17	1.5	28 (100%)	11 (39%)	13 (46%)	0 (0%)
2544468, 2821395)		2017	0.002	30	30	100	0.28	0.15	0.26	0.6	30 (100%)	6 (20%)	12 (40%)	0 (0%)
Point-du-Jour	Potatoes, corn,	2010	0.001	27	27	100	0.033	0.009	0.033	0.056	18 (67%)	0 (0%)	0 (0%)	0 (0%)
Creek	soybeans, cereals	2012	0.002	28	28	100	0.12	0.061	0.11	0.33	28 (100%)	0 (0%)	1 (4%)	0 (0%)
(PMRA# 2523837,		2017	0.002	29	29	100	0.17	0.05	0.16	0.33	29 (100%)	0 (0%)	1 (3%)	0 (0%)
2544468, 2821395)														
Déversant-du-Lac	Orchards, corn,	2010	0.001	30	22	73	0.01	0.031	0.005	0.17	1 (3%)	0 (0%)	0 (0%)	0 (0%)
Creek	soybeans, cereals	2011	0.001	31	27	87	0.016	0.037	0.004	0.17	3 (10%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523837,		2015	0.001	28	14	50	0.006	0.009	0.001	0.042	1 (4%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	nples) exceed	ing the toxicity	endpoints
(Data source)	•		(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
2544468, 2821394, 2821395)		2016	0.001	30	12	40	0.005	0.012	0.0005	0.053	2 (7%)	1 (3%)	0 (0%)	0 (0%)
Rousse Creek	Orchards, corn,	2010	0.001	29	2	7	0.001	0.001	0.0005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523837,	soybeans,	2011	0.001	27	5	19	0.005	0.015	0.0005	0.07	2 (7%)	0 (0%)	0 (0%)	0 (0%)
2544468, 2821394,	vegetables	2015	0.001	29	19	66	0.025	0.084	0.007	0.46	4 (14%)	1 (3%)	1 (3%)	0 (0%)
2821395)		2016	0.001	30	24	80	0.021	0.022	0.015	0.077	7 (23%)	0 (0%)	0 (0%)	0 (0%)
Gibeault-Delisle Creek (PMRA# 2709793, 2821394)	Vegetables, potatoes, corn, soybeans	2013	0.001	30	30	100	0.27 0.066	0.82	0.034	0.46	16 (57%) 16 (53%)	3 (11%) 1 (3%)	3 (11%) 1 (3%)	0 (0%)
Norton Creek	Vegetables,	2013	0.001	27	26	96	0.015	0.02	0.006	0.074	3 (11%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2709793, 2821394)	potatoes, corn, soybeans	2014	0.001	30	29	97	0.031	0.026	0.028	0.088	16 (53%)	0 (0%)	0 (0%)	0 (0%)
Yamaska River	Mixed	2014	0.001	10	10	100	0.039	0.049	0.017	0.16	4 (40%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2561884,	cropsMixed	2016	0.001	9	9	100	0.053	0.11	0.013	0.33	2 (22%)	0 (0%)	1 (11%)	0 (0%)
2821395)	crops, corn, soybeans	2017	0.002	9	9	100	0.016	0.019	0.009	0.063	1 (11%)	0 (0%)	0 (0%)	0 (0%)
À la Barbue River (PMRA# 2561884)	Mixed cropsMixed crops, corn, soybeans	2013	0.001	10	9	90	0.042	0.069	0.017	0.23	4 (40%)	0 (0%)	0 (0%)	0 (0%)
Bécancour River (PMRA# 2561884)	Mixed cropsMixed crops, corn, soybeans	2014	0.001	11	6	55	0.008	0.015	0.001	0.043	2 (18%)	0 (0%)	0 (0%)	0 (0%)
La Chaloupe River (PMRA# 2523837, 2561884)	Mixed cropsMixed crops, corn, soybeans	2012	0.002	10	8	80	0.047	0.082	0.014	0.27	4 (40%)	0 (0%)	0 (0%)	0 (0%)
Châteauguay River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	0.002	11	11	100	0.02	0.026	0.009	0.089	2 (18%)	0 (0%)	0 (0%)	0 (0%)
De l'Achigan River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	0.002	10	9	90	0.018	0.018	0.009	0.059	3 (30%)	0 (0%)	0 (0%)	0 (0%)
L'Assomption River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	0.002	11	8	73	0.011	0.013	0.006	0.045	1 (9%)	0 (0%)	0 (0%)	0 (0%)
Du Loup River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	0.001	10	5	50	0.002	0.002	0.001	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	º/o	Average	Stdev	Median	Max	N (% of san	nples) exceed	ing the toxicity	endpoints
(Data source)	·		(μg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
Gentilly River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	0.001	11	6	55	0.006	0.01	0.002	0.033	1 (9%)	0 (0%)	0 (0%)	0 (0%)
L'Acadie River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	0.001	10	10	100	0.068	0.06	0.036	0.16	7 (70%)	0 (0%)	0 (0%)	0 (0%)
Mascouche River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	0.001	10	10	100	0.024	0.033	0.014	0.11	3 (30%)	0 (0%)	0 (0%)	0 (0%)
Nicolet River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	0.001	11	5	45	0.005	0.013	0.0005	0.043	1 (9%)	0 (0%)	0 (0%)	0 (0%)
Saint-François River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	0.001	11	6	55	0.005	0.009	0.001	0.03	1 (9%)	0 (0%)	0 (0%)	0 (0%)
À l'Ours River (PMRA# 2821395)	Mixed crops	2017	0.002	10	10	100	0.064	0.099	0.023	0.34	3 (30%)	0 (0%)	1 (10%)	0 (0%)
Beaurivage River (PMRA# 2709792)	Mixed crops	2015	0.001	11	11	100	0.017	0.018	0.006	0.049	3 (27%)	0 (0%)	0 (0%)	0 (0%)
Boyer River (PMRA# 2709792, 2821395)	Mixed crops	2016	0.001	11	10	91	0.028	0.029	0.015	0.095	4 (36%)	0 (0%)	0 (0%)	0 (0%)
Chaudière River (2 sites) (PMRA# 2709792	Mixed crops	2015	0.001	11	7	64	0.007	0.015	0.002	0.051	1 (9%)	0 (0%)	0 (0%)	0 (0%)
Du Chêne River (PMRA# 2709792)	Mixed crops	2015	0.001	11	10	91	0.017	0.028	0.005	0.079	2 (18%)	0 (0%)	0 (0%)	0 (0%)
Du Sud River (PMRA# 2709792, 2821395)	Mixed crops	2016	0.001	11	4	36	0.004	0.006	0.0005	0.019	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Etchemin River (PMRA# 2709792)	Mixed crops	2015	0.001	11	10	91	0.015	0.025	0.004	0.076	2 (18%)	0 (0%)	0 (0%)	0 (0%)
Le Bras River (PMRA# 2709792)	Mixed crops	2015	0.001	11	11	100	0.04	0.047	0.021	0.14	5 (45%)	0 (0%)	0 (0%)	0 (0%)
Mistassini River (PMRA# 2821395)	Mixed crops	2017	0.002	11	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mistouk River (PMRA# 2821395)	Mixed crops	2017	0.002	11	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Moreau River (PMRA# 2821395)	Mixed crops	2017	0.002	11	4	36	0.003	0.003	0.001	0.012	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Richelieu River (PMRA# 2709792, 2821395)	Mixed crops	2016	0.001	10	5	50	0.005	0.01	0.001	0.032	1 (10%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceedi	ng the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
Ruisseau puant près du rang Sainte- Anne (PMRA# 2821395)	Mixed crops	2017	0.002	11	1	9	0.001	0.001	0.001	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Ticouapé River (PMRA# 2821395)	Mixed crops	2017	0.002	11	1	9	0.002	0.003	0.001	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Saint-Pierre Lake (3 sites) (PMRA# 2821395)	Corn, soybeans, wheat, potatoes, urban	2017	0.002	33	13	39	0.003	0.003	0.001	0.015	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Ditch (PMRA# 2548877)	Agriculture	2013	0.001 (LOQ)	1	0	0	0.0005	NA	0.0005	0.0005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Stream (PMRA# 2548876)	Agriculture	2014	0.0008	1	1	100	0.007	NA	0.007	0.007	0 (0%)	0 (0%)	0 (0%)	0 (0%)

LOD = limit of detection; N = sample size; Stdev = standard deviation; Chronic HC₅ = the 5th percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC₁₀ = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC₅ = the 5th percentile of the species sensitivity distribution for the LC₅₀ (the median lethal concentration) at 50% confidence intervals; LOQ = limit of quantification; NA = not applicable

Table A.7-4 Risk quotients for thiamethoxam measured in waterbodies located in Quebec.

- -Shaded cells indicate the level of concern is exceeded, meaning that the risk quotient is equal to or greater than a value of 1.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred one to three times per week between May and August. Sampling at two sites occurred only once, and values measured at these sites may not represent concentrations throughout the growing season.

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce tic toxicity endpo	ntrations and	calculated usin	onic Risk Quotien ng median ² concen nic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Chibouet River	Corn, soybeans	2014	29	1.4	0.1	0.1	0.7	< 0.1	0.1	< 0.1
(PMRA#		2015	26	1.8	0.1	0.2	1.1	0.1	0.1	< 0.1
2561884,		2016	30	1.4	0.1	0.1	1.2	0.1	0.1	< 0.1
2709791, 2821395)		2017	22	1.4	0.1	0.1	0.6	< 0.1	0.1	< 0.1

(PMRA# 2561884, 2709791, 2821395) Saint-Régis River (PMRA# 2561884, 2709791, 2821395)	Corn, soybeans Corn, soybeans	2014 2015 2016 2017	30 27 29 23	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC	Chronic HC5	Chronic EC ₁₀	Mesocosm	acute toxicity endpoint Acute HC ₅
(PMRA# 2561884, 2709791, 2821395) Saint-Régis River (PMRA# 2561884, 2709791, 2821395) Saint-Zéphirin C		2015 2016 2017	27 29			(0.3 μg/L)	(0.026 μg/L)	(0.43 μg/L)	NOEC (0.3 μg/L)	(9 μg/L)
(PMRA# 2561884, 2709791, 2821395) Saint-Régis River (PMRA# 2561884, 2709791, 2821395) Saint-Zéphirin C		2015 2016 2017	27 29		0.1	0.1	0.7	< 0.1	0.1	< 0.1
2709791, 2821395) Saint-Régis River (PMRA# 2561884, 2709791, 2821395) Saint-Zéphirin C	Corn, soybeans	2017	29	1.7	0.1	0.1	1	0.1	0.1	< 0.1
2821395) Saint-Régis River (PMRA# 2561884, 2709791, 2821395) Saint-Zéphirin C	Corn, soybeans	2017		1.5	0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
(PMRA# 2561884, 2709791, 2821395) Saint-Zéphirin C	Corn, soybeans	0011		0.9	0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
2561884, 2709791, 2821395) Saint-Zéphirin C	·	2014	29	6	0.4	0.5	4.2	0.3	0.4	0.1
2709791, 2821395) Saint-Zéphirin C		2015	27	15	0.9	1.3	4.6	0.3	0.4	0.5
2821395) Saint-Zéphirin C		2016	30	7	0.4	0.6	4.6	0.3	0.4	0.1
		2017	24	7	0.4	0.6	2.5	0.2	0.2	0.1
Director (DMD A#	Corn, soybeans	2014	29	1.3	0.1	0.1	0.8	< 0.1	0.1	< 0.1
		2015	27	2.2	0.1	0.2	1	0.1	0.1	< 0.1
2561884,		2016	30	1.5	0.1	0.1	1.2	0.1	0.1	< 0.1
2709791, 2821395)		2017	23	1.7	0.1	0.2	1.1	0.1	0.1	< 0.1
Blanche River P	Potatoes, corn,	2012	24	2.3	0.1	0.2	2.3	0.1	0.2	< 0.1
(PMRA# C4 2544468, 2821395)	cereals	2017	30	2.9	0.2	0.2	3	0.2	0.3	< 0.1
Chartier Creek P	Potatoes, corn,	2010	27	7.6	0.5	0.7	4.6	0.3	0.4	0.1
(PMRA# co	cereals	2012	28	16	1	1.4	6.3	0.4	0.6	0.2
2523837, 2544468, 2821395)		2017	30	11	0.7	0.9	10	0.6	0.9	0.1
	Potatoes, corn,	2010	27	1.3	0.1	0.1	1.3	0.1	0.1	< 0.1
	soybeans, cereals	2012	28	4.5	0.3	0.4	4	0.2	0.4	< 0.1
2523837, 2544468, 2821395)		2017	29	6.4	0.4	0.6	6.2	0.4	0.5	< 0.1
	Orchards, corn,	2010	30	0.4	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Creek so	soybeans, cereals	2011	31	0.6	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA#		2015	28	0.2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2523837, 2544468, 2821394, 2821395)		2016	30	0.2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
L		2010	29	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	201
(PMRA# se	Orchards, corn,	2011	27		1			➤ V. I	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated usin	onic Risk Quotien og median ² concen nic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
2523837, 2544468, 2821394, 2821395)	vegetables	2015 2016	30	0.8	0.1 < 0.1	0.1	0.3	< 0.1 < 0.1	< 0.1 < 0.1	< 0.1 < 0.1
Gibeault-Delisle Creek (PMRA# 2709793, 2821394)	Vegetables, potatoes, corn, soybeans	2013 2014	28 30	10 2.6	0.6 0.2	0.9	1.3	0.1 0.1	0.1	0.5 < 0.1
Norton Creek (PMRA# 2709793, 2821394)	Vegetables, potatoes, corn, soybeans	2013 2014	30	0.6	< 0.1	< 0.1 0.1	0.2	< 0.1 0.1	< 0.1	< 0.1 < 0.1
Yamaska River (PMRA# 2561884, 2821395)	Mixed crops, corn, soybeans	2014 2016 2017	10 9 9	1.5 2 0.6	0.1 0.1 < 0.1	0.1 0.2 0.1	0.7 0.5 0.4	< 0.1 < 0.1 < 0.1	0.1 < 0.1 < 0.1	< 0.1 < 0.1 < 0.1
À la Barbue River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	10	1.6	0.1	0.1	0.7	< 0.1	0.1	< 0.1
Bécancour River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	11	0.3	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
La Chaloupe River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	10	1.8	0.1	0.2	0.5	< 0.1	< 0.1	< 0.1
Châteauguay River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	11	0.8	< 0.1	0.1	0.4	< 0.1	< 0.1	< 0.1
De l'Achigan River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	10	0.7	< 0.1	0.1	0.3	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average² conce lic toxicity endpo	ntrations and	calculated usin	onic Risk Quotien ng median² concen nic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
L'Assomption River (PMRA# 2523837, 2561884)	Mixed crops, corn, soybeans	2012	11	0.4	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Du Loup River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	10	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Gentilly River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	11	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
L'Acadie River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	10	2.6	0.2	0.2	1.4	0.1	0.1	< 0.1
Mascouche River (PMRA# 2561884)	Mixed crops, corn, soybeans	2013	10	0.9	0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
Nicolet River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	11	0.2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Saint-François River (PMRA# 2561884)	Mixed crops, corn, soybeans	2014	11	0.2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
À l'Ours River (PMRA# 2821395)	Mixed crops	2017	10	2.5	0.1	0.2	0.9	0.1	0.1	< 0.1
Beaurivage River (PMRA# 2709792)	Mixed crops	2015	11	0.7	< 0.1	0.1	0.2	< 0.1	< 0.1	< 0.1
Boyer River (PMRA# 2709792, 2821395)	Mixed crops	2016	11	1.1	0.1	0.1	0.6	< 0.1	< 0.1	< 0.1
Chaudière River (2 sites) (PMRA# 2709792	Mixed crops	2015	11	0.3	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce tic toxicity endpo	ntrations and	calculated usi	onic Risk Quotien ng median ² concen nic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Du Chêne River (PMRA# 2709792)	Mixed crops	2015	11	0.7	< 0.1	0.1	0.2	< 0.1	< 0.1	< 0.1
Du Sud River (PMRA# 2709792, 2821395)	Mixed crops	2016	11	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Etchemin River (PMRA# 2709792)	Mixed crops	2015	11	0.6	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Le Bras River (PMRA# 2709792)	Mixed crops	2015	11	1.6	0.1	0.1	0.8	< 0.1	0.1	< 0.1
Mistassini River (PMRA# 2821395)	Mixed crops	2017	11	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Mistouk River (PMRA# 2821395)	Mixed crops	2017	11	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Moreau River (PMRA# 2821395)	Mixed crops	2017	11	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Richelieu River (PMRA# 2709792, 2821395)	Mixed crops	2016	10	0.2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Ruisseau puant près du rang Sainte-Anne (PMRA# 2821395)	Mixed crops	2017	11	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Ticouapé River (PMRA# 2821395)	Mixed crops	2017	11	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated usi	onic Risk Quotien ng median ² concen nic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Saint-Pierre Lake (3 sites) (PMRA# 2821395)	Corn, soybeans, wheat, potatoes, urban	2017	33	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Ditch (PMRA# 2548877)	Agriculture	2013	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Stream (PMRA# 2548876)	Agriculture	2014	1	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1

N = sample size, Chronic HC₅ = the 5th percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC₁₀ = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam), NOEC = no observable effect concentration; Acute HC₅ = the 5th percentile of the species sensitivity distribution for the LC₅₀ (the median lethal concentration) at 50% confidence intervals

Table A.7-5 Summary statistics for thiamethoxam measured in waterbodies from Ontario.

- -In calculations, the PMRA assigned a value equal to half the limit of detection to samples that showed no detection.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred one to four times per month between April and November. Sampling at some sites occurred only a few times over a short time period, and values measured may not represent concentrations throughout the growing season.

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)			(μg/L)		detects	Detection	(µg/L)		(µg/L)	(μg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 μg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Two Mile Creek	Vineyards,	2012	0.00139	15	3	20	0.002	0.003	0.0007	0.008	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,	orchards	2013	0.00139	14	1	7	0.001	0.0004	0.0007	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,		2014	0.00139	12	12	100	0.013	0.009	0.01	0.025	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287)		2015	0.00139	13	10	77	0.002	0.003	0.002	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
		2016	0.00139	6	3	50	0.01	0.021	0.002	0.052	1 (17%)	0 (0%)	0 (0%)	0 (0%)
Twenty Mile Creek	Soybeans, corn	2011	0.00139	1	1	100	0.17	NA	0.17	0.17	1 (100%)	0 (0%)	0 (0%)	0 (0%)

¹Risk Quotient = concentration ÷ toxicity endpoint

²Average, median and maximum concentrations over the sampling period are reported in Table A.7-3.

³Because monitoring may not capture peak concentrations, maximum concentrations may be underestimated.

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)	•		(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 μg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
(3 sites)		2012	0.00139	11	11	100	0.033	0.031	0.025	0.1	5 (45%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,		2013	0.00139	12	12	100	0.3	0.38	0.18	1.3	10 (83%)	3 (25%)	4 (33%)	0 (0%)
2532563, 2681876,		2014	0.00139	14	14	100	0.2	0.22	0.084	0.64	12 (86%)	4 (29%)	4 (29%)	0 (0%)
2703534, 2834287;		2015	0.00139	14	14	100	0.15	0.29	0.027	1.1	7 (50%)	1 (7%)	3 (21%)	0 (0%)
2011 data are from ECCC, as cited in PMRA# 2526820)		2016	0.00139	5	5	100	0.013	0.016	0.007	0.041	1 (20%)	0 (0%)	0 (0%)	0 (0%)
Four Mile Creek	Vineyards,	2012	0.00139	14	7	50	0.013	0.026	0.001	0.097	2 (14%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,	orchards,	2013	0.00139	12	9	75	0.032	0.041	0.015	0.12	4 (33%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,	soybeans	2014	0.00139	14	9	64	0.013	0.02	0.004	0.073	2 (14%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287)		2015	0.00139	13	8	62	0.031	0.056	0.003	0.2	4 (31%)	0 (0%)	0 (0%)	0 (0%)
		2016	0.00139	6	2	33	0.002	0.003	0.0007	0.007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Big Creek (PMRA# 2523839, 2703534, 2834287)	Corn, soybeans, wheat	2012	0.00139	14	6	43	0.004	0.011	0.0007	0.042	1 (7%)	0 (0%)	0 (0%)	0 (0%)
Innisfil Creek	Soybeans, corn,	2011	0.00139	1	1	100	0.006	NA	0.006	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,	wheat	2012	0.00139	13	13	100	0.014	0.018	0.01	0.073	1 (8%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2013	0.00139	11	11	100	0.029	0.046	0.012	0.14	2 (18%)	0 (0%)	0 (0%)	0 (0%)
Lebo Drain	Soybeans, corn,	2013	0.00139	12	12	100	0.15	0.12	0.1	0.37	12 (100%)	0 (0%)	3 (25%)	0 (0%)
(PMRA# 2523839,	tomatoes, wheat,	2014	0.00139	14	14	100	0.13	0.13	0.11	0.55	14 (100%)	1 (7%)	1 (7%)	0 (0%)
2532563, 2681876,	greenhouses	2015	0.00139	13	13	100	0.11	0.22	0.042	0.85	11 (85%)	1 (8%)	1 (8%)	0 (0%)
2703534, 2834287)		2016	0.00139	6	6	100	0.052	0.021	0.056	0.079	5 (83%)	0 (0%)	0 (0%)	0 (0%)
Lebo Drain 1 (PMRA# 2818733)	Corn, soybeans, greenhouses	2017	0.002	13	13	100	0.11	0.15	0.061	0.6	1 (8%)	1 (8%)	1 (8%)	0 (0%)
Lebo Drain 10 (PMRA# 2818733)	Greenhouses, soybeans, tomatoes	2017	0.002	9	8	89	0.077	0.11	0.02	0.3	4 (44%)	0 (0%)	1 (11%)	0 (0%)
Lebo Drain 2 (PMRA# 2818733)	Soybeans, tomatoes, greenhouses	2017	0.002	13	13	100	0.087	0.063	0.066	0.26	12 (92%)	0 (0%)	0 (0%)	0 (0%)
Site 200m downstream from Lebo Drain 2 (PMRA# 2818733)	Soybeans, tomatoes, greenhouses	2017	0.002	5	5	100	0.044	0.015	0.049	0.057	4 (80%)	0 (0%)	0 (0%)	0 (0%)
Lebo Drain 3 (PMRA# 2818733)	Soybeans, wheat, tomatoes	2017	0.002	8	8	100	0.22	0.48	0.035	1.4	5 (63%)	1 (13%)	1 (13%)	0 (0%)
Lebo Drain 4 (PMRA# 2818733)	Soybeans, tomatoes	2017	0.002	13	13	100	0.11	0.074	0.075	0.27	13 (100%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	9/6	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)			(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
Lebo Drain 5 (PMRA# 2818733)	Greenhouses, soybeans, tomatoes	2017	0.002	12	12	100	0.072	0.084	0.042	0.33	10 (83%)	0 (0%)	1 (8%)	0 (0%)
Lebo Drain 6 (PMRA# 2818733)	Soybeans, tomatoes, wheat	2017	0.002	11	11	100	0.052	0.031	0.04	0.13	10 (91%)	0 (0%)	0 (0%)	0 (0%)
Lebo Drain 7 (PMRA# 2818733)	Corn, tomatoes	2017	0.002	10	10	100	0.083	0.09	0.054	0.33	10 (100%)	0 (0%)	1 (10%)	0 (0%)
Lebo Drain 8 (PMRA# 2818733)	Greenhouses, tomatoes, corn	2017	0.002	10	10	100	0.14	0.22	0.054	0.76	9 (90%)	1 (10%)	1 (10%)	0 (0%)
Lebo Drain 9 (PMRA# 2818733)	Greenhouses, soybeans, corn	2017	0.002	9	9	100	0.12	0.2	0.046	0.65	7 (78%)	1 (11%)	1 (11%)	0 (0%)
Nissouri Creek	Corn, soybeans	2013	0.00139	12	8	67	0.004	0.005	0.002	0.018	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,		2015	0.00139	12	7	58	0.006	0.009	0.002	0.026	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2681876, 2703534, 2834287)		2016	0.00139	6	4	67	0.003	0.002	0.003	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Nottawasaga River	Soybeans, corn,	2012	0.00139	13	13	100	0.016	0.017	0.008	0.053	3 (23%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839, 2703534, 2834287)	wheat	2013	0.00139	11	11	100	0.032	0.032	0.013	0.084	4 (36%)	0 (0%)	0 (0%)	0 (0%)
Prudhomme Creek	Orchards,	2011	0.00139	1	1	100	0.001	NA	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(Old Vineland	vineyards,	2012	0.00139	13	8	62	0.015	0.039	0.003	0.14	1 (8%)	0 (0%)	0 (0%)	0 (0%)
Creek)	urban/developed	2013	0.00139	11	10	91	0.031	0.054	0.005	0.14	2 (18%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,		2014	0.00139	14	9	64	0.003	0.003	0.003	0.009	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,		2015	0.00139	14	7	50	0.003	0.006	0.001	0.023	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2016	0.00139	6	2	33	0.003	0.004	0.001	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Sturgeon Creek	Soybeans, corn,	2012	0.00139	12	12	100	0.018	0.013	0.013	0.048	2 (17%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,	greenhouses,	2013	0.00139	12	12	100	0.013	0.012	0.008	0.04	2 (17%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,	wheat, tomatoes	2014	0.00139	14	13	93	0.008	0.005	0.008	0.021	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287)		2015	0.00139	13	13	100	0.051	0.095	0.009	0.29	4 (31%)	0 (0%)	0 (0%)	0 (0%)
		2016	0.00139	6	6	100	0.032	0.015	0.029	0.061	5 (83%)	0 (0%)	0 (0%)	0 (0%)
Sturgeon Creek 1 (PMRA# 2818733)	Greenhouses, soybeans, tomatoes	2017	0.002	13	13	100	0.12	0.14	0.058	0.54	12 (92%)	1 (8%)	1 (8%)	0 (0%)
Sturgeon Creek 2 (PMRA# 2818733)	Soybeans	2017	0.002	8	7	88	0.032	0.034	0.022	0.085	4 (50%)	0 (0%)	0 (0%)	0 (0%)
Sturgeon Creek 3 (PMRA# 2818733)	Greenhouses, soybeans	2017	0.002	13	13	100	0.04	0.023	0.045	0.091	8 (62%)	0 (0%)	0 (0%)	0 (0%)
Sturgeon Creek 4 (PMRA# 2818733)	Greenhouses, tomatoes	2017	0.002	9	9	100	0.039	0.025	0.032	0.078	6 (67%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)			(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC5 of	Chronic EC ₁₀	Mesocosm	Acute
											0.026 μg/L	of 0.43 μg/L	NOEC of	HC5 of
													0.3 μg/L	9 μg/L
LE1 (PMRA# 2818733)	Corn, tomatoes	2017	0.002	13	13	100	0.046	0.049	0.033	0.2	9 (69%)	0 (0%)	0 (0%)	0 (0%)
Sydenham River	Soybeans, corn,	2012	0.00139	17	16	94	0.087	0.19	0.008	0.74	7 (41%)	1 (6%)	2 (12%)	0 (0%)
(PMRA# 2523839,	wheat	2013	0.00139	10	10	100	0.076	0.15	0.032	0.5	6 (60%)	1 (10%)	1 (10%)	0 (0%)
2532563, 2681876,		2014	0.00139	14	14	100	0.028	0.046	0.012	0.18	4 (29%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287)		2015	0.00139	13	13	100	0.065	0.15	0.006	0.57	6 (46%)	1 (8%)	1 (8%)	0 (0%)
		2016	0.00139	6	6	100	0.016	0.021	0.009	0.058	1 (17%)	0 (0%)	0 (0%)	0 (0%)
Thames River	Corn, soybeans,	2011	0.00139	1	1	100	0.008	NA	0.008	0.008	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,	wheat	2012	0.00139	17	17	100	0.018	0.02	0.007	0.058	5 (29%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,		2013	0.00139	11	11	100	0.038	0.043	0.025	0.13	5 (45%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287,		2015	0.00139	12	12	100	0.1	0.24	0.011	0.83	4 (33%)	1 (8%)	1 (8%)	0 (0%)
2011 data are from ECCC, as cited in PMRA# 2526820)		2016	0.00139	6	6	100	0.009	0.006	0.007	0.018	0 (0%)	0 (0%)	0 (0%)	0 (0%)
West Holland River (PMRA# 2523839, 2703534, 2834287)	Soybeans, corn, vegetables, wheat	2013	0.00139	13	13	100	0.021	0.022	0.016	0.079	3 (23%)	0 (0%)	0 (0%)	0 (0%)
Indian Creek	Urban/developed	2011	0.00139	2	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,	•	2012	0.00139	14	4	29	0.001	0.001	0.0007	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,		2013	0.00139	11	3	27	0.018	0.054	0.0007	0.18	1 (9%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287;		2014	0.00139	8	2	25	0.001	0.001	0.0007	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2011 data are from		2015	0.00139	12	3	25	0.003	0.005	0.0007	0.016	0 (0%)	0 (0%)	0 (0%)	0 (0%)
ECCC, as cited in PMRA# 2526820)		2016	0.00139	5	2	40	0.001	0.001	0.0007	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Credit River (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	0.00139	1	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Highland Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	0.00139	1	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Kossuth (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	0.00139	1	1	100	0.001	NA	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Lake Erie (4 stations) (PMRA# 2523839)	Not applicable; sites were not near the shore	2013	0.00139	4	1	25	0.003	0.005	0.0007	0.011	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	º/o	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	lpoints
(Data source)	•		(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC, of 0.026 µg/L	Chronic EC ₁₀ of 0.43 μg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
Lgrand (data from ECCC, as cited in PMRA# 2526820)	Row crops	2011	0.00139	1	1	100	0.002	NA	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mimico Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	0.00139	1	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Nott-baxter and Nott-SR10 sites (2 sites) (data from ECCC, as cited in PMRA# 2526820)	Potatoes	2011	0.00139	2	2	100	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Spencer Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	0.00139	1	1	100	0.001	NA	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Spring Creek	Reference site	2012	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2523839,		2013	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2532563, 2681876,		2014	0.00139	7	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2703534, 2834287)		2015	0.00139	6	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
T1 C1-	TT-1C	2016	0.00139	4	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Taylor Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	0.00139	1	0	0	0.0007	NA	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Welland (data from ECCC, as cited in PMRA# 2526820)	Row crops	2011	0.00139	1	1	100	0.006	NA	0.006	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Batteaux River (PMRA# 2523836, 2759002)	Urban, shrubland, forest	2012- 2014	0.09	18	0	0	0.045	0	0.045	0.045	0 detects, 18 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Boomer Creek (PMRA# 2523836, 2759002)	Corn, pasture, wheat, hemp	2012- 2014	0.09	18	0	0	0.045	0	0.045	0.045	0 detects, 18 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Decker Creek (PMRA# 2523836, 2759002)	Corn, soybean cereals, orchards	2012- 2014	0.09	17	1	6	0.069	0.098	0.045	0.45	1 detect, 17 samples (100%) ¹	1 (6%)	1 (6%)	0 (0%)
Don River (PMRA# 2523836, 2759002)	Urban	2012	0.09	1	0	0	0.045	0	0.045	0.045	0 detects, 1 sample (100%) ¹	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	lpoints
(Data source)	·		(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Four Mile Creek (PMRA# 2523836, 2759002)	Orchards, corn, soybeans, vineyards, greenhouses	2012- 2014	0.09	18	0	0	0.045	0	0.045	0.045	0 detects, 18 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Grand River (PMRA# 2523836, 2759002)	Urban, forest, pasture, corn, soybeans	2012- 2014	0.09	17	0	0	0.045	0	0.045	0.045	0 detects, 17 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Gregory Creek (PMRA# 2523836, 2759002)	Corn, soybeans, wheat, cereals	2012- 2014	0.09	14	0	0	0.045	0	0.045	0.045	0 detects, 14 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Griffins Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	0.09	16	0	0	0.045	0	0.045	0.045	0 detects, 16 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Humber River (PMRA# 2523836, 2759002)	Urban	2012- 2014	0.09	20	0	0	0.045	0	0.045	0.045	0 detects, 20 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Lebo Drain (PMRA# 2523836, 2759002)	Corn, soybeans, wheat, vegetables	2012- 2014	0.09	16	3	19	0.097	0.11	0.045	0.35	3 detects, 16 samples (100%) ¹	0 (0%)	2 (13%)	0 (0%)
Little Ausable River (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, hemp	2012	0.09	2	0	0	0.045	0	0.045	0.045	0 detects, 2 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
McGregor Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, vegetables	2012- 2014	0.09	18	2	11	0.15	0.3	0.045	1.1	2 detects, 18 samples (100%) ¹	2 (11%)	2 (11%)	0 (0%)
McKillop Drain (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	0.09	18	3	17	0.11	0.15	0.045	0.52	3 detects, 18 samples (100%) ¹	1 (6%)	3 (17%)	0 (0%)
Nissouri Creek (PMRA# 2523836, 2759002)	Corn, soybeans, wheat, pasture	2013	0.09	2	0	0	0.045	0	0.045	0.045	0 detects, 2 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Otter Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	0.09	16	0	0	0.045	0	0.045	0.045	0 detects, 16 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Reynolds Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat, hemp	2012- 2014	0.09	17	0	0	0.045	0	0.045	0.045	0 detects, 17 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Saugeen River (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	0.09	17	0	0	0.045	0	0.045	0.045	0 detects, 17 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC, of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Thames River (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	0.09	18	0	0	0.045	0	0.045	0.045	0 detects, 18 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Venison Creek (PMRA# 2523836, 2759002)	Corn, soybeans, forest, wheat, orchards	2012- 2014	0.09	17	0	0	0.045	0	0.045	0.045	0 detects, 17 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Whitemans Creek (PMRA# 2523836, 2759002)	Corn, soybeans, tobacco, other crops	2012- 2014	0.09	18	0	0	0.045	0	0.045	0.045	0 detects, 18 samples (100%) ¹	0 (0%)	0 (0%)	0 (0%)
Big Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	0.005	23	23	100	0.15	0.31	0.068	1.5	19 (83%)	1 (4%)	2 (9%)	0 (0%)
Garvey Glenn (PMRA# 2712893)	Corn, soybeans, wheat	2015	0.005	19	14	74	0.042	0.12	0.007	0.51	5 (26%)	1 (5%)	1 (5%)	0 (0%)
Little Ausable Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	0.005	17	10	59	0.07	0.15	0.011	0.58	7 (41%)	1 (6%)	1 (6%)	0 (0%)
North Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	0.005	19	17	89	0.37	0.76	0.057	2.7	12 (63%)	3 (16%)	4 (21%)	0 (0%)
White Ash Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	0.005	18	8	44	0.021	0.055	0.003	0.24	3 (17%)	0 (0%)	0 (0%)	0 (0%)
Hamilton Harbour, WWTP influent and effluent (PMRA# 2710505)	Urban	2016	0.005	6	0	0	0.003	0.003	0	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Grand River, WWTP influent and effluent (PMRA# 2710505)	Urban, corn, soybeans	2016	0.005	12	1	8	0.003	0.001	0.003	0.007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Detroit River, WWTP influent and effluent (PMRA# 2710505)	Urban	2016	0.005	6	0	0	0.003	0	0.003	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Little River, WWTP influent and effluent (PMRA# 2710505)	Urban	2016	0.005	6	0	0	0.003	0	0.003	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Presqu'île Bay, WWTP influent and effluent (PMRA# 2710505)	Urban, corn, soybeans	2016	0.005	2	0	0	0.003	0	0.003	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Cootes Paradise, WWTP influent and effluent (PMRA# 2710505)	Urban, forest, corn, soybeans	2016	0.005	6	0	0	0.003	0	0.003	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)			(μg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 μg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Ditches around corn fields ² (PMRA# 2526184)	Corn	2013	0.004	22	22	100	1.22	1.82	0.392	7.52	17 (77%) ²	11 (50%) ²	11 (50%) ²	0 (0%)2
Drainage tile outlets around corn fields ² (PMRA# 2526184)	Corn	2013	0.004	8	7	88	0.77 ²	0.96 ²	0.342	2.62	6 (75%) ²	4 (50%) ²	4 (50%) ²	0 (0%) ²
Creeks, streams, ponds	Agriculture	2013	0.001 (LOQ)	42	5	12	0.002	0.006	0.005	0.034	1 (2%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2548877)		2014	0.0008	14	10	71	0.001	0.005	0.002	0.016	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Streams, culverts, ditches (PMRA# 2548876)	Agriculture	2014	0.0008	5	2	40	0.011	0.022	0.0004	0.05	1 (20%)	0 (0%)	0 (0%)	0 (0%)
Black Creek	Corn, soybeans,	2015	0.00009	2	2	100	0.036	0.026	0.036	0.054	1 (50%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	2	100	0.011	0.004	0.011	0.013	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Beckstead	Corn, soybeans,	2015	0.00009	2	1	50	0.003	0.004	0.003	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	1	50	0.001	0.001	0.001	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
East Branch Scotch	Forest, corn,	2015	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
River (PMRA# 2785041)	soybeans, wheat	2016	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
East Castor	Corn, soybeans,	2015	0.00009	2	2	100	0.042	0.032	0.042	0.065	1 (50%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	pasture, wheat	2016	0.00009	2	2	100	0.005	0.001	0.005	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Greenough	Corn, soybeans,	2015	0.00009	1	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	pasture	2016	0.00009	2	1	50	0.001	0.001	0.001	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Kirkwood	Corn, soybeans,	2015	0.00009	2	2	100	0.007	0.001	0.007	0.007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	2	100	0.007	0.004	0.007	0.009	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Little Castor	Corn, soybeans,	2015	0.00009	2	2	100	0.019	0.016	0.019	0.03	1 (50%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	2	100	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
McLeod	Corn, soybeans,	2015	0.00009	2	1	50	0.001	0.002	0.001	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Middle Castor River	Corn, soybeans,	2015	0.00009	2	1	50	0.017	0.024	0.017	0.035	1 (50%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	2	100	0.002	0.001	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
North Branch South	Corn, soybeans,	2015	0.00009	2	2	100	0.011	0.011	0.011	0.019	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Nation (PMRA# 2785041)	wheat	2016	0.00009	2	1	50	0.0005	0.001	0.0005	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Nugent	Corn, soybeans,	2015	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	1	50	0.001	0.001	0.001	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Payne River	Corn, soybeans,	2015	0.00009	2	2	100	0.035	0.03	0.035	0.056	1 (50%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	2	100	0.036	0.037	0.036	0.062	1 (50%)	0 (0%)	0 (0%)	0 (0%)
Shane	Corn, soybeans,	2015	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	1	50	0.0003	0.0003	0.0003	0.0005	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of samp	les) exceeding th	e toxicity end	points
(Data source)			(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 μg/L	Mesocosm NOEC of 0.3 ug/L	Acute HC ₅ of 9 μg/L
St. Edouard Road	Corn, soybeans,	2015	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)	wheat	2016	0.00009	2	0	0	0.00005	0	0.00005	0.00005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
West Branch Scotch	Corn, soybeans,	2015	0.00009	2	2	100	0.017	0.017	0.017	0.029	1 (50%)	0 (0%)	0 (0%)	0 (0%)
River (PMRA#	wheat	2016	0.00009	2	2	100	0.017	0.018	0.017	0.029	1 (50%)	0 (0%)	0 (0%)	0 (0%)
2785041)														
Whittaker	Corn, soybeans	2015	0.00009	2	1	50	0.002	0.002	0.002	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2785041)		2016	0.00009	2	1	50	0.003	0.004	0.003	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)

LOD = limit of detection, N = sample size; Stdev = standard deviation; Chronic HC_5 = the 5^{th} percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC_{10} = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC_5 = the 5^{th} percentile of the species sensitivity distribution for the LC_{50} (the median lethal concentration) at 50% confidence intervals; ECCC = Environment and Climate Change Canada; WWTP = waste water treatment plant; LOQ = limit of quantification

Table A.7-6 Risk quotients for thiamethoxam measured in waterbodies located in Ontario.

NOTES:

- -Shaded cells indicate the level of concern is exceeded, meaning that the risk quotient is equal to or greater than a value of 1.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred one to four times per month between April and November. Sampling at some sites occurred only a few times over a short time period, and values measured may not represent concentrations throughout the growing season.

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Two Mile Creek	Vineyards,	2012	15	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA#	orchards	2013	14	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2523839,		2014	12	0.5	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
2532563,		2015	13	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
2681876, 2703534, 2834287)		2016	6	0.4	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Twenty Mile	Soybeans, corn	2011	1	6.7	0.4	0.6	6.7	0.4	0.6	< 0.1

¹The LOD is more than two times higher than the chronic HC_5 of 0.026 μ g/L. Assigning half the limit of detection to non-detected samples still results in a concentration exceeding the toxicity endpoint. Thus, all samples, including non-detects at half the limit of detection, exceed the toxicity endpoint.

²Ditches and tile drain outlets around corn fields may not represent aquatic habitat.

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Creek (3 sites)		2012	11	1.3	0.1	0.1	0.9	0.1	0.1	< 0.1
(PMRA#		2013	12	12	0.7	1	6.7	0.4	0.6	0.1
2523839,		2014	14	7.7	0.5	0.7	3.2	0.2	0.3	0.1
2532563,		2015	14	5.9	0.4	0.5	1	0.1	0.1	0.1
2681876, 2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2016	5	0.5	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
Four Mile Creek	Vineyards,	2012	14	0.5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA#	orchards,	2013	12	1.2	0.1	0.1	0.6	< 0.1	0.1	< 0.1
2523839,	soybeans	2014	14	0.5	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
2532563,		2015	13	1.2	0.1	0.1	0.1	< 0.1	< 0.1	< 0.1
2681876, 2703534, 2834287)		2016	6	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Big Creek (PMRA# 2523839, 2703534, 2834287)	Corn, soybeans, wheat	2012	14	0.2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Innisfil Creek	Soybeans, corn,	2011	1	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
(PMRA#	wheat	2012	13	0.6	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
2523839, 2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2013	11	1.1	0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
Lebo Drain	Soybeans, corn,	2013	12	5.8	0.4	0.5	4	0.2	0.3	< 0.1
(PMRA#	tomatoes, wheat,	2014	14	5	0.3	0.4	4.1	0.2	0.4	0.1
2523839,	greenhouses	2015	13	4.2	0.3	0.4	1.6	0.1	0.1	0.1
2532563, 2681876, 2703534, 2834287)		2016	6	2	0.1	0.2	2.2	0.1	0.2	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concer ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Lebo Drain 1 (PMRA# 2818733)	Corn, soybeans, greenhouses	2017	13	4.3	0.3	0.4	2.3	0.1	0.2	0.1
Lebo Drain 10 (PMRA# 2818733)	Greenhouses, soybeans, tomatoes	2017	9	3	0.2	0.3	0.8	< 0.1	0.1	< 0.1
Lebo Drain 2 (PMRA# 2818733)	Soybeans, tomatoes, greenhouses	2017	13	3.3	0.2	0.3	2.5	0.2	0.2	< 0.1
Site 200m downstream from Lebo Drain 2 (PMRA# 2818733)	Soybeans, tomatoes, greenhouses	2017	5	1.7	0.1	0.1	1.9	0.1	0.2	< 0.1
Lebo Drain 3 (PMRA# 2818733)	Soybeans, wheat, tomatoes	2017	8	8.3	0.5	0.7	1.4	0.1	0.1	0.2
Lebo Drain 4 (PMRA# 2818733)	Soybeans, tomatoes	2017	13	4.1	0.2	0.4	2.9	0.2	0.2	< 0.1
Lebo Drain 5 (PMRA# 2818733)	Greenhouses, soybeans, tomatoes	2017	12	2.8	0.2	0.2	1.6	0.1	0.1	< 0.1
Lebo Drain 6 (PMRA# 2818733)	Soybeans, tomatoes, wheat	2017	11	2	0.1	0.2	1.5	0.1	0.1	< 0.1
Lebo Drain 7 (PMRA# 2818733)	Corn, tomatoes	2017	10	3.2	0.2	0.3	2.1	0.1	0.2	< 0.1
Lebo Drain 8 (PMRA# 2818733)	Greenhouses, tomatoes, corn	2017	10	5.2	0.3	0.5	2.1	0.1	0.2	0.1
Lebo Drain 9 (PMRA# 2818733)	Greenhouses, soybeans, corn	2017	9	4.6	0.3	0.4	1.8	0.1	0.2	0.1
Nissouri Creek (PMRA#	Corn, soybeans	2013 2015	12 12	0.2	< 0.1 < 0.1	< 0.1 < 0.1	0.1	< 0.1 < 0.1	< 0.1 < 0.1	< 0.1 < 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien gmedian ² concen c toxicity endpoi	strations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
2523839, 2681876, 2703534, 2834287)		2016	6	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Nottawasaga	Soybeans, corn,	2012	13	0.6	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
River (PMRA# 2523839, 2703534, 2834287)	wheat	2013	11	1.2	0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
Prudhomme Creek		2011	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
(Old Vineland	vineyards,	2012	13	0.6	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Creek) (PMRA#	urban/developed	2013	11	1.2	0.1	0.1	0.2	< 0.1	< 0.1	< 0.1
2523839,		2014	14	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
2532563,		2015	14	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2681876, 2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2016	6	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Sturgeon Creek	Soybeans, corn,	2012	12	0.7	< 0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
(PMRA#	greenhouses,	2013	12	0.5	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
2523839,	wheat, tomatoes	2014	14	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
2532563,		2015	13	1.9	0.1	0.2	0.4	< 0.1	< 0.1	< 0.1
2681876, 2703534, 2834287)		2016	6	1.2	0.1	0.1	1.1	0.1	0.1	< 0.1
Sturgeon Creek 1 (PMRA# 2818733)	Greenhouses, soybeans, tomatoes	2017	13	4.6	0.3	0.4	2.2	0.1	0.2	0.1
Sturgeon Creek 2 (PMRA# 2818733)	Soybeans	2017	8	1.2	0.1	0.1	0.9	0.1	0.1	< 0.1
Sturgeon Creek 3 (PMRA# 2818733)	Greenhouses, soybeans	2017	13	1.5	0.1	0.1	1.7	0.1	0.2	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concer ic toxicity endpo	itrations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Sturgeon Creek 4 (PMRA# 2818733)	Greenhouses, tomatoes	2017	9	1.5	0.1	0.1	1.2	0.1	0.1	< 0.1
LE1 (PMRA# 2818733)	Corn, tomatoes	2017	13	1.8	0.1	0.2	1.3	0.1	0.1	< 0.1
Sydenham River	Soybeans, corn,	2012	17	3.3	0.2	0.3	0.3	< 0.1	< 0.1	0.1
(PMRA#	wheat	2013	10	2.9	0.2	0.3	1.2	0.1	0.1	0.1
2523839,		2014	14	1.1	0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
2532563,		2015	13	2.5	0.2	0.2	0.2	< 0.1	< 0.1	0.1
2681876, 2703534, 2834287)	681876, 703534, 834287) hames River	2016	6	0.6	< 0.1	0.1	0.4	< 0.1	< 0.1	< 0.1
Thames River	MRA# wheat	2011	1	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
(PMRA#		2012	17	0.7	< 0.1	0.1	0.3	< 0.1	< 0.1	< 0.1
2523839,		2013	11	1.5	0.1	0.1	1	0.1	0.1	< 0.1
2532563,		2015	12	- 4	0.2	0.3	0.4	< 0.1	< 0.1	0.1
2681876, 2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2016	6	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
West Holland River (PMRA# 2523839, 2703534, 2834287)	Soybeans, corn, vegetables, wheat	2013	13	0.8	< 0.1	0.1	0.6	< 0.1	0.1	< 0.1
Indian Creek	Urban/developed	2011	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA#		2012	14	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2523839,		2013	11	0.7	< 0.1	0.1	< 0.1	< 0.1	< 0.1	< 0.1
2532563,		2014	8	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2681876,		2015	12	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2703534, 2834287; 2011 data are from ECCC, as cited in PMRA# 2526820)		2016	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotien g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Credit River (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Highland Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Kossuth (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Lake Erie (4 stations) (PMRA# 2523839)	Not applicable; sites were not near the shore	2013	4	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Lgrand (data from ECCC, as cited in PMRA# 2526820)	Row crops	2011	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Mimico Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Nott-baxter and Nott-SR10 sites (2 sites) (data from ECCC, as cited in PMRA# 2526820)		2011	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Spencer Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Spring Creek	Reference site	2012	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA#		2013	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2523839,		2014	7	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
2532563,		2015	6	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce- nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien median ² concer c toxicity endpo	itrations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
2681876, 2703534, 2834287)		2016	4	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Taylor Creek (data from ECCC, as cited in PMRA# 2526820)	Urban or turf	2011	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Welland (data from ECCC, as cited in PMRA# 2526820)	Row crops	2011	1	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Batteaux River (PMRA# 2523836, 2759002)	Urban, shrubland, forest	2012- 2014	18	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1
Boomer Creek (PMRA# 2523836, 2759002)	Corn, pasture, wheat, hemp	2012- 2014	18	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1
Decker Creek (PMRA# 2523836, 2759002)	Corn, soybean cereals, orchards	2012- 2014	17	2.65	0.2	0.2	1.75	0.1	0.2	< 0.1
Don River (PMRA# 2523836, 2759002)	Urban	2012	1	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1
Four Mile Creek (PMRA# 2523836, 2759002)	Orchards, corn, soybeans, vineyards, greenhouses	2012- 2014	18	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Grand River (PMRA# 2523836, 2759002)	Urban, forest, pasture, corn, soybeans	2012- 2014	17	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotien g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Gregory Creek (PMRA# 2523836, 2759002)	Corn, soybeans, wheat, cereals	2012- 2014	14	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Griffins Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	16	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1
Humber River (PMRA# 2523836, 2759002)	Urban	2012- 2014	20	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1
Lebo Drain (PMRA# 2523836, 2759002)	Corn, soybeans, wheat, vegetables	2012- 2014	16	3.75	0.2	0.3	1.75	0.1	0.2	< 0.1
Little Ausable River (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, hemp	2012	2	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1
McGregor Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, vegetables	2012- 2014	18	5.65	0.3	0.5	1.75	0.1	0.2	0.1
McKillop Drain (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	18	4.15	0.3	0.4	1.75	0.1	0.2	0.1
Nissouri Creek (PMRA# 2523836, 2759002)	Corn, soybeans, wheat, pasture	2013	2	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Otter Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	16	1.71	0.1	0.2	1.74	0.1	0.2	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier ng average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Reynolds Creek (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat, hemp	2012- 2014	17	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Saugeen River (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	17	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Thames River (PMRA# 2523836, 2759002)	Corn, soybeans, cereals, wheat	2012- 2014	18	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Venison Creek (PMRA# 2523836, 2759002)	Corn, soybeans, forest, wheat, orchards	2012- 2014	17	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Whitemans Creek (PMRA# 2523836, 2759002)	Corn, soybeans, tobacco, other crops	2012- 2014	18	1.74	0.1	0.2	1.74	0.1	0.2	< 0.1
Big Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	23	5.9	0.4	0.5	2.6	0.2	0.2	0.2
Garvey Glenn (PMRA# 2712893)	Corn, soybeans, wheat	2015	19	1.6	0.1	0.1	0.3	< 0.1	< 0.1	0.1
Little Ausable Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	17	2.7	0.2	0.2	0.4	< 0.1	< 0.1	0.1
North Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	19	14	0.9	1.2	2.2	0.1	0.2	0.3
White Ash Creek (PMRA# 2712893)	Corn, soybeans, wheat	2015	18	0.8	< 0.1	0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier ig average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concer ic toxicity endpo	itrations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Hamilton Harbour, WWTP influent and effluent (PMRA# 2710505)	Urban	2016	6	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Grand River, WWTP influent and effluent (PMRA# 2710505)	Urban, corn, soybeans	2016	12	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Detroit River, WWTP influent and effluent (PMRA# 2710505)	Urban	2016	6	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Little River, WWTP influent and effluent (PMRA# 2710505)	Urban	2016	6	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Presqu'île Bay, WWTP influent and effluent (PMRA# 2710505)	Urban, corn, soybeans	2016	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Cootes Paradise, WWTP influent and effluent (PMRA# 2710505)	Urban, forest, corn, soybeans	2016	6	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Ditches around corn fields ⁶ (PMRA# 2526184)	Corn	2013	22	45°	2.76	3.96	156	0.96	1.36	0.86

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotier g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Drainage tile outlets around corn fields ⁶ (PMRA# 2526184)	Corn	2013	8	29 ⁶	1.8°	2.66	13 ⁶	0.86	1.16	0.3 ⁶
Creeks, streams,	Agriculture	2013	42	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
ponds (PMRA# 2548877)		2014	12	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Streams, culverts, ditches (PMRA# 2548876)	Agriculture	2014	5	0.4	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Black Creek	Corn, soybeans,	2015	2	1.4	0.1	0.1	1.4	0.1	0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	0.4	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
Beckstead	Corn, soybeans,	2015	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
East Branch	Forest, corn,	2015	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Scotch River (PMRA# 2785041)	soybeans, wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
East Castor	Corn, soybeans,	2015	2	1.6	0.1	0.1	1.6	0.1	0.1	< 0.1
(PMRA# 2785041)	pasture, wheat	2016	2	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Greenough	Corn, soybeans,	2015	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)	pasture	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Kirkwood	Corn, soybeans,	2015	2	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
Little Castor	Corn, soybeans,	2015	2	0.7	< 0.1	0.1	0.7	< 0.1	0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
McLeod	Corn, soybeans,	2015	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usin	onic Risk Quotien g average ² conce nic toxicity endpo	ntrations and	calculated using	nic Risk Quotien g median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
(PMRA# 2785041)	wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Middle Castor	Corn, soybeans,	2015	2	0.7	< 0.1	0.1	0.7	< 0.1	0.1	< 0.1
River (PMRA# 2785041)	2785041) North Branch Corn, soybeans,	2016	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
North Branch	Corn, soybeans,	2015	2	0.4	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
South Nation (PMRA# 2785041)	wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Nugent	Corn, soybeans,	2015	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Payne River	Corn, soybeans,	2015	2	1.3	0.1	0.1	1.3	0.1	0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	1.4	0.1	0.1	1.4	0.1	0.1	< 0.1
Shane	Corn, soybeans,	2015	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
St. Edouard Road	Corn, soybeans,	2015	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)	wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
West Branch	Corn, soybeans,	2015	2	0.6	< 0.1	0.1	0.6	< 0.1	0.1	< 0.1
Scotch River (PMRA# 2785041)	wheat	2016	2	0.6	< 0.1	0.1	0.6	< 0.1	0.1	< 0.1
Whittaker	Corn, soybeans	2015	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2785041)		2016	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

N = sample size; LOD = limit of detection; Stdev = standard deviation; Chronic HC₅ = the 5^{th} percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC₁₀ = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC₅ = the 5^{th} percentile of the species sensitivity distribution for the LC₅₀ (the median lethal concentration) at 50% confidence intervals; ECCC = Environment and Climate Change Canada; WWTP = waste water treatment plant

¹Risk Quotient = concentration ÷ toxicity endpoint

²Average, median and maximum concentrations over the sampling period are reported in Table A.7-5.

³Because monitoring may not capture peak concentrations, maximum concentrations may be underestimated.

⁴The limit of detection for these samples was more than two times higher than the chronic endpoint. Even though thiamethoxam was not detected in any samples, assigning half the limit of detection to non-detected samples still results in average and median concentrations which exceed the chronic toxicity endpoint. Thus, calculated risk quotients exceed the level of concern.

Table A.7-7 Summary statistics for thiamethoxam measured in waterbodies from Manitoba, Saskatchewan and Alberta.

NOTES:

- -In calculations, the PMRA assigned a value equal to half the limit of detection to samples that showed no detection.
- -The frequency of sampling and the length of the sampling period varied between data sets. Some waterbodies were sampled one to three times between May and October, while others were sampled one to three times per month between April and December. Values measured at sites where only a few samples were collected may not represent concentrations throughout the growing season.

Waterbody	Major land use	Year	LOD	N	N	º/o	Average	Stdev	Median	Max	N (% of san	nples) exceed	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
						Mar	iitoba							
Red River at Emerson	Soybeans, wheat,	2014	0.00139	7	7	100	0.017	0.018	0.014	0.052	2 (29%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2745819)	canola, oats, corn	2015	0.00139	6	5	83	0.016	0.011	0.017	0.031	1 (17%)	0 (0%)	0 (0%)	0 (0%)
		2016	0.00139	1	1	100	0.009	NA	0.009	0.009	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red River at Emerson (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats, corn	2017	0.0027	3	3	100	0.02	0.018	0.013	0.04	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Red River at Selkirk (PMRA# 2745819)	Soybeans, wheat, canola, oats	2014	0.00139	1	1	100	0.013	NA	0.013	0.013	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red River at Selkirk (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats	2017	0.0027	3	2	67	0.019	0.027	0.007	0.05	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Red River at Norbert (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	0.0027	3	3	100	0.019	0.018	0.011	0.04	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Assiniboine River Northwest of Treesbank (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, corn	2017	0.0027	3	2	67	0.003	0.002	0.003	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Assiniboine River at Happy Hollow Farm (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, corn	2017	0.0027	3	2	67	0.003	0.002	0.003	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)

⁵The limit of detection for these samples was more than two times higher than the chronic endpoint. Even though thiamethoxam was not detected in most samples, assigning half the limit of detection to non-detected samples still results in average and median concentrations which exceed the chronic toxicity endpoint. Thus, calculated risk quotients exceed the level of concern.

⁶Ditches and tile drain outlets around corn fields may not represent aquatic habitat.

Waterbody	Major land use	Year	LOD	N	N	º/o	Average	Stdev	Median	Max	N (% of san	ples) exceed	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
Assiniboine River downstream of Portage la Prairie (PMRA# 2849359, 2849370)	Soybeans, wheat, canola	2017	0.0027	3	2	67	0.006	0.005	0.004	0.011	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Assiniboine River at Headingley (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, oats, barley, corn	2017	0.0027	3	2	67	0.005	0.004	0.005	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Assiniboine River at Provincial Trunk Highway 21, North of Griswold (PMRA# 2849359, 2849370)	Canola, wheat, soy beans, barley	2017	0.0027	3	2	67	0.013	0.014	0.009	0.028	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Assiniboine River at Provincial Trunk Highway 83, South of Miniota (PMRA# 2849359, 2849370)	Canola, wheat, soy beans, barley	2017	0.0027	3	2	67	0.052	0.081	0.01	0.15	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Boyne River (PMRA# 2849359, 2849370)	Soybeans, corn, wheat, canola, oats	2017	0.0027	3	3	100	0.043	0.049	0.025	0.099	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Cooks Creek at Rural Municipality Boundary Road (PMRA# 2849359, 2849370)	Soybeans, canola, oats, corn, wheat	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Cooks Creek south of Millbrook (PMRA# 2849359, 2849370)	Soybeans, canola, oats, corn, wheat	2017	0.0027	3	2	67	0.049	0.079	0.006	0.14	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Edwards Creek (PMRA# 2849359, 2849370)	Canola, soybeans, wheat	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Icelandic River (EMWG)	Soybeans, wheat, canola, oats	2017	0.0027	3	1	33	0.018	0.028	0.001	0.05	1 (33%)	0 (0%)	0 (0%)	0 (0%)
La Salle River at the town of La Salle (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, oats, corn	2017	0.0027	3	2	67	0.006	0.006	0.004	0.013	0 (0%)	0 (0%)	0 (0%)	0 (0%)
La Salle River at La Barriere	Soybeans, wheat, canola, corn, oats	2017	0.0027	3	2	67	0.008	0.01	0.004	0.019	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceed	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
(PMRA# 2849359, 2849370)														
Lake Manitoba (PMRA# 2849359, 2849370)	Soybeans, wheat, canola	2017	0.0027	2	1	50	0.003	0.003	0.003	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Lake Winnipeg (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats	2017	0.0027	3	2	67	0.006	0.005	0.007	0.011	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Little Saskatchewan River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, barley	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Morris River (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, corn, oats	2017	0.0027	3	3	100	0.039	0.037	0.034	0.078	2 (67%)	0 (0%)	0 (0%)	0 (0%)
Oak River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, barley	2017	0.0027	3	2	67	0.018	0.026	0.005	0.048	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Pelican Lake (PMRA# 2849359, 2849370)	Wheat, canola, soybeans	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Pipestone Creek (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, barley	2017	0.0027	3	1	33	0.009	0.013	0.001	0.023	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Rat River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	0.0027	3	1	33	0.003	0.002	0.001	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Rock Lake (PMRA# 2849359, 2849370)	Canola, soybeans, wheat, barley	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Roseau River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats, corn	2017	0.0027	3	1	33	0.008	0.012	0.001	0.023	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Seine River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	0.0027	2	1	50	0.002	0.002	0.002	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Souris River at the Town of Souris (PMRA# 2849359, 2849370)	Canola, soybeans, wheat, corn	2017	0.0027	2	1	50	0.005	0.006	0.005	0.009	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Souris River at Melita	Canola, wheat,	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceed	ing the toxicity	endpoints
(Data source)	•		(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
(PMRA# 2849359, 2849370)	soybeans, oats													
Sturgeon Creek (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, oats, barley, corn	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Swan River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans	2017	0.0027	3	1	33	0.009	0.013	0.001	0.024	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Willow Creek (PMRA# 2849359, 2849370)	Soybeans, wheat, canola	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Woody River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2849370) Seasonal (Class III) and semi-permanent (Class IV) wetlands ^{1,2} (PMRA# 2847073, 2847083)	Canola, wheat, oats, pasture, corn	Summe r 2017	0.002	12	8	67	NC	NC	NC	Overall range: 0.001 – 0.76; Range of detects: 0.003 – 0.76	3 wetlands (25%)	2 wetlands (17%)	2 wetlands (17%)	0 wetlands (0%)
		Fall 2017	0.002	5	0	0	NC	NC	NC	0.001	0 wetlands (0%)	0 wetlands (0%)	0 wetlands (0%)	0 wetlands (0%)
Creek (PMRA# 2548877)	Agriculture	2013	0.001 (LOQ)	1	0	0	0.0005	NA	0.0005	0.0005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Streams, culverts, ditches (PMRA# 2548876)	Agriculture	2014	0.0008	3	0	0	0.0004	0	0.0004	0.0004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
	•					Saska	tchewan							
Assiniboine River	Canola and	2014	0.00139	6	6	100	0.017	0.016	0.003	0.047	1 (17%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2745819)	rapeseed, wheat	2015	0.00139	8	3	38	0.001	0.001	0.003	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Battle River (PMRA#	Canola and	2015	0.00139	6	2	33	0.002	0.002	0.0007	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
2745819)	rapeseed, rye, wheat	2016	0.00139	2	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Avonlea Creek (PMRA# 2849265, 2849266)	Canola, peas, lentils, wheat	2017	0.0027	8	1	13	0.002	0.001	0.001	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceed	ng the toxicity	endpoints
(Data source)	,		(μg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC5 of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Lanigan Creek (PMRA# 2849265, 2849266)	Mainly canola, with some peas and wheat	2017	0.0027	8	3	38	0.006	0.01	0.001	0.028	1 (13%)	0 (0%)	0 (0%)	0 (0%)
Lightning Creek (PMRA# 2849265, 2849266)	Canola with some soybeans and wheat	2017	0.0027	10	6	60	0.011	0.014	0.006	0.046	1 (10%)	0 (0%)	0 (0%)	0 (0%)
McDonald Creek (PMRA# 2849265, 2849266)	Mainly lentils, with wheat	2017	0.0027	7	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Moose Jaw River (PMRA# 2849265, 2849266)	Lentils, canola, with wheat	2017	0.0027	9	5	56	0.005	0.005	0.005	0.016	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Moose Mountain Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	0.0027	9	2	22	0.019	0.053	0.001	0.16	1 (11%)	0 (0%)	0 (0%)	0 (0%)
Oscar Creek (PMRA# 2849265, 2849266)	Mainly canola	2017	0.0027	10	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Pipestone Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	0.0027	12	7	58	0.008	0.009	0.005	0.033	1 (8%)	0 (0%)	0 (0%)	0 (0%)
Saline Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	0.0027	10	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Souris River (PMRA# 2849265, 2849266)	Mainly canola, lentils, with wheat	2017	0.0027	9	6	67	0.01	0.011	0.006	0.034	1 (11%)	0 (0%)	0 (0%)	0 (0%)
Spirit Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	0.0027	10	9	90	0.017	0.017	0.008	0.049	3 (30%)	0 (0%)	0 (0%)	0 (0%)
Swift Current Creek below Rock Creek (PMRA# 2849265, 2849266)	Mainly lentils, with some peas, canola and wheat	2017	0.0027	8	1	13	0.002	0.002	0.001	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Swift Current Creek near Leinon (PMRA# 2849265, 2849266)	Mainly lentils, with some peas, canola and wheat	2017	0.0027	9	3	33	0.003	0.004	0.001	0.014	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Willowbrook Creek (PMRA# 2849265, 2849266)	Mainly canola	2017	0.0027	9	2	22	0.002	0.001	0.001	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Wood River (PMRA# 2849265, 2849266)	Mainly lentils, peas, with wheat	2017	0.0027	9	1	11	0.002	0.001	0.001	0.003	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max		ples) exceed	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic	Chronic	Mesocosm	Acute
											HC ₅ of 0.026 μg/L	EC ₁₀ of 0.43 μg/L	NOEC of 0.3 µg/L	HC ₅ of 9 μg/L
Temporary (Class II),	Barley, canola,	Spring	0.0018	138	13	9	NC	NC	NC	Overall	2 wetlands	0 wetlands	0 wetlands	0
seasonal (Class III),	oats, wheat,	(pre-	(LOQ)							range:	(1%)	(0%)	(0%)	wetlands
semi-permanent (Class	grassland	seed)								0.0009				(0%)
IV) and permanent (Class V) wetlands ^{1,2}	(previous year's crops)	2012								-0.032; Range				
(PMRA# 2526133,	(crops)									of				
2572395, 2608629,										detects:				
2612760, 2612761,										0.007 -				
2612762, 2712896)										0.032				
	Barley, canola,	Summe	0.0018	134	26	19	NC	NC	NC	Overall		1 wetland	1 wetland	0
	oats, wheat, peas, grassland	r 2012	(LOQ)							range: 0.0009	(10%)	(1%)	(1%)	wetlands (0%)
	grassianu									-1.5;				(076)
										Range				
										of				
										detects:				
										0.006 -				
	Barley, canola,	Fall	0.0018	80	5	6	NC	NC	NC	1.5 Overall	1 wetland	0 wetlands	0 wetlands	0
	oats, wheat, peas,	2012	(LOQ)	80	'		INC	INC	INC	range:	(1%)	(0%)	(0%)	wetlands
	grassland		(234)							0.0009	(273)	(67.6)	(0,0)	(0%)
										-0.1;				, ,
										Range				
										of detects:				
										0.01 -				
										0.1				
	Barley, canola,	Spring	0.0056	90	21	23	NC	NC	NC	Overall		0 wetlands	0 wetlands	0
	oats, wheat, peas,	(pre-	(LOQ)							range:	(20%)	(0%)	(0%)	wetlands
	grassland (previous year's	seed) 2013								0.003 - 0.11;				(0%)
	(previous year's crops)	2013								Range				
	(Crops)									of				
										detects:				
										0.011 -				
	Dorlay canala	Cumanas	0.0018	144	75	52	NC	NC	NC	0.11	53 wetlands	1 wetland	4 wetlands	0
	Barley, canola, oats, peas, wheat,	Summe r 2013	(LOQ)	144	/3	32	I NC	I NC	I NC	range:	(37%)	(1%)	4 wettands (3%)	wetlands
	flax, grassland,	12013	(100)							0.0009	(3770)	(1/0)	(370)	(0%)
	chemfallow									-0.48;				
										Range				
										of				

Waterbody	Major land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of san	iples) exceed	ing the toxicity	endpoints
(Data source)	•		(µg/L)		detects		(μg/L)		(µg/L)	(µg/L)	Chronic HC5 of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
										detects: 0.006 - 0.48				
	Canola, oats (previous year's crops)	Spring (pre- seed) 2014	0.002	16	10	63	NC	NC	NC	Overall range: 0.001 – 0.084; Range of detects: 0.002 – 0.084	1 wetland (6%)	0 wetlands (0%)	0 wetlands (0%)	0 wetlands (0%)
	Barley, canola,	Summe	0.0017-	A 11 year	tlande				<u> </u>	0.084	L			
	flax, oats, lentils, wheat, peas, soy beans, chemfallow, pasture, grassland	r 2014	0.0018	115	50	43	NC	NC	NC	Overall range: 0.0009 - 0.86; Range of detects: 0.004 - 0.86	32 wetlands (28%)	3 wetlands (3%)	6 wetlands (5%)	0 wetlands (0%)
											MRA# 287057			
				46	13	28	NC	NC	NC	Overall range: 0.0009 – 0.45; Range of detects: 0.016 – 0.45	10 wetlands (22%)	1 wetland (2%)	2 wetlands (4%)	0 wetlands (0%)
Seasonal (Class III) and semi-permanent (Class IV) wetlands ^{1,2} (PMRA# 2847073, 2847083)	Wheat, canola, barley, pasture, lentils, summer fallow	Summe r 2017	0.002	30	6	20	NC	NC	NC	Overall range: 0.001 – 0.31; Range of detects: 0.004 – 0.31	4 wetlands (13%)	0 wetlands (0%)	1 wetland (3%)	0 wetlands (0%)
1	1	1	1	1	1	1								1

Waterbody	Major land use	Year	LOD	N	N	0/0	Average	Stdev	Median	Max	N (% of san	iples) exceed	ing the toxicity	endpoints
(Data source)	•		(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
		2017									(0%)	(0%)	(0%)	wetlands (0%)
							oerta							
South Saskatchewan River (PMRA# 2745819)	Grassland, peas, wheat	2014	0.00139	5	3	60	0.003	0.002	0.002	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Oldman River (3 sites) (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	0.0027	12	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
South Saskatchewan River (PMRA# 2842307, 2842433)	Low disturbance, developed land, agriculture	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Bow River (4 sites) (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	0.0027	16	0	0	0.001	0	0.0005	0.0005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Elbow River (PMRA# 2842307, 2842433)	Developed land, low disturbance	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red Deer River (PMRA# 2745819)	Grassland, peas, wheat, canola and rapeseed	2015	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red Deer River at Sundre (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red Deer River 1 kilometre upstream of Highway 2 Bridge (PMRA# 2842307, 2842433)	Agriculture, developed land, low disturbance	2017	0.0027	4	1	25	0.002	0.002	0.001	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red Deer River at Nevis Bridge (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	0.0027	4	1	25	0.003	0.003	0.001	0.008	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red Deer River at Morrin Bridge (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	0.0027	4	1	25	0.005	0.007	0.001	0.014	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Red Deer River downstream of Dinosaur Provincial Park	Low disturbance, agriculture	2017	0.0027	4	1	25	0.005	0.006	0.001	0.014	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceed	ng the toxicity	endpoints
(Data source)	,		(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
(PMRA# 2842307, 2842433)														
North Saskatchewan River (3 sites) (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	0.0027	11	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Battle River downstream of Highway 53 (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Battle River at the North end of Dried Meat Lake (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	0.0027	4	1	25	0.002	0.001	0.001	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Beaver River (3 sites) (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	0.0027	12	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Athabasca River (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Peace River (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Wapiti River (2 sites) (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	0.0027	8	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Smoky River (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Milk River (PMRA# 2842307, 2842433)	Low disturbance, agriculture	2017	0.0027	4	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Bigknife Creek (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	0.0027	1	1	100	0.045	NA	0.045	0.045	1 (100%)	0 (0%)	0 (0%)	0 (0%)
Birch Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Buffalo Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceed	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(μg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Beaverhill Creek (PMRA# 2842307, 2842433)	Canola, cereals, mixed animal use, low disturbance	2017	0.0027	3	3	100	0.019	0.008	0.017	0.028	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Big Valley Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Egg Creek (PMRA# 2842307, 2842433)	Canola, cereals, pulse crops, mixed animal use, low disturbance, developed land	2017	0.0027	3	3	100	0.009	0.008	0.006	0.018	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Grizzlybear Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Haynes Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance	2017	0.0027	3	2	67	0.006	0.004	0.008	0.008	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Kneehills Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Michichi Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mosquito Creek (PMRA# 2842307, 2842433)	Canola, cereals, pulse crops, mixed animal use, low disturbance	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Meeting Creek (PMRA# 2842307, 2842433)	Cereals, mixed animal use, low disturbance	2017	0.0027	2	1	50	0.003	0.003	0.003	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceed	ng the toxicity	endpoints
(Data source)	•		(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
Seven Persons Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Pipestone Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance, developed land	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Parlby Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, unknown agricultural use, low disturbance	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Pothole Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance	2017	0.0027	1	1	100	0.005	NA	0.005	0.005	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Queenie Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	3	1	33	0.002	0.001	0.001	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Ray Creek (PMRA# 2842307, 2842433)	Canola, cereals, mixed animal use, low disturbance	2017	0.0027	3	1	33	0.004	0.005	0.001	0.011	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Ribstone Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	3	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Redwillow Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Rosebud Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	2	1	50	0.005	0.005	0.005	0.009	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Scandia Creek	Cereals, mixed	2017	0.0027	1	0	0	0.001	NA	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	nples) exceed	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 µg/L
(PMRA# 2842307, 2842433)	animal use, low disturbance													
Sturgeon River (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	3	2	67	0.006	0.008	0.002	0.015	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Serviceberry Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	2	0	0	0.001	0	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Threehills Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	3	2	67	0.01	0.008	0.012	0.018	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Vermilion River (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	3	3	100	0.02	0.021	0.009	0.043	1 (33%)	0 (0%)	0 (0%)	0 (0%)
Weiller Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	0.0027	2	1	50	0.009	0.011	0.009	0.016	0 (0%)	0 (0%)	0 (0%)	0 (0%)
West Michichi Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	0.0027	2	1	50	0.004	0.004	0.004	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Yellow Lake Tributary (PMRA# 2842307, 2842433)	Cereals, sugar beet, pulse crops, potatoes, mixed animal use, low disturbance	2017	0.0027	1	0	0	0.001	NA	0.001	0.001	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Surface water from 23 watersheds (PMRA# 2523835)	Agriculture	2004, 2006	0.05	245	0	0	0.025	0	0.025	0.025	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Surface water (PMRA# 2523834)	Agriculture	2004- 2013	0.05	2577	1	0	0.025	0	0.025	0.025	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Seasonal (Class III) and semi-permanent (Class	Wheat, canola, oats, barley,	Summe r 2017	0.002	18	4	22	NC	NC	NC	Overall range:	2 wetlands (11%)	0 wetlands (0%)	0 wetlands (0%)	0 wetlands

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of san	iples) exceedi	ing the toxicity	endpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 µg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
IV) wetlands ^{1,2} (PMRA# 2847073, 2847083)	pasture									0.001 – 0.19; Range of detects: 0.004 – 0.19				(0%)
		Fall 2017	0.002	10	0	0	NC	NC	NC	0.001	0 wetlands (0%)	0 wetlands (0%)	0 wetlands (0%)	0 wetlands (0%)
50 irrigation water sites ³ (PMRA# 2842307, 2842433)	Agriculture	2017	0.0027	194	3	2	0.002^3	0.001 ³	0.001 ³	0.016 ³	0 (0%) ³	0 (0%) ³	0 (0%) ³	0 (0%)3
3 tile drain sites ³ (PMRA# 2842307, 2842433)	Irrigated agricultural area	2017	0.0027	8	4	50	0.0233	0.0323	0.0043	0.089^3	3 (38%) ³	0 (0%)3	0 (0%)3	0 (0%)3

LOD = limit of detection; N = sample size; Stdev = standard deviation; Chronic HC_5 = the 5^{th} percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC_{10} = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC_5 = the 5^{th} percentile of the species sensitivity distribution for the LC_{50} (the median lethal concentration) at 50% confidence intervals; LOQ = limit of quantification; NA = not applicable; NC = not calculated

¹The wetlands were classified by the researchers using the classification system defined in Stewart, R.E. and H.A. Kantrud. 1971. Classification of natural ponds and lakes in the glaciated prairie region. Bureau of Sport Fisheries and Wildlife, U.S. Fish and Wildlife Service, Washington, D.C., USA. Resource Publication 92. 57 pp.

²Each wetland in these data sets was sampled only once during the time period, with the following exceptions:

a) For summer 2013 in the data set from PMRA# 2526133 and 2612760, 11 wetlands in canola-growing areas were sampled three times between the months of June and July 2013. The average of the three values was used in calculations for each of the wetlands to represent concentrations for the sampling period.

b) For spring 2014 in the data set from PMRA# 2572395, 2612761, 16 wetlands were sampled three to five times between May and June 2014. The averages over the four-week period were used in calculations for each of the wetlands to represent concentrations for the sampling period.

Average, standard deviation and median concentrations to estimate chronic exposure concentrations were not calculated because most wetlands were sampled only once during each time period.

³Irrigation water and tile drain sites may not represent aquatic habitat.

Table A.7-8 Risk quotients for thiamethoxam measured in waterbodies located in Manitoba, Saskatchewan and Alberta.

NOTES:

- -Shaded cells indicate the level of concern is exceeded, meaning that the risk quotient is equal to or greater than a value of 1.
- -The frequency of sampling and the length of the sampling period varied between data sets. Some waterbodies were sampled one to three times between May and October, while others were sampled one to three times per month between April and December. Values measured at sites where only a few samples were collected may not represent concentrations throughout the growing season.

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotie ng average ² conce onic toxicity endpo	entrations and	calculated using	nic Risk Quotient g median ² concentic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
					Manitob)ង				
Red River at Emerson	Soybeans, wheat,	2014	7	0.7	< 0.1	0.1	0.6	< 0.1	< 0.1	< 0.1
(PMRA# 2745819)	canola, oats, corn	2015	6	0.6	< 0.1	< 0.1	0.7	< 0.1	0.1	< 0.1
		2016	1	0.4	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
Red River at Emerson (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats, corn	2017	3	0.8	< 0.1	0.1	0.5	< 0.1	< 0.1	< 0.1
Red River at Selkirk (PMRA# 2745819)	Soybeans, wheat, canola, oats	2014	1	0.5	< 0.1	< 0.1	0.5	< 0.1	< 0.1	< 0.1
Red River at Selkirk (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats	2017	3	0.7	< 0.1	0.1	0.3	< 0.1	< 0.1	< 0.1
Red River at Norbert (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	3	0.7	< 0.1	0.1	0.4	< 0.1	< 0.1	< 0.1
Assiniboine River Northwest of Treesbank (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, corn	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Assiniboine River at Happy Hollow Farm (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, corn	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotie ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Assiniboine River downstream of Portage la Prairie (PMRA# 2849359, 2849370)	Soybeans, wheat, canola	2017	3	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Assiniboine River at Headingley (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, oats, barley, corn	2017	3	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Assiniboine River at Provincial Trunk Highway 21, North of Griswold (PMRA# 2849359, 2849370)	Canola, wheat, soy beans, barley	2017	3	0.5	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
Assiniboine River at Provincial Trunk Highway 83, South of Miniota (PMRA# 2849359, 2849370)	Canola, wheat, soy beans, barley	2017	3	2	0.1	0.2	0.4	< 0.1	< 0.1	< 0.1
Boyne River (PMRA# 2849359, 2849370)	Soybeans, corn, wheat, canola, oats	2017	3	1.6	0.1	0.1	0.9	0.1	0.1	< 0.1
Cooks Creek at Rural Municipality Boundary Road (PMRA# 2849359, 2849370)	Soybeans, canola, oats, corn, wheat	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Cooks Creek south of Millbrook (PMRA# 2849359, 2849370)	Soybeans, canola, oats, corn, wheat	2017	3	1.9	0.1	0.2	0.2	< 0.1	< 0.1	< 0.1
Edwards Creek (PMRA# 2849359, 2849370)	Canola, soybeans, wheat	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	conic Risk Quotie ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Icelandic River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats	2017	3	0.7	< 0.1	0.1	0.1	< 0.1	< 0.1	< 0.1
La Salle River at the town of La Salle (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, oats, corn	2017	3	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
La Salle River at La Barriere (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	3	0.3	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Lake Manitoba (PMRA# 2849359, 2849370)	Soybeans, wheat, canola	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Lake Winnipeg (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats	2017	3	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
Little Saskatchewan River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, barley	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Morris River (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, corn, oats	2017	3	1.5	0.1	0.1	1.3	0.1	0.1	< 0.1
Oak River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, barley	2017	3	0.7	< 0.1	0.1	0.2	< 0.1	< 0.1	< 0.1
Pelican Lake (PMRA# 2849359, 2849370)	Wheat, canola, soybeans	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Pipestone Creek (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, barley	2017	3	0.3	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Rat River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	Chronic Risk Quotients ¹ calculated using average ² concentrations and chronic toxicity endpoints			Chronic Risk Quotients ¹ calculated using median ² concentrations and chronic toxicity endpoints			Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Rock Lake (PMRA# 2849359, 2849370)	Canola, soybeans, wheat, barley	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Roseau River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, oats, corn	2017	3	0.3	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Seine River (PMRA# 2849359, 2849370)	Soybeans, wheat, canola, corn, oats	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Souris River at the Town of Souris (PMRA# 2849359, 2849370)	Canola, soybeans, wheat, corn	2017	2	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Souris River at Melita (PMRA# 2849359, 2849370)	Canola, wheat, soybeans, oats	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Sturgeon Creek (PMRA# 2849359, 2849370)	Soybeans, canola, wheat, oats, barley, corn	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Swan River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans	2017	3	0.3	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Willow Creek (PMRA# 2849359, 2849370)	Soybeans, wheat, canola	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Woody River (PMRA# 2849359, 2849370)	Canola, wheat, soybeans	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Seasonal (Class III) and semi-permanent (Class IV) wetlands ^{4,5} (PMRA# 2847073, 2847083)	Canola, wheat, oats, pasture, corn	Summer 2017	12	Using range of concentrations: < 0.1 – 29	Using range of concentrations: < 0.1 – 1.8	Using range of concentrations: < 0.1 – 2.5	Using range of concentrations: < 0.1 – 29	Using range of concentrations: < 0.1 – 1.8	Using range of concentrations: < 0.1 - 2.5	Using range of concentrations: < 0.1 – 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotiong average ² concomic toxicity endp	entrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
		Fall 2017	5	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrati ons:	Using range of concentrations: < 0.1
Creek (PMRA# 2548877)	Agriculture	2013	1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Streams, culverts, ditches (PMRA# 2548876)	Agriculture	2014	3	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
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Assiniboine River	Canola and	2014	6	0.7	< 0.1	0.1	0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2745819)	rapeseed, wheat	2015	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Battle River	Canola and	2015	6	0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2745819)	rapeseed, rye, wheat	2016	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Avonlea Creek (PMRA# 2849265, 2849266)	Canola, peas, lentils, wheat	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Lanigan Creek (PMRA# 2849265, 2849266)	Mainly canola, with some peas and wheat	2017	8	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Lightning Creek (PMRA# 2849265, 2849266)	Canola with some soybeans and wheat	2017	10	0.4	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
McDonald Creek (PMRA# 2849265, 2849266)	Mainly lentils, with wheat	2017	7	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Moose Jaw River (PMRA# 2849265, 2849266)	Lentils, canola, with wheat	2017	9	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Moose Mountain Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	9	0.8	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotiong average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Oscar Creek (PMRA# 2849265, 2849266)	Mainly canola	2017	10	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Pipestone Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	12	0.3	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Saline Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	10	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Souris River (PMRA# 2849265, 2849266)	Mainly canola, lentils, with wheat	2017	9	0.4	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Spirit Creek (PMRA# 2849265, 2849266)	Mainly canola, with wheat	2017	10	0.7	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
Swift Current Creek below Rock Creek (PMRA# 2849265, 2849266)	Mainly lentils, with some peas, canola and wheat	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Swift Current Creek near Leinon (PMRA# 2849265, 2849266)	Mainly lentils, with some peas, canola and wheat	2017	9	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Willowbrook Creek (PMRA# 2849265, 2849266)	Mainly canola	2017	9	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Wood River (PMRA# 2849265, 2849266)	Mainly lentils, peas, with wheat	2017	9	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Temporary (Class II), seasonal (Class III), semi-permanent (Class IV) and permanent (Class V) wetlands ^{4,5}	Barley, canola, oats, wheat, grassland (previous year's crops)	Spring (pre-seed) 2012	138	Using range of concentrations: < 0.1 – 1.2	Using range of concentrations: < 0.1 – 0.1	Using range of concentrations: < 0.1 – 0.1	Using range of concentrations: < 0.1 – 1.2	Using range of concentrations: < 0.1 – 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotion ong average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient g median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
(PMRA# 2526133, 2572395, 2608629, 2612760, 2612761, 2612762, 2712896)	Barley, canola, oats, wheat, peas, grassland	Summer 2012	134	Using range of concentrations: < 0.1 – 57	Using range of concentrations: < 0.1 – 3.5	Using range of concentrations: < 0.1 – 5	Using range of concentrations: < 0.1 – 57	Using range of concentrations: < 0.1 – 3.5	Using range of concentrations: < 0.1 -5	Using range of concentrations: < 0.1 – 0.2
	Barley, canola, oats, wheat, peas, grassland	Fall 2012	80	concentrations: $< 0.1 - 3.8$ concentrations: $< 0.1 - 0.2$		Using range of concentrations: < 0.1 – 0.3	Using range of concentrations: < 0.1 – 3.8	Using range of concentrations: < 0.1 – 0.2	Using range of concentrations: < 0.1 - 0.3	Using range of concentrations: < 0.1
	Barley, canola, oats, wheat, peas, grassland (previous year's crops)	Spring (pre-seed) 2013	90	Using range of concentrations: < 0.1 – 4.1	Using range of concentrations: < 0.1 – 0.3	Using range of concentrations: < 0.1 – 0.4	Using range of concentrations: < 0.1 – 4.1	Using range of concentrations: < 0.1 – 0.3	Using range of concentrations: < 0.1 - 0.4	Using range of concentrations:
	Barley, canola, oats, peas, wheat, flax, grassland, chemfallow	Summer 2013	144	Using range of concentrations: < 0.1 – 18	Using range of concentrations: < 0.1 – 1.1	Using range of concentrations: < 0.1 – 1.6	Using range of concentrations: < 0.1 – 18	Using range of concentrations: < 0.1 – 1.1	Using range of concentrations: < 0.1 - 1.6	Using range of concentrations: < 0.1
	Canola, oats (previous year's crops)	Spring (pre-seed) 2014	16	Using range of concentrations: < 0.1 – 3.2	Using range of concentrations: < 0.1 – 0.2	Using range of concentrations: < 0.1 – 0.3	Using range of concentrations: < 0.1 – 3.2	Using range of concentrations: < 0.1 – 0.2	Using range of concentrati ons: < 0.1 - 0.3	Using range of concentrations: < 0.1
 	Barley, canola,	Summer	All we			•		*		
	flax, oats, lentils, wheat, peas, soybeans, chemfallow, pasture, grassland	2014	115	Using range of concentrations: < 0.1 – 33	Using range of concentrations: < 0.1 - 2	Using range of concentrations: < 0.1 – 2.9	Using range of concentrations: < 0.1 – 33	Using range of concentrations: < 0.1 - 2	Using range of concentrations: < 0.1 - 2.9	Using range of concentrations: < 0.1 – 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotiong average ² conconnic toxicity endp	entrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
			46	Using range of concentrations: < 0.1 – 17	Using range of concentrations: < 0.1 – 1.1	Using range of concentrations: < 0.1 – 1.5	Using range of concentrations: < 0.1 – 17	Using range of concentrations: < 0.1 – 1.1	Using range of concentrations: < 0.1 -1.5	Using range of concentrations: < 0.1 – 0.1
Seasonal (Class III) and semi-permanent (Class IV) wetlands ^{4,5} (PMRA# 2847073, 2847083)	Wheat, canola, barley, pasture, lentils, summer fallow	Summer 2017	30	Using range of concentrations: < 0.1 – 12	Using range of concentrations: < 0.1 – 0.7	Using range of concentrations: < 0.1 - 1	Using range of concentrations: < 0.1 – 12	Using range of concentrations: < 0.1 – 0.7	Using range of concentrations:	Using range of concentrations: < 0.1
		Fall 2017	8	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1
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South Saskatchewan River (PMRA# 2745819)	Grassland, peas, wheat	2014	5	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Oldman River (3 sites) (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	12	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
South Saskatchewan River (PMRA# 2842307, 2842433)	Low disturbance, developed land, agriculture	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Bow River (4 sites) (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	16	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Elbow River (PMRA# 2842307, 2842433)	Developed land, low disturbance	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Red Deer River (PMRA# 2745819)	Grassland, peas, wheat, canola and rapeseed	2015	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotie ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Red Deer River at Sundre (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Red Deer River 1 kilometre upstream of Highway 2 Bridge (PMRA# 2842307, 2842433)	Agriculture, developed land, low disturbance	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Red Deer River at Nevis Bridge (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Red Deer River at Morrin Bridge (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	4	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Red Deer River downstream of Dinosaur Provincial Park (PMRA# 2842307, 2842433)	Low disturbance, agriculture	2017	4	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
North Saskatchewan River (3 sites) (PMRA# 2842307, 2842433)	Low disturbance, agriculture, developed land	2017	11	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Battle River downstream of Highway 53 (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Battle River at the North end of Dried Meat Lake (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotie ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient median ² concen c toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Beaver River (3 sites) (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	12	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Athabasca River (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Peace River (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Wapiti River (2 sites) (PMRA# 2842307, 2842433)	Agriculture, low disturbance, developed land	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Smoky River (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Milk River (PMRA# 2842307, 2842433)	Low disturbance, agriculture	2017	4	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Bigknife Creek (PMRA# 2842307, 2842433)	Agriculture, low disturbance	2017	1	1.7	0.1	0.1	1.7	0.1	0.1	< 0.1
Birch Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Buffalo Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Beaverhill Creek (PMRA# 2842307, 2842433)	Canola, cereals, mixed animal use, low disturbance	2017	3	0.7	< 0.1	0.1	0.7	< 0.1	0.1	< 0.1
Big Valley Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotie ing average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Egg Creek (PMRA# 2842307, 2842433)	Canola, cereals, pulse crops, mixed animal use, low disturbance, developed land	2017	3	0.4	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Grizzlybear Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Haynes Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance	2017	3	0.2	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
Kneehills Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Michichi Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Mosquito Creek (PMRA# 2842307, 2842433)	Canola, cereals, pulse crops, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Meeting Creek (PMRA# 2842307, 2842433)	Cereals, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Seven Persons Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Pipestone Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance, developed land	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotie ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient gmedian ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Parlby Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, unknown agricultural use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Pothole Creek (PMRA# 2842307, 2842433)	Cereals, canola, pulse crops, mixed animal use, low disturbance	2017	1	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Queenie Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Ray Creek (PMRA# 2842307, 2842433)	Canola, cereals, mixed animal use, low disturbance	2017	3	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Ribstone Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	3	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Redwillow Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Rosebud Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	2	0.2	< 0.1	< 0.1	0.2	< 0.1	< 0.1	< 0.1
Scandia Creek (PMRA# 2842307, 2842433)	Cereals, mixed animal use, low disturbance	2017	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Sturgeon River (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	3	0.2	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Serviceberry Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotion ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotien g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Threehills Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	3	0.4	< 0.1	< 0.1	0.5	< 0.1	< 0.1	< 0.1
Vermilion River (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	3	0.8	< 0.1	< 0.1	0.4	< 0.1	< 0.1	< 0.1
Weiller Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance, developed land	2017	2	0.3	< 0.1	< 0.1	0.3	< 0.1	< 0.1	< 0.1
West Michichi Creek (PMRA# 2842307, 2842433)	Cereals, canola, mixed animal use, low disturbance	2017	2	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Yellow Lake Tributary (PMRA# 2842307, 2842433)	Cereals, sugar beet, pulse crops, potatoes, mixed animal use, low disturbance	2017	1	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Surface water from 23 watersheds (PMRA# 2523835)	Agriculture	2004, 2006	245	1	0.1	0.1	1	0.1	0.1	< 0.1
Surface water (PMRA# 2523834)	Agriculture	2004- 2013	2577	1	0.1	0.1	1	0.1	0.1	< 0.1
Seasonal (Class III) and semi-permanent (Class IV) wetlands ^{4,5} (PMRA# 2847073,	Wheat, canola, oats, barley, pasture	Summer 2017	18	Using range of concentrations: 7.3	Using range of concentrations: 0.4	Using range of concentrations: 0.6	Using range of concentrations: 7.3	Using range of concentrations: 0.4	Using range of concentrati ons: 0.6	Using range of concentrations: < 0.1
2847083)		Fall 2017	10	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1	Using range of concentrations: < 0.1
50 irrigation water sites ⁶ (PMRA# 2842307, 2842433)	Agriculture	2017	194	0.16	< 0.16	< 0.16	0.16	< 0.16	< 0.16	< 0.16

Waterbody (Data source)	Major land use	Year	N	calculated usi	ronic Risk Quotic ng average ² conc onic toxicity endp	entrations and	calculated using	nic Risk Quotient median ² concent c toxicity endpoi	trations and	maximum ^{2,3} concentrations and the acute toxicity endpoint		
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.45 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)		
3 tile drain sites ⁶ (PMRA# 2842307, 2842433)	Irrigated agricultural area	2017	8	0.9 ⁶	0.16	0.16	0.16	< 0.16	< 0.16	< 0.1 ⁶		

N = sample size; Chronic HC₅ = the 5th percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC₁₀ = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC₅ = the 5th percentile of the species sensitivity distribution for the LC₅₀ (the median lethal concentration) at 50% confidence intervals

Average, standard deviation and median concentrations to estimate chronic exposure concentrations were not calculated because most wetlands were sampled only once during each time period. Risk quotients were calculated using the single highest concentration, in the absence of a chronic exposure level.

⁶Irrigation water and tile drain sites may not represent aquatic habitat.

¹Risk Quotient = concentration ÷ toxicity endpoint

²Average, median and maximum concentrations over the sampling period are reported in Table A.7-7.

³Because monitoring may not capture peak concentrations, maximum concentrations may be underestimated.

⁴The wetlands were classified by the researchers using the classification system defined in Stewart, R.E. and H.A. Kantrud. 1971. Classification of natural ponds and lakes in the glaciated prairie region. Bureau of Sport Fisheries and Wildlife, U.S. Fish and Wildlife Service, Washington, D.C., USA. Resource Publication 92. 57 pp.

⁵Each wetland in these data sets was sampled only once during the time period, with the following exceptions:

a) For summer 2013 in the data set from PMRA# 2526133 and 2612760, 11 wetlands in canola-growing areas were sampled three times between the months of June and July 2013. The average of the three values was used in calculations for each of the wetlands.

b) For spring 2014 in the data set from PMRA# 2572395, 2612761, 16 wetlands were sampled three to five times between May and June 2014. The averages over the four-week period were used in calculations for each of the wetlands.

Table A.7-9 Summary statistics for thiamethoxam measured in waterbodies from British Columbia.

NOTES:

- -In calculations, the PMRA assigned a value equal to half the limit of detection to samples that showed no detection.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred one to three times per month between May and December. Sampling at some sites occurred only a few times over a short time period, and values measured may not represent concentrations throughout the growing season.

Waterbody	Major land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of sampl	es) exceeding	the toxicity end	lpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
Alouette River	Urban, corn,	2014	0.00139	7	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2707947)	berries	2015	0.00139	9	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Chilliwack River (PMRA# 2707947)	Urban, forest	2015	0.00139	9	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Coquitlam River (PMRA# 2707947)	Urban, forest	2014	0.00139	7	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Fishtrap Creek	Berries, corn,	2014	0.00139	7	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2707947)	greenhouses	2015	0.00139	8	3	38	0.001	0.0004	0.0007	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Harrison River (PMRA# 2707947)	Agriculture	2015	0.00139	9	1	11	0.001	0.0003	0.0007	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Hope Slough	Urban, forest,	2014	0.00139	7	2	29	0.001	0.0005	0.0007	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
(PMRA# 2707947)	corn	2015	0.00139	8	7	88	0.79	1.9	0.088	5.5	5 (63%)	1 (13%)	2 (25%)	0 (0%)
Murdo Creek (PMRA# 2707947)	Forest	2014	0.00139	7	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Okanagan River (PMRA# 2707947)	Orchards, vineyards, vegetables, fruit	2015	0.00139	2	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Okanagan River, upstream (PMRA# 2842180)	Urban, forest, corn, blueberries	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	9/0	Average	Stdev	Median	Max	N (% of samp	les) exceeding	the toxicity end	lpoints
(Data source)	·		(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Okanagan River; downstream (PMRA# 2842180)	Fruit trees, grapes	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Vedder Canal (PMRA# 2707947)	Urban, forest, agriculture	2015	0.00139	9	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Gold Creek (PMRA# 2889992)	No agriculture in the watershed	2016	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Katzie Slough (PMRA# 2889992)	Berries, grass, forage, ornamentals and shrubs	2016	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Matsqui Slough (PMRA# 2889992)	Berries, grass, forage, corn, nurseries	2016	0.00139	5	3	60	0.002	0.002	0.002	0.006	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Scott Creek (PMRA# 2889992)	Residential, golf course	2016	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Sumas Drainage Canal (PMRA# 2889992)	Potatoes, vegetables, forage crops (corn or peas), berries, turf, sweet corn, cereals, oilseed and fallow, floriculture, nurseries	2016	0.00139	5	3	60	0.001	0.0004	0.001	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Sumas Lake Canal; upstream (PMRA# 2842180)	Urban, forest, corn, blueberries, potatoes, vegetables	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Sumas Lake Canal; downstream (PMRA# 2842180)	Potatoes, vegetables, corn, berries, cereals, oilseeds	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N N	9/0	Average	Stdev	Median	Max	N (% of samp	les) exceeding	the toxicity end	lpoints
(Data source)	•		(µg/L)		detects	Detection	(μg/L)		(µg/L)	(μg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 µg/L	Acute HC ₅ of 9 μg/L
Sumas River at the Border (PMRA# 2889992)	River flows into Canada from the United States	2016	0.00139	5	0	0	0.0007	0	0.0007	0.0007	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Cohilukthan Slough (PMRA# 2842180)	Potatoes, vegetables, berries, cereals, oilseeds, corn	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Middle Vernon Creek; upstream (PMRA# 2842180)	Urban, wheat, orchards	2017	0.004	8	1	13	0.002	0.001	0.002	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Middle Vernon Creek; downstream (PMRA# 2842180)	Fruit trees, berries, grapes, potatoes, vegetables	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mission Creek; upstream (PMRA# 2842180)	Urban, forest, wheat, orchards	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Mission Creek; downstream (PMRA# 2842180)	Fruit trees, grapes	2017	0.004	8	1	13	0.002	0.001	0.002	0.004	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Naramata Creek; upstream (PMRA# 2842180)	Urban, forest, orchards, vineyards	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Naramata Creek; downstream (PMRA# 2842180)	Grapes, fruit trees	2017	0.004	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Nicomekl River; upstream (PMRA# 2842180)	Berries, nurseries and ornamentals	2017	0.004	8	1	13	0.025	0.065	0.002	0.19	1 (13%)	0 (0%)	0 (0%)	0 (0%)
Nicomekl River; downstream (PMRA# 2842180)	Berries, potatoes, vegetables, corn	2017	0.004	8	1	13	0.003	0.003	0.002	0.01	0 (0%)	0 (0%)	0 (0%)	0 (0%)

Waterbody	Major land use	Year	LOD	N	N	%	Average	Stdev	Median	Max	N (% of samp	les) exceeding	the toxicity end	lpoints
(Data source)			(µg/L)		detects	Detection	(µg/L)		(µg/L)	(µg/L)	Chronic HC ₅ of 0.026 μg/L	Chronic EC ₁₀ of 0.43 µg/L	Mesocosm NOEC of 0.3 μg/L	Acute HC ₅ of 9 μg/L
Trout Creek; upstream (PMRA# 2842180)	Wheat, forest, shrubland	2017	0.005	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Trout Creek; downstream (PMRA# 2842180)	Fruit trees, grapes, potatoes, vegetables	2017	0.005	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Flowing waterbody with no pesticide application (PMRA# 2842180)	No crops	2017	0.005	8	0	0	0.002	0	0.002	0.002	0 (0%)	0 (0%)	0 (0%)	0 (0%)
Slough, water at the edge of a field (PMRA# 2548876)	Agriculture	2014	0.0008	2	1	50	0.035	0.049	0.035	0.069	1 (50%)	0 (0%)	0 (0%)	0 (0%)

LOD = limit of detection; N = sample size; Stdev = standard deviation; Chronic HC_5 = the 5^{th} percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals; EC_{10} = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC_5 = the 5^{th} percentile of the species sensitivity distribution for the LC_{50} (the median lethal concentration) at 50% confidence intervals

Table A.7-10 Risk quotients for thiamethoxam measured in waterbodies located in British Columbia.

NOTE:

- -Shaded cells indicate the level of concern is exceeded, meaning that the risk quotient is equal to or greater than a value of 1.
- -The frequency of sampling and the length of the sampling period varied between data sets. Sampling generally occurred one to three times per month between May and December. Sampling at some sites occurred only a few times over a short time period, and values measured may not represent concentrations throughout the growing season.

Waterbody	Land use	Year	N	calculated usi	ronic Risk Quotien ing average ² concel onic toxicity endpo	ntrations and	calculated using	nic Risk Quotient g median ² concent ic toxicity endpoin	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Alouette River	Urban, corn,	2014	7	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2707947)	berries	2015	9	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Chilliwack River (PMRA# 2707947)	Urban, forest	2015	9	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Coquitlam River (PMRA# 2707947)	Urban, forest	2014	7	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Fishtrap Creek	Berries, corn,	2014	7	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2707947)	greenhouses	2015	8	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Harrison River (PMRA# 2707947)	Agriculture	2015	9	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Hope Slough	Urban, forest,	2014	7	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
(PMRA# 2707947)	corn	2015	8	30	1.8	2.6	3.4	0.2	0.3	0.6
Murdo Creek (PMRA# 2707947)	Forest	2014	7	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Okanagan River (PMRA# 2707947)	Orchards, vineyards, vegetables, fruit	2015	2	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Okanagan River; upstream (PMRA# 2842180)	Urban, forest, corn, blueberries	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Okanagan River, downstream (PMRA# 2842180)	Fruit trees, grapes	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Vedder Canal (PMRA# 2707947)	Urban, forest, agriculture	2015	9	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1

Waterbody	dy Land use	Year	N	calculated us	ronic Risk Quotien ing average ² concer onic toxicity endpo	ntrations and	calculated using	nic Risk Quotient g median ² concen ic toxicity endpoi	trations and	Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
					Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)
Gold Creek (PMRA# 2889992)	No agriculture in the watershed	2016	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Katzie Slough (PMRA# 2889992)	Berries, grass, forage, ornamentals and shrubs	2016	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Matsqui Slough (PMRA# 2889992)	Berries, grass, forage, corn, nurseries	2016	5	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Scott Creek (PMRA# 2889992)	Residential, golf course	2016	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Sumas Drainage Canal (PMRA# 2889992)	Potatoes, vegetables, forage crops (corn or peas), berries, turf, sweet corn, cereals, oilseed and fallow, floriculture, nurseries	2016	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Sumas Lake Canal; upstream (PMRA# 2842180)	Urban, forest, corn, blueberries, potatoes, vegetables	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Sumas Lake Canal; downstream (PMRA# 2842180)	Potatoes, vegetables, corn, berries, cereals, oilseeds	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1
Sumas River at the Border (PMRA# 2889992)	River flows into Canada from the United States	2016	5	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1
Cohilukthan Slough (PMRA# 2842180)	Potatoes, vegetables, berries, cereals, oilseeds, corn	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1

Waterbody	Land use	Year	N	calculated usi	ronic Risk Quotier ng average ² conce onic toxicity endpo	ntrations and	Chronic Risk Quotients ¹ calculated using median ² concentrations and chronic toxicity endpoints			Acute Risk Quotients ¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint			
							Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 µg/L)	Acute HC ₅ (9 μg/L)
Middle Vernon Creek; upstream (PMRA# 2842180)	Urban, wheat, orchards	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Middle Vernon Creek; downstream (PMRA# 2842180)	Fruit trees, berries, grapes, potatoes, vegetables	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Mission Creek; upstream (PMRA# 2842180)	Urban, forest, wheat, orchards	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Mission Creek; downstream (PMRA# 2842180)	Fruit trees, grapes	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Naramata Creek; upstream (PMRA# 2842180)	Urban, forest, orchards, vineyards	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Naramata Creek; downstream (PMRA# 2842180)	Grapes, fruit trees	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Nicomekl River; upstream (PMRA# 2842180)	Berries, nurseries and ornamentals	2017	8	1	< 0.1	0.1	0.1	< 0.1	< 0.1	< 0.1			
Nicomekl River; downstream (PMRA# 2842180)	Berries, potatoes, vegetables, corn	2017	8	0.1	< 0.1	0.1	0.1	< 0.1	< 0.1	< 0.1			
Trout Creek; upstream (PMRA# 2842180)	Wheat, forest, shrubland	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Trout Creek; downstream (PMRA# 2842180)	Fruit trees, grapes, potatoes, vegetables	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			
Flowing waterbody with no pesticide application (PMRA# 2842180)	No crops	2017	8	0.1	< 0.1	< 0.1	0.1	< 0.1	< 0.1	< 0.1			

Waterbody	Land use	Year	N	calculated us	ronic Risk Quotier ing average ² conce onic toxicity endpo	ntrations and	calculated using	nic Risk Quotient g median ² concent ic toxicity endpoir	trations and	Acute Risk Quotients¹ calculated using maximum ^{2,3} concentrations and the acute toxicity endpoint
				Chronic HC ₅ (0.026 µg/L)	Chronic EC ₁₀ (0.43 μg/L)	Mesocosm NOEC (0.3 μg/L)	Chronic HC ₅ (0.026 μg/L)	Chronic EC ₁₀ (0.43 µg/L)	Mesocosm NOEC (0.3 μg/L)	Acute HC ₅ (9 μg/L)
Slough, water at the edge of a field (PMRA# 2548876)	Agriculture	2014	2	1.3	0.1	0.1	1.3	0.1	0.1	< 0.1

N = sample size, Chronic HC_5 = the 5th percentile of the species sensitivity distribution for the NOEC at 50% confidence intervals, EC_{10} = effective concentration on 10% of the population (it is the most sensitive single species chronic endpoint for thiamethoxam); NOEC = no observable effect concentration; Acute HC_5 = the 5th percentile of the species sensitivity distribution for the LC_{50} (the median lethal concentration) at 50% confidence intervals

¹Risk Quotient = concentration ÷ toxicity endpoint

²Average, median and maximum concentrations over the sampling period are reported in Table A.7-9.

³Because monitoring may not capture peak concentrations, maximum concentrations may be underestimated.

Appendix VIII Proposed Label Amendments for Products Containing Thiamethoxam

The label amendments proposed below do not include all label requirements for individual products, such as disposal statements, and precautionary statements. Information on labels of currently registered products should not be removed unless it contradicts the following label statements.

Add to ENVIRONMENTAL PRECAUTIONS:

TOXIC to aquatic organisms and non-target terrestrial plants. Observe buffer zones specified under DIRECTIONS FOR USE.

Toxic to non-target terrestrial plants. This product contains an active ingredient and aromatic petroleum distillates that are TOXIC to aquatic organisms. Observe buffer zones specified under DIRECTIONS FOR USE. [for PCP# 30404]

To reduce runoff from treated areas into aquatic habitats avoid application to areas with a moderate to steep slope, compacted soil, or clay.

Avoid application when heavy rain is forecast.

Contamination of aquatic areas as a result of runoff may be reduced by including a vegetative strip between the treated area and the edge of the water body.

Add to DIRECTIONS FOR USE:

As this product is not registered for the control of pests in aquatic systems, DO NOT use to control aquatic pests.

DO NOT contaminate irrigation or drinking water supplies or aquatic habitats by cleaning of equipment or disposal of wastes.

DO NOT allow effluent or runoff from greenhouses or mushroom houses containing this product to enter lakes, streams, ponds or other waters. [for PCP#s 30723 and 30901]

<u>Field sprayer application</u>: **DO NOT** apply during periods of dead calm. Avoid application of this product when winds are gusty. **DO NOT** apply with spray droplets smaller than the American Society of Agricultural Engineers (ASAE S572.1) fine classification. Boom height must be 60 cm or less above the crop or ground.

<u>Airblast application</u>: **DO NOT** apply during periods of dead calm. Avoid application of this product when winds are gusty. **DO NOT** direct spray above plants to be treated. Turn off outward pointing nozzles at row ends and outer rows. **DO NOT** apply when wind speed is greater than 16 km/h at the application site as measured outside of the treatment area on the upwind side.

Aerial application: **DO NOT** apply during periods of dead calm. Avoid application of

this product when winds are gusty. **DO NOT** apply when wind speed is greater than 16 km/h at flying height at the site of application. **DO NOT** apply with spray droplets smaller than the American Society of Agricultural Engineers (ASAE S572.1) fine classification. Reduce drift caused by turbulent wingtip vortices. Nozzle distribution along the spray boom length MUST NOT exceed 65% of the wing- or rotorspan.

Buffer zones:

Spot treatments using hand-held equipment **DO NOT** require a buffer zone. In-furrow application and soil drench or soil incorporation **DO NOT** require a buffer zone.

The buffer zones specified in the table below are required between the point of direct application and the closest downwind edge of sensitive terrestrial habitats (such as grasslands, forested areas, shelter belts, woodlots, hedgerows, riparian areas and shrublands), and sensitive freshwater habitats (such as lakes, rivers, sloughs, ponds, prairie potholes, creeks, marshes, streams, reservoirs and wetlands).

			Buffer Zones (metres) Required for the Protection of:						
Method of	Crop		Freshwater H	labitat of Depths:					
application			Less than 1 m	Greater than 1 m	Terrestrial Habitat:				
	Potato (foliar), Crop group 11 vegetables, Crop group 4 Lea		4	2	1				
	Soybean, bean (dry), outdoor landscapes, Viburnum	nurseries and	5	3	1				
Field sprayer	Crop group 8 Fruiting vegeta	bles	5	4	1				
Sp. 17, 12	Pepper, celeriac, Crop group Crop group 13-07B Bush ber 07G Low growing berries		10	4	2				
	Outdoor ornamentals		10	5	2				
		Early growth stage	20	15	3				
4:11	Cherries (sweet and sour)	Late growth stage	15	5	2				
Airblast	Apple, crab apple, pear,	Early growth stage	30	20	10				
	Oriental pear	Late growth stage	20	15	4				
	Potato	Fixed wing	35	10	15				
	rotato	Rotary wing	30	10	15				
Aerial	Caybaan baan (day)	Fixed wing	70	20	20				
	Soybean, bean (dry)	Rotary wing	55	15	20				

For tank mixes, consult the labels of the tank-mix partners and observe the largest (most restrictive) buffer zone of the products involved in the tank mixture and apply using the coarsest spray (ASAE) category indicated on the labels for those tank mix partners.



List of References

A. Registrant Submitted Studies/Information

A.1 Environmental Fate and Effects Assessment

Published Information

PMRA Document Number	Reference
2712665	Li, D. et al., 2013, Acute immobilization of four neonicotinoid insecticides to Daphnia magna Straus, DACO: 9.3.2
2712686	Barbee, G.C. and Stout, M.J., 2009, Comparative acute toxicity of neonicotinoid and pyrethroid insecticides to non-target crayfish (Procambarus clarkia) associated with rice-crayfish crop rotations, DACO: 9.3.4
2712687	Cavallaro, M.C., C.A. Morrissey, J.V. Headley, K.M. Peru, and K. Liber, 2017, Comparative chronic toxicity of imidacloprid, clothianidin, and thiamethoxam to Chironomus dilutus and estimation of toxic equivalency factors, Environmental Toxicology and Chemistry. 36(2): 372-382, DACO: 9.3.4
2712688	Prosser, R.S. et al., 2016, Sensitivity of the early-life stages of freshwater mollusks to neonicotinoid and butenolide insecticides, DACO: 9.3.4
2712689	Riaz, M.A. et al., 2013, Molecular mechanisms associated with increased tolerance to the neonicotinoid insecticide imidacloprid in the dengue vector Aedes aegypti, DACO: 9.3.4
2712705	Stevens, M.M., Helliwell, S. and Hughes, P.A., 2005, Toxicity of Bacillus thuringiensis var. israelensis formulations, spinosad, and selected synthetic insecticides to Chironomus tepperi larvae, DACO: 9.3.5
2712706	Ugurlu, Unlu and Satar, 2015, The toxicological effects of Thiamethoxam on Gammarus kischineffensis (Schellenberg 1937) (Crustacea: Amphipoda), DACO: 9.3.5
2712707	Van den Brink, P.J. et al., 2016, Acute and chronic toxicity of neonicotinoids to nymphs of a mayfly species and some notes on seasonal differences, DACO: 9.3.5

Unpublished Information

PMRA	Reference
Document	
Number	
1074854	2004, Dissipation Trial to Determine Persistence and Leaching Movement of CGA 293343 and its Significant Soil Degradation Products After
1108402	Application of ACTARA 25WG or ACTARA 240SC., DACO: 8.3.2.1 2003, A Small-Scale Prospective Ground Water Monitoring Study for Platinum 2SC (Thiamethozam, CGA-293343) in St. Joseph County, Michigan - Progress Report No. 2, DACO: 8.3

1178192	1997, CGA 293343: Hydrolysis Of 14c-Guanidine-CGA-293343 Under Laboratory Conditions, DACO: 8.2.3.2
1178193	1998, CGA 293343: Hydrolysis Of 2-14c-Thiazolyl-CGA-293343 Under Laboratory Conditions, DACO: 8.2.3.2
1178196	1998, CGA 293343: Aerobic Soil Metabolism Of [14c-(Guanidine)] CGA-293343, DACO: 8.2.3.4.2
1178197	1998, CGA 293343: Aerobic Soil Metabolism Of [14c-Thiazole] CGA-293343, DACO: 8.2.3.4.2
1178198	1998, CGA 293343: Metabolism Of 14c-Guanidine-CGA-293343 In Viable And Sterile Clay Loam Soil Under Aerobic Conditions, DACO: 8.2.3.4.2
1178199	1998, CGA 293343: Soil Adsorption/Desorption Of [14c-Guanidine] CGA-293343 By The Batch Equilibrium Method, DACO: 8.2.4.2
1178249	1998, Cont'd From Roll 1776 - CGA 293343: Aged Leaching Of [14c-Guanidine]CGA-293343 In Representative Agricultural Soils, DACO: 8.2.4.3.2
1178359	1998, Four Soil Dissipation Trials To Determine Persistence And Leaching Movement Of CGA 293343, CGA 169374, CGA 173506, CGA 329351 And Their Significant Soil Metabolites After Application Of Helix As A Seed Treatment On Canola, DACO: 8.3.2.1
1196650	1998, Anaerobic Aquatic Metabolism Of [14c-Guanidine] CGA-293343 At Low Temperature, DACO: 8.2.3.5.6
1196651	1998, Part 1 Of 2: Aerobic Aquatic Metabolism Of [14c-Guanidine] CGA-293343 and [14C-Thiazole] CGA-293343, DACO: 8.2.3.5.2
1196652	2000, Time Dependent Sorption Of Technical And Of 2sc Formulated (Thiazolyl-2-14c)-Labeled CGA-293343 In Two Different Soils, DACO: 8.2.4.2
1196653	1997, Photodegradation Of 14c-[Guanidine]-CGA-293343 In Ph 5 Buffered Solution Under Artificial Light, DACO: 8.2.3.3.2
1196654	1998, Photodegradation Of 14c-[Thiazolyl]-CGA-293343 In Ph 5 Buffered Solution Under Artificial Light, DACO: 8.2.3.3.2
1196656	1997, Photodegradation Of 14c-[Thiazolyzl]-CGA-293343 On Soil Under Artificial Light, K. Sparrow, Completed July 7, 1997 (Abr-97011;372-96) [Thiamethoxam Technical;Subn.#1998-1753;Regn.#26665;Submitted December 4, 1998;Us Epa Submission: Volume 69 Of 123], DACO: 8.2.3.3.1
1196657	1997, Photodegradation Of 14c-[Guanidine]-CGA-293343 On Soil Under Artificial Light, DACO: 8.2.3.3.1
1196658	1998, Anaerobic Aquatic Metabolism Of [14c-Guanidine] CGA-293343, DACO: 8.2.3.5.6
1196659	1998, Anaerobic Aquatic Metabolism Of [14c-Thiazole] CGA-293343, DACO: 8.2.3.5.6
1196660	1998, Aerobic Aquatic Metabolism Of [14c-Guanidine] CGA-293343 And [14c-Thiazole] CGA-293343, DACO: 8.2.3.5.2
1196663	1998, CGA-293343 Technical: A 48-Hour Static Acute Toxicity Test With The Midge (Chironomus Riparius), DACO: 9.3.4
1196664	1996, Acute Toxicity Test Of CGA-293343 Technical To The Cladoceran

Daphnia Magna Straus Under Static Conditions, DACO: 9.3.2

1196666	1998, Soil Adsroption And Desorption Of Oxadiazinyl-14c-CGA-353042 By
	The Batch Equilibrium Method, Final Report, M. Scott, Completed
	November 24, 1998 (629-98) [Thiamethoxam Technical; Subn.#1998-
	1753;Regn.#26665;Submitted December 4, 1998;Us Epa Submission:
	Volume 78 Of 123], DACO: 8.2.4.2
1196667	1998, Soil Adsroption And Desorption Of [Thiazolyl-2-14c]-Noa-407475 By
1170007	The Batch Equilibrium Method, DACO: 8.2.4.2
1196669	1998, Soil Adsorption/Desorption Of [14c-Thiazole]CGA-322704 By The
1190009	Batch Equilibrium Method., DACO: 8.2.4.2
1106670	1
1196670	1998, Soil Adsorption/Desorption Of [14c]CGA-355190 By The Batch
	Equilibrium Method, M. Concha, T. Hathcock, Completed November 30,
	1998 (691w) [Thiamethoxam Technical; Subn.#1998-
	1753;Regn.#26665;Submitted December 4, 1998;Us Epa Submission:
	Volume 82 Of 123], DACO: 8.2.4.2
1196674	1997, CGA-293343: A 96-Hour Shell Deposition Test With The Eastern
	Oyster (Crassostrea Virginica), DACO: 9.4.4
1196685	1997, CGA-293343: A 96-Hour Flow-Through Acute Toxicity Test With
	The Saltwater Mysid (Mysidopsis Bahia), DACO: 9.4.2
1196696	1997, Daphnia Magna Reproduction Test: Effects Of CGA-293343 On The
	Reproduction Of The Cladoceran Daphnia Magna Straus In A Semi-Static
	Laboratory Test, DACO: 9.3.3
1196701	1998, Toxicity Test Of CGA-293343 Technical On Sediment-Dwelling
	Chironomus Riparius (Syn. Chironomus Thummi) Under Static Conditions,
	DACO: 9.3.4
1529715	2007, Environmental Chemistry and Fate Summary - Tier II, DACO:
10207110	12.7,8.1,8.2.1,8.2.3.1,8.2.4.1,8.3.1
1529718	2007, CGA 355190: n-Octanol / Water Partition Coefficient, DACO: 8.2.1
1327710	2007, CG/1 333170. If Country Water Landidon Coefficient, D/1CO. 0.2.1
1529731	1999, Hydrolysis of 14C-Guanidine-CGA 322704 Under Laboratory
	Conditions, DACO: 8.2.3.2
1529737	1998, Quantum Yield of the Photochemical Degradation of CGA 322704,
	DACO: 8.2.3.3.2
1529738	1997, Rate of Degradation of CGA 293343 in Soil Under Various
	Conditions, DACO: 8.2.3.4.2
1529739	2001, Degradation of [Oxadiazin-4-14C]-labelled NOA 407475 in Soil
	Gartenacker Under Aerobic Conditions at 20C, DACO: 8.2.3.4.2
1529740	2004, Rate of Degradation of [Oxadiazin-4-14C]-labelled NOA 407475
1323710	(Metabolite of Thiamethoxam) in Various Soils Under Aerobic Laboratory
	Conditions at 20C, DACO: 8.2.3.4.2
1529741	1996, Degradation of 14C-Thiazolring Labelled CGA 293343 in Various
1329741	Soils Under Laboratory Conditions, DACO: 8.2.3.4.2
1529745	
1329743	1999, Degradation of 14C-Thiazol Labelled CGA 293343 and its Major
	Metabolite CGA 322704 (14C-Thiazole Labelled) in Borstel Soil Under
1500546	Aerobic Conditions at 20C, DACO: 8.2.3.4.2
1529746	1999, Degradation of 14C-Thiazol Labelled CGA 322704 in Schwaderloch
	Soil Under Aerobic Conditions at 20C, DACO: 8.2.3.4.2
1529747	2001, Rate of Degradation of [Thiazol-2-14C]-CGA 322704 in Birkenheide
	Soil, DACO: 8.2.3.4.2

1529748	2004, Rate of Degradation of [Thiazol-2-14C]-labelled CGA 355190 in
	Various Soils Under Aerobic Laboratory Conditions at 20C, DACO:
1500750	8.2.3.4.2
1529750	2000, Anaerobic Degradation of 14C-Thiazol-Labelled CGA-322704 in Soil DACO: 8.2.3.4.4
1529752	1998, Degradation and Metabolism of 14C-Thiazolring Labelled CGA
1347134	293343 in Two Aerobic Aquatic Systems Under Laboratory Conditions,
	DACO: 8.2.3.5.4
1529753	1998, Degradation and Metabolism of 14C-Oxadiazinring Labelled CGA
	293343 in Two Aerobic Aquatic Systems Under Laboratory Conditions,
	DACO: 8.2.3.5.4
1529754	2000, Degradation and Metabolism of 14C-Thiazolring Labelled CGA
	322704 in Two Aerobic Aquatic Systems Under Laboratory Conditions,
	DACO: 8.2.3.5.4
1529758	1996, Adsorption/Desorption of CGA 293343 in Various Soil Types,
1.500550	DACO: 8.2.4.2
1529759	2001, Time Dependent Sorption of [Thiazolyl-2-14C]-CGA 322704 in
1529765	Birkenheide Soil, DACO: 8.2.4.2 2002, Adsorption/Desorption of [Thiazol-2-14C]-NOA 459602 in Various
1329703	Soils and Time Dependent Sorption, DACO: 8.2.4.2
1529766	2003, 14C-Labelled SYN 501406: Time Dependent Sorption in Three Soils,
1327700	DACO: 8.2.4.2
1529769	2001, Time Dependent Sorption of [Thiazolyl-2-14C]-CGA 293343 in
	Various Soils, DACO: 8.2.4.2
1529770	2001, Adsorption/Desorption of [Oxadiazin-4-14C]-CGA 293343 in
	Birkenheide Soil, DACO: 8.2.4.2
1529771	2001, Time Dependent Sorption of [Oxadiazin-4-14C]-CGA 293343 on
	Birkenheide Soil, DACO: 8.2.4.2
1529772	1997, Adsorption/Desorption of CGA 322704 in Various Soil Types,
1500774	DACO: 8.2.4.2
1529774	2001, Adsorption/Desorption of [Thiazolyl-2-14C]-CGA 322704 on Birkenheide Soil, DACO: 8.2.4.2
1529775	1998, Degradation and Leaching of 14C-CGA 293343 in Two Sand
1327113	Lysimeters Under Outdoor Conditions, DACO: 8.2.4.3
1529776	2005, [Thiazol-2-14C]-CGA 293343: Mobility and Degradation in Soil in
10230	Outdoor Lysimeters Following Seed Treatment Application, DACO: 8.2.4.3
1529777	1996, Leaching Model Study With CGA 293343 in Four Soils Under
	Laboratory Conditions, DACO: 8.2.4.3.1
1529778	1996, Leaching Characteristics of Aged Soil Residues of Thiazol- and
	Guanidine-14C-CGA 293343 in Two Soils After Percolation of 200 mm
	Artificial Rainfall, DACO: 8.2.4.3.2
1529779	1996, CGA 293343: Volatization of 14C-Thiazolring-Labelled CGA 293343
1500700	from Soil Surface Under Controlled Laboratory Conditions, DACO: 8.2.4.5
1529782	1997, Field Dissipation of CGA 293343 After Bareground Application of
1529783	[Thiazol-2-14C] Labelled Material, DACO: 8.2.3.4.2,8.3.2.3 1998, Residue Study With CGA 293343 In or On Soil in South of France,
1347/03	DACO: 8.3.2.3
	D1100, 0.3.4.3

1529784	1998, Residue Study With CGA 293343 In or On Soil in South of France, DACO: 8.3.2.3
1529785	1998, Determination of Residues of CGA 293343 and the Metabolite CGA
1529787	322704 in Soil, DACO: 8.3.2.3 1998, Study with GGA 293343 in or on Soil in Denmark, DACO: 8.3.2.3
1529788	1998, Study with GGA 293343 in or on Soil in Sweden, DACO: 8.3.2.3
1529789	2003, Field Soil Dissipation of Thiamethoxam In Bare Soil (Broadcast Application) and a Plot Cultivated With Tomatoes (Drench Application) at 1 Location in Spain, 200, DACO: 8.3.2.3
1529793	1996, Determination of the Residues of CGA 293343 and its Metabolite CGA 322704 in Soil After Application As WG-25 - Field Trial, DACO: 8.3.2.3
1529794	1997, Magnitude of the Residues of CGA 293343 and CGA 322704 in Soil After Application of Formulation A-9584C, DACO: 8.3.2.3
1529795	1997, Magnitude of the Residues of CGA 293343 and CGA 322704 in Soil After Application of Formulation A-9584C, DACO: 8.3.2.3
1529796	1996, Magnitude of the Residues of CGA 293343 and CGA 322704 in Soil After Application of Formulation A-9584C, DACO: 8.3.2.3
1529797	1998, Residue Study With CGA 293343 In or On Soil in North of France, DACO: 8.3.2.3
1529799	1998, Atmospheric Oxidation of CGA 293343 by Hydroxyl Radicals; Rate Estimation, DACO: 8.5
1529800	2004, CGA 322704: Calculation of Half-Life by Reaction With Atmospheric Hydroxyl Radicals, DACO: 8.5
1529851	2007, CGA-355190 - Acute Toxicity to Chironomus riparius Under Static Conditions, DACO: 9.3.4
1529852	2003, Effects of CGA 353042 (Metabolite of CGA 293343) on the Development of Sediment Dwelling Larvae of Chironomus riparius in a Water-Sediment System, DACO: 9.3.4
1529853	2007, NOA404617 - Acute Toxicity to Chironomus riparius Under Static Conditions, DACO: 9.3.4
1529854	2000, Toxicity Test of NOA-407475 (Metabolite of CGA 293343) on Sediment Dwelling Chironomus riparius (syn. Chironomus thummi) Under Static Conditions, DACO: 9.3.4
1751758	2008, A Small-Scale Prospective Groundwater Monitoring Study for Platinum 2SC (Thiamethoxam, CGA-293343) in St. Joseph County, Michigan-Final report, DACO: 8.5
1751760	2008, A Small-Scale Prospective Groundwater Monitoring Study for Thiamethoxam Insecticide (Platinum) in Macon County, Georgia, DACO: 8.5
2296375	2000, Acute Toxicity Test of CGA 293343 Tech. to the Ephemeroptera Cloeon sp. Under Static Conditions, DACO: 9.3.4
2446844	2002, Rate of Degradation of [Thiazol-2-14C]-Labelled CGA293343 and of [Imidazolidin-2-14C]-Labelled CGA256084 in Gartenacker Soil under Aerobic Laboratory Conditions at 20C, DACO: 8.2.3.4.2
2446849	2000, Degradation of Thiamethoxam in Soil after Barley Seed Treatment

	with [Thiazol-2-14C] CGA293343 under Indoor Conditions, DACO:
2446853	8.2.3.4.2,8.3.2 2006, Long Term Residue Study with Thiamethoxam (CGA293343) in or on
2440033	Different Vegetables and Soil in Switzerland - Final Report, DACO:
	6.3,8.3.2
2446854	2013, Dissipation of Thiamethoxam Applied In-Furrow at Spinach Planting
	And as a Foliar Application in the Central Valley of California, DACO: 8.3.2
2446857	2004, Residue Study with Thiamethoxam (CGA 293343) in or on Soil
	Cultivated with Seed Treated Winter Barley in Switzerland, DACO:
	7.4.1,8.3.2
2446859	2004, Residue Study with Thiamethoxam (CGA 293343) in or on Soil
	Cultivated with Seed Treated Winter Barley in France (South), DACO:
	7.4.1,8.3.2
2446861	2007, Thiamethoxam (CGA293343): Dissipation Study with A9584C,
0.500001	25WG, in or on Cultivated Soil in Switzerland, DACO: 8.3.2
2529331	2014, Thiamethoxam - Aerobic Aquatic Sediment Metabolism of
	[oxadiazine-4-14C]-Thiamethoxam and [thiazole-2-14C]-
2529332	Thiamethoxam, DACO: 8.2.3.5.4 2015, Thiamethoxam - Anaerobic Aquatic Sediment Metabolism of
2327332	[thiazole-2-14C]-Thiamethoxam and [oxadiazine-4-14C]-
	Thiamethoxam, DACO: 8.2.3.5.6
2681280	2016, Thiamethoxam WG (A9584C) - Effects of Chronic Exposure on
	Mayflies in an Outdoor Mesocosm, DACO: 9.3.6
2712668	1999, A9795B- A 48 hour static acute toxicity test with the Cladoceran
	(Daphnia magna), DACO: 9.3.2
2712669	2007, Actara 75WG (A9549C) - Acute toxicity to water fleas (Daphnia
	magna) under static conditions following OECD Guideline 202, DACO:
	9.3.2
2712670	1996, Acute toxicity test of CA2343A (intermediate of CGA293343) to the
2712672	Cladoceran Daphnia magna STRAUS under static conditions, DACO: 9.3.2
2712672	1998, Acute toxicity of NOA407475 (Metabolite of CGA293343) to
2712674	Daphnia magna in a 48-hour immobilization test, DACO: 9.3.2 1997, Acute toxicity of CGA322704 (Metabolite of CGA293343) to the
2/120/4	Cladoceran Daphnia magna STRAUS under static conditions, DACO: 9.3.2
2712675	1998, Acute toxicity test of CGA293343 WG25 (A-9584C) to the
	Cladoceran Daphnia magna STRAUS in the static system, DACO: 9.3.2
2712676	2000, Acute toxicity of CGA293343 FS600 (A9765C) to the Cladoceran
	Daphnia magna STRAUS in the static system, DACO: 9.3.2
2712677	2002, NOA459602 (Thiamethoxam metabolite)- Acute toxicity to Daphnia
	magna, DACO: 9.3.2
2712678	1998, Acute toxicity of A9700B Daphnia magna (Immobilisation text),
	DACO: 9.3.2
2712679	1998, Acute toxicity of CGA355190 to Daphnia magna (Immobilisation
2712690	test), DACO: 9.3.2
2712680	2001, CA2343A: Chronic toxicity to Daphnia magna, DACO: 9.3.3
2712681	2008, Acute toxicity to red swamp crayfish (Procambarus clarkia), under
	static-renewal conditions Final Report, DACO: 9.3.4

2712682	1998, Toxicity test of CGA322704 (metabolite of CGA293343) on
	sediment-dwelling Chironomus riparius (syn. Chironomus thummi) under
	static conditions, DACO: 9.3.4
2712683	2003, SYN501406 (Thiamethoxam metabolite): Toxicity to the sediment
	dweller Chironomus riparius using spiked water, DACO: 9.3.4
2712684	2002, CGA293343 (Thiamethoxam): Acute toxicity to aquatic invertebrates,
	DACO: 9.3.4
2712685	2003, CGA293343 (Thiamethoxam technical) and CGA322704
	(Thiamethoxam metabolite) Acute toxicity to a range of aquatic
	invertebrates, DACO: 9.3.4
2712691	2015, CGA282149 - Acute toxicity to larvae of Chironomus riparius in a 48-
2.120>1	hour immobilization test Final report, DACO: 9.3.4
2712692	2004, Thiamethoxam metabolite (NOA 421275): Acute toxicity to first instar
2712032	larvae of Chironomus riparius in a 48-hour test, DACO: 9.3.4
2712693	2015, 10- day toxicity test exposing midge (Chironomus dilutes) to
2/120/5	Thiamethoxam applied to sediment under static-renewal conditions- Final
	report, DACO: 9.3.4
2712695	1998, Toxicity test of CGA322704 (metabolite of CGA293343) on
2712033	sediment-dwelling Chironomus riparius (syn. Chironomus thummi) under
	static conditions, DACO: 9.3.4
2712696	2000, Acute toxicity test (24h) of CGA293343 tech to three invertebrate
2712070	species Daphnia pulex Leydig, Thamnocephalus platyurus, and Brachionus
	calyciflorus under static conditions, DACO: 9.3.4
2712697	2000, Acute toxicity test of CGA293343 tech to Gammarus sp. under static
2/12057	conditions, DACO: 9.3.4
2712699	2000, Acute toxicity test of CGA293343 tech to individual invertebrate
2712000	species and molluses from a natural pond assemblage under static conditions,
	DACO: 9.3.4
2712700	2005, Supplementary comments to Brixham Env. Lab Report BL7987B
	(2033605), DACO: 9.3.4
2712701	2002, Toxicity test of CGA293343 tech on Chaoborus sp. (invertebrate,
2,12,01	insect) under static conditions in a sediment-water-test system, DACO: 9.3.4
2712702	2005, Thiamethoxam WG25 (A9584C) Acute toxicity to Chironomus
2,12,02	riparius under static conditions, DACO: 9.3.5
2712703	2007, Thiamethoxam FS formulation (A9765N)- Acute toxicity to first-instar
2712703	larvae of Chironomus riparius, DACO: 9.3.5
2712704	2007, Thiamethoxam SC formulation (A9795B)- Acute toxicity to first-
2/12/04	instar larvae of Chironomus riparius, DACO: 9.3.5
2712708	2014, Thiamethoxam FS (A9765R) - Acute toxicity to first-instar larvae of
2/12/00	Chironomus riparius - Final report Amendment 1, DACO: 9.3.5
2712709	2003, Outdoor microcosm study to assess effects on aquatic organisms,
2/12/09	DACO: 9.3.6
2712710	
<i>4/14/1</i> 0	2004, Evaluation of the report Thiamethoxam 25WG (A9584C): Outdoor microcosm study to assess the effects on aquatic organisms, DACO: 9.3.6
2712712	
2712712	2015, Thiamethoxam - Life cycle toxicity test with mysids (Americamysis
	bahia) Final report, DACO: 9.4.5

A.2 Water Monitoring Assessment

Unpublished Information

PMRA	Reference
Document	
Number	
2818731	2017, Additional Ancillary Data for Ontario Water Monitoring Studies
	Conducted from 2012 to 2016, DACO: 8.6.1,8.6.2
2818733	2017, Additional Ancillary Data for Ontario Water Monitoring Studies
	Conducted from 2012 to 2016, DACO: 8.6.1,8.6.2
2870577	2018, Relevancy of Monitoring Sites for Aquatic Invertebrate Risk
	Assessment Classification of 2014 Water Sampling Locations (Morrissey),
	DACO: 8.6
2870578	2018, Relevancy of Monitoring Sites for Aquatic Invertebrate Risk
	Assessment Classification of 2014 Water Sampling Locations (Morrissey),
	DACO: 8.6

B. Additional Information Considered

B.1 Environmental Fate and Effects Assessment

Published Information

Belanger, S., M. Barron, P. Craig, S. Dyer, M. Galay-Burgos, M. Hamer, S. Marshall, L. Posthuma, S. Raimondo and P. Whitehouse. 2017. Future needs and recommendations in the development of species sensitivity distributions: Estimating toxicity thresholds for aquatic ecological communities and assessing impacts of chemical exposures. Integrated Environmental Assessment and Management 13(4): 664 – 674.

EFSA PPR Panel (EFSA Panel on Plant Protection Products and their Residues). 2013. Guidance on tiered risk assessment for plant protection products for aquatic organisms in edge-of-field surface waters. EFSA Journal 2013; 11 (7):3290, 268 pp. doi:10.2903/j.efsa.2013.3290.

USEPA (United States Environmental Protection Agency). 2011. Registration Review: Problem Formulation for the Environmental Fate, Ecological Risk, Endangered Species, and Drinking Water Exposure Assessments for Thiamethoxam. United States Environmental Protection Agency. Office of Pesticide Programs, Environmental Fate and Effects Division, Washington DC 20460. December 13, 2011.

Young, D. and M. Fry. 2017. Field –scale evaluation of pesticide uptake into runoff using a mixing cell and a non-uniform uptake model. Environmental Modelling & Software. 1-8. In Press. https://doi.org/10.1016/j.envsoft.2017.09.007

PMRA Document Number	Reference
2538669	C.A. Morrissey, P. Mineau, J.H. Devries, F. Sanchez-Bayo, M. Liess, M.C. Cavallaro and K. Liber. 2015. Neonicotinoid contamination of global surface waters and associated risk to aquatic invertebrates: A review. Environment International 74: 291-303, DACO: 8.6
2720027	Saraiva, A.S., R.A. Sarmento, A.C.M. Rodrigues, D. Campos, G. Federova, V. Zlabek, C. Gravato, J.L.T. Pestana and A.M.V.M. Soares., 2017, Assessment of thiamethoxam toxicity to Chironomus riparius., Ecotoxicology and Environmental Safety, 137: 240-246., DACO: 9.3.4,9.9
2764640	Finnegan, M.S., Baxter, L.R., Maul, J., Hanson M.L. and P.F. Hoekstra, 2017, Comprehensive characterization of the acute and chronic toxicity of the neonicotinoid insecticide thiamethoxam to a suite of primary producers, invertebrates, and fish, DACO: 9.9
2818524	Maloney, E.M., C.A. Morrissey, J.V. Headley, K.M. Peru and K. Liber, 2017, Cumulative toxicity of neonicotinoid insecticide mixtures to Chironomus dilutus under acute exposure scenarios, Maloney, E.M., C.A. Morrissey, J.V. Headley, K.M. Peru and K. Liber. 2017. Cumulative toxicity of neonicotinoid insecticide mixtures to Chironomus dilutus under acute exposure scenarios. Environ. Toxicol. and Chem. 36 (11): 3091-3101., DACO: 9.9
2841145	Ahmed, M.A.I and F. Matsumura, 2012, Synergistic actions of formamidine insecticides on the activity of pyrethroids and neonicotinoids against Aedes aegypti (Diptera: Culicidae), Ahmed, M.A.I and F. Matsumura. 2012. Synergistic actions of formamidine insecticides on the activity of pyrethroids and neonicotinoids against Aedes aegypti (Diptera: Culicidae). J. Med. Entomol. 49(6):1405-1410., DACO: 9.3.4,9.9
2841146	Uragayala S., V. Verma, E. Natarajan, P.S. Velamuri and R. Kamaraju, 2015, Adulticidal and larvicidal efficacy of three neonicotinoids against insecticide susceptible and resistant mosquito strains aegypti (Diptera: Culicidae), Uragayala S., V. Verma, E. Natarajan, P.S. Velamuri and R. Kamaraju.2015. Adulticidal and larvicidal efficacy of three neonicotinoids against insecticide susceptible and resistant mosquito strains. Indian J. Med. Res. 142(Supplement): 64-70., DACO: 9.3.4,9.9
2842540	Raby, M., M. Nowierski, D. Perlov, X. Zhao, C. Hao, D.G. Poirier and P.K. Sibley, 2018, Acute toxicity of six neonicotinoid insecticides to freshwater invertebrates, Raby, M., M. Nowierski, D. Perlov, X. Zhao, C. Hao, D.G. Poirier and P.K. Sibley. 2018. Acute toxicity of six neonicotinoid insecticides to freshwater invertebrates. Environ. Toxicol. and Chem. Accepted Article, DOI: 10.1002/etc.4088., DACO: 9.3.4,9.9
2861091	Bartlett, A.J., A.M. Hedges, K.D. Intini, L.R. Brown, F.J. Maisonneuve, S.A. Robinson, P.L. Gillis and S.R. de Solla, 2018, Lethal and sublethal toxicity of neonicotinoid and butenolide insecticides to the mayfly, Hexagenia spp, DACO: 9.3.4,9.9
2861918	Basley, K. and D. Goulson, 2018, Neonicotinoids thiamethoxam and clothianidin adversely affect the colonisation of invertebrate populations in

2862805	aquatic microcosms, Basley, K. and D. Goulson. 2018. Neonicotinoids thiamethoxam and clothianidin adversely affect the colonisation of invertebrate populations in aquatic microcosms. Environ. Sci. Poll. Res. https://doi.org/10.1007/s11356-017-1125-5, DACO: 9.3.4,9.9 Whiteside, M., P. Mineau, C. Morrison and L.D. Knopper, 2008, Comparison of a score-based approach with risk-based ranking of in-use agricultural pesticides in Canada to aquatic receptors, Whiteside, M., P. Mineau, C. Morrison and L.D. Knopper. 2008. Comparison of a score-based approach with risk-based ranking of in-use agricultural pesticides in Canada to aquatic receptors. Integr Environ Ass Mgmt. 4(2):215-236., DACO: 9.9
2862809	United States Environmental Protection Agency, 2017, Preliminary Risk
	Assessment to Support the Registration Review of Thiamethoxam, DACO: 12.5.8,12.5.9
2866915	Hladik, M.L., S. Bradbury, L.A. Schulte, M. Helmers, C. Witte, D.W.
	Kolpin, J.D. Garrett and M. Harris, 2017, Neonicotinoid insecticide removal by prairie strips in row-cropped watersheds with historical seed coating use, DACO: 8.6
2873503	Maloney, E.M., C.A. Morrissey, J.V. Headley, K.M. Perua and K. Liber,
	2018, Can chronic exposure to imidacloprid, clothianidin, and thiamethoxam mixtures exert greater than additive toxicity to Chironomus dilutus?, DACO: 9.3.4,9.9

Unpublished Information

PMRA	Reference
Document	
Number	
2518467	2004, An assessment of buffer zone effectiveness in reducing pesticide runoff from potato fields in Prince Edward Island (2001-2002). Environment Canada, Environmental Protection Branch, Atlantic Region., DACO: 8.6
2753706	2017, Final Progress Report (2014-2017) to the Ontario Ministry of the Environment and Climate Change. Grant Funding Agreement STF14-087 with Environment and Climate Change Canada: Assessment of acute and chronic toxicity of neonicotinoid insecticides to non-target aquatic species., see comments, DACO: 9.3.4

B.2 Water Monitoring Assessment

Published Information

PMRA Document Number	Reference
2526133	Main, A.R., J.V. Headley, K.M. Peru, N.L. Michel, A.J. Cessna, and C.A. Morrissey, 2014, Widespread use and frequent detection of neonicotinoid

insecticides in wetlands of Canada's Prairie Pothole Region. PLoS ONE 9(3): e92821, DACO: 8.6 2526184 Schaafsma, A., V. Limay-Rios, T. Beaute, J. Smith, and Y. Xue, 2015, Neonicotinoid insecticide residue in surface water and soil associated with commercial maize (corn) fields in Southwestern Ontario. PLoS ONE 10(2): e0118139, DACO: 8.6 2526820 Mineau, P., and C. Palmer, 2013, The impact of the Nation's most widely used insecticides on birds. American Bird Conservancy, March 2013. 96 pp., **DACO: 8.6** 2544468 Giroux, I., 2014, Présence de pesticides dans l'eau au Québec - Zones de vergers et de pommes de terre, 2010 à 2012. Québec, Ministère du Développement durable, de l'Environnement et de la Lutte contre les changements climatiques. Direction du suivi de l'état de l'environnement, ISBN 978-2-550-71747-8 (PDF), DACO: 8.6 2561884 Giroux, I., 2015, Présence de pesticides dans l'eau au Québec: Portrait et tendances dans les zones de maïs et de soya - 2011 à 2014, Québec, Ministère du Développement durable, de l'Environnement et de la Lutte contre les changements climatiques, Direction du suivi de l'état de l'environnement, ISBN 978-2-550-73603-5, Available: http://www.mddelcc.gouv.qc.ca/eau/flrivlac/pesticides.htm, DACO: 8.6 2572395 Main, A.R., N.L. Michel, M.C. Cavallaro, J.V. Headley, K.M. Peru, and C.A. Morrissey, 2016, Snowmelt transport of neonicotinoid insecticides to Canadian Prairie wetlands. Agriculture, Ecosystems and Environment 215: 76-84, DACO: 8.6 2608629 Main, A.R., N.L. Michel, J.V. Headley, K.M. Peru, and C.A. Morrissey, 2015, Ecological and landscape drivers of neonicotinoid insecticide detections and concentrations in Canada's Prairie Wetlands. Environmental Science & Technology 49:8367-8376, DACO: 8.6 2703534 Struger, J., J. Grabuski, S. Cagampan, E. Sverko, D. McGoldrick, and C.H. Marvin, 2017, Factors influencing the occurrence and distribution of neonicotinoid insecticides in surface waters of southern Ontario, Canada. Chemosphere 169: 516-523, DACO: 8.6 2821394 Giroux, I., 2017, Présence de pesticides dans l'eau de surface au Québec -Zones de vergers et de cultures maraîchères, 2013 à 2016. ISBN 978-2-550-78847-8, DACO: 8.6

Unpublished Information

PMRA Reference Document

Number	
2468268	Government of Prince Edward Island, 2014, Summary of pesticide detections in groundwater, surface water and sediment from the PEI Pesticide Monitoring Program (2004-2014). Downloaded from www.gov.pe.ca/pesticidemonitoring on October 24, 2014, DACO: 8.6
2523834	Alberta Environment and Sustainable Resource Development, 2013, Unpublished monitoring data on neonicotinoids in Alberta, received from Alberta Environment and Sustainable Resource Development. Information received on December 11, 2013 following the PMRA's request for monitoring data on neonicotinoids, DACO: 8.6
2523835	Alberta Agriculture and Rural Development, 2013, Summary of monitoring data on neonicotinoids in Alberta received from the Water Quality Branch of Alberta Agriculture and Rural Development. Information received on November 22, 2013, following the PMRA's September 29, 2013 request for monitoring data on neonicotinoids, DACO: 8.6
2523836	Ontario Ministry of the Environment and Ontario Ministry of Agriculture and Food, 2013, Unpublished water monitoring data for neonicotinoids in streams in Southern Ontario. Data received from a joint Ontario Ministry of Environment-Ministry of Agriculture and Food pesticide monitoring program. Data received on November 26, 2013 following the PMRA's request for monitoring data on neonicotinoids, DACO: 8.6
2523837	Ministère du Développement durable, de l'Environnement, de la Faune et des Parcs, 2013, Unpublished water monitoring data on neonicotinoids in Quebec water bodies from 2010 to 2012. Data received from the Ministère du Développement durable, de l'Environnement, de la Faune et des Parcs du Québec on November 27, 2013 following the PMRA's request for water monitoring data on neonicotinoids, DACO: 8.6
2523839	Environment Canada, 2014, Unpublished monitoring data on neonicotinoids in Ontario surface water in 2012 and 2013, from Environment Canada's Water Quality Monitoring and Surveillance Division in Burlington. Information received on January 15, 2014 following the PMRA's request for monitoring data on neonicotinoids, DACO: 8.6
2532563	Environment Canada, 2015, Unpublished monitoring data on neonicotinoids in Ontario surface water in 2014, from Environment Canada's Water Quality Monitoring and Surveillance Division in Burlington. Information received on May 13, 2015, DACO: 8.6
2548876	Pest Management Regulatory Agency, Pesticides detected in water and soil samples collected as part of the Hive Monitoring Program in 2014, Health Canada. Unpublished, DACO: 8.6

Pest Management Regulatory Agency, Pesticides detected in water and soil 2548877 samples collected during Bee Mortality Incidents in 2013 and 2014, Health Canada. Unpublished, DACO: 8.6 2612760 Main, A.R., J.V. Headley, K.M. Peru, N.L. Michel, A.J. Cessna, and C.A. Morrissey, 2014, RAW DATA for PMRA 2526133 - Widespread use and frequent detection of neonicotinoid insecticides in wetlands of Canada's Prairie Pothole Region. PLoS ONE 9(3): e92821. Raw data received from C.A. Morrissey on February 12, 2016, DACO: 8.6 2612761 Main, A.R., N.L. Michel, M.C. Cavallaro, J.V. Headley, K.M. Peru, and C.A. Morrissey, 2016, RAW DATA for PMRA 2572395 - Snowmelt transport of neonicotinoid insecticides to Canadian Prairie wetlands. Agriculture, Ecosystems and Environment 215: 76-84. Raw data received from C.A. Morrissey on February 4, 2016, DACO: 8.6 2612762 Main, A.R., N.L. Michel, J.V. Headley, K.M. Peru, and C.A. Morrissey, 2015, RAW DATA for PMRA 2608629 - Ecological and landscape drivers of neonicotinoid insecticide detections and concentrations in Canada's Prairie Wetlands. Environmental Science & Technology 49: 8367-8376. File also contains additional unpublished data for the summer of 2013. Data received from C.A. Morrissey on February 12, 2016, DACO: 8.6 2681876 Environment Canada, 2016, Unpublished monitoring data for neonicotinoid insecticides, fungicides (strobins and conazoles), acid herbicides, neutral herbicides, op insecticides, sulfonyls herbicides and carbamate pesticides in Ontario surface water in 2015, DACO: 8.6 2707947 Environment and Climate Change Canada, 2016, Unpublished water monitoring data for neonicotinoids in waterbodies from the Pacific Region of Canada from 2014 to 2015, DACO: 8.6 2709791 Ministère du Développement durable, de l'Environnement et de la Lutte contre les changements climatiques, 2016, Clothianidine Thiamethoxame 2015-2016 Projet 226, DACO: 8.6 2709792 Ministère du Développement durable, de l'Environnement et de la Lutte contre les changements climatiques, 2016, Clothianidine Thiamethoxame 2015-2016 Projet 4214, DACO: 8.6 2709793 Ministère du Développement durable, de l'Environnement et de la Lutte contre les changements climatiques, 2016, Clothianidine Thiamethoxame Gilbeault Delisle Norton, DACO: 8.6 2710505 Ontario Ministry of Environment and Climate Change, 2016, WWTP neonic data and sampling info, DACO: 8.6 2712893 Ontario Ministry of Environment and Climate Change, 2016, OMOECC monitoring studies for the year 2015 on pesticides, including neonicotinoids,

	in pollen, drinking water, soil, streams, and bumblebees, as well as baseline aquatic invertebrate community assemblages in southwestern Ontario, https://www.ontario.ca/search/data-catalogue?sort=asc&query=neonicotinoids, DACO: 8.6
2712896	Morrissey, C., 2016, Unpublished monitoring data on neonicotinoids in wetlands sampled in the summer of 2014 along breeding bird survey routes across Saskatchewan, DACO: 8.6
2745506	Prince Edward Island Department of Communities, Land and Environment, 2016, PEI Pesticide Monitoring Program's Stream Water Pesticide Analysis, 2009-2015. Available at: https://www.princeedwardisland.ca/en/service/pesticide-analysis-streamwater-open-data. Downloaded March 28, 2017, DACO: 8.6
2745819	Environment and Climate Change Canada, 2017, Water monitoring data for neonicotinoids from the Prairie provinces, 2014-2016. Data received through the Environmental Monitoring Working Group of Agriculture and Agri-Food Canada's Multi-stakeholder Forum on January 27, 2017, DACO: 8.6
2759002	Ontario Ministry of the Environment and Climate Change, Pesticide Network 2012-2014 Neonic Data, DACO: 8.6
2785041	Environment and Climate Change Canada, 2017, Water sampling from drainage ditches, streams and ponds around the Ottawa area, DACO: 8.6
2821395	Ministère du Développement durable, de l'Environnement et de la Lutte contre les changements climatiques, 2017, Unpublished water monitoring data for clothianidin and thiamethoxam in Quebec surface water in 2016 and 2017, DACO: 8.6
2834287	Environment and Climate Change Canada, 2017, Unpublished water monitoring data for pesticides in Great Lakes Tributaries, from 2007 to 2016, DACO: 8.6
2834289	Environment and Climate Change Canada, 2017, Unpublished water monitoring data for pesticides in the Atlantic region from 2013 to 2016, DACO: 8.6
2842169	Water Monitoring for Neonicotinoid Pesticides in British Columbia - 2017, DACO: 8.6
2842180	Neonicotinoid Water Monitoring Data for British Columbia in 2017, DACO: 8.6
2842307	Neonicotinoid Water Monitoring Data for Alberta in 2017, DACO: 8.6
2842433	Neonicotinoids in Surface Water from Alberta's Agricultural Areas: 2017

	Report, DACO: 8.6
2842449	Saskatchewan Water Monitoring Program for Neonicotinoid Pesticides 2017, DACO: 8.6
2842595	Neonicotinoid monitoring in surface and ground water in Manitoba 2017, DACO: 8.6
2845169	Neonicotinoid Water Monitoring Data for Prince Edward Island in 2017, DACO: 8.6
2847073	2017, Final Report - Prairie Wetland Neonicotinoid Monitoring Program, DACO: 8.6
2847083	EMWG - Data Collection - PPR Final 2017, DACO: 8.6
2849265	2017 Saskatchewan Neonicotinoid water sampling program, DACO: 8.6
2849266	Saskatchewan Neonicotinoid stream survey 2017 - 2014-2017 crop types, DACO: 8.6
2849359	Manitoba Neonic Monitoring Raw Data 2017, DACO: 8.6
2849370	Manitoba Crop Composition by Rural Municipality 2017, DACO: 8.6
2889992	2017, Unpublished water monitoring data for neonicotinoids in waterbodies from the Pacific Region of Canada in 2016, DACO: 8.6